

NOTICE

All drawings located at the end of the document.

**Draft Final
Corrective Measures Study/
Feasibility Study**

**Rocky Flats Environmental Technology Site
881 Hillside Area**

(Operable Unit No 1)



August 1994

ADMIN RECORD

A-OU01-000708

EXECUTIVE SUMMARY

This report documents the Corrective Measures Study/Feasibility Study (CMS/FS) that was performed for the 881 Hillside Area Operable Unit 1 (OU 1) of the Rocky Flats Environmental Technology Site (RFETS). The study was conducted in accordance with the requirements of the Rocky Flats Interagency Agreement (IAG) of January 1991. This agreement was signed between the U.S. Department of Energy (DOE), the U.S. Environmental Protection Agency (EPA) and the Colorado Department of Public Health and the Environment (CDPHE). The agreement specifies that the CMS/FS shall be conducted following appropriate Comprehensive Environmental Response, Compensation and Liability Act (CERCLA) and Resource Conservation and Recovery Act (RCRA) guidance.

The primary source of guidance used in the preparation of this report was EPA's *Guidance for Conducting Remedial Investigations and Feasibility Studies Under CERCLA*, which outlines and describes the requirements of the National Oil and Hazardous Substances Pollution Contingency Plan (NCP). In preparing this report, data on OU 1 were obtained from both the *Phase III RCRA Facility Investigation/Remedial Investigation (RFI/RI) Report* and the Rocky Flats Environmental Database System (RFEDS) directly. Where appropriate, the more recent RFEDS data were used to revise contaminant distribution maps and site depictions.

Following standard CERCLA guidelines, results of the Phase III RFI/RI report were first examined to determine primary site contaminants and exposure pathways. Once these risk drivers were identified, remedial action objectives (RAOs) and preliminary remediation goals (PRGs) were developed to address risks to human health and the environment. In the case of OU 1, the Environmental Evaluation (EE) portion of the Baseline Risk Assessment (BRA) did not identify any current or future risks to environmental receptors. Therefore, this report focuses on minimizing the risk to human receptors as identified in the Public Health Evaluation (PHE) portion of the BRA. The RAOs identified for OU 1 are listed below:

- 1) Prevent the inhalation of, ingestion of, and/or dermal contact with volatile organic compounds (VOCs) and inorganic contaminants in groundwater that would result

in a total excess cancer risk greater than 10^{-4} to 10^{-6} for carcinogens and/or a hazard index greater than or equal to one for non carcinogens

- 2) Minimize further degradation of groundwater beneath OU 1 by eliminating and/or containing residual subsurface soil dense non aqueous phase liquids (DNAPLs) to the maximum extent practicable
- 3) Prevent the inhalation of ingestion of and/or dermal contact with polynuclear aromatic hydrocarbons (PAHs) polychlorinated biphenyls (PCBs) and radionuclides in surface soils that would result in a total excess cancer risk greater than 10^{-4} to 10^{-6} for carcinogens and/or a hazard index greater than or equal to one for non-carcinogens
- 4) Prevent exposure to carcinogenic radionuclides in surface soil hotspots that would result in an excessive short-term risk to a human receptor

These RAOs presented to EPA and CDPHE in Technical Memorandum (TM) #10 were selected to address the primary risk exposure pathways identified for OU 1 the pathways associated with groundwater and surface soils. Because surface soil risks already fall within the acceptable risk range of 10^{-4} to 10^{-6} and because surface soil hotspots are being addressed through a recent Proposed Action Memorandum (PAM) alternatives were not developed for this medium as the RAOs are already achieved. PRGs for RAOs dealing with groundwater were identified by examining both risk- and applicable or relevant and appropriate requirement (ARAR)-based values. The exposure route of groundwater ingestion resulted in the highest potential risk to a future on site resident. Therefore State maximum contaminant levels (MCLs) were selected as appropriate PRGs for OU 1 groundwater.

After selecting appropriate PRGs for OU 1 remedial action alternatives for groundwater were assembled that would provide various conceptual approaches for cleanup of the site. The alternatives presented to EPA and CDPHE in TM #11 and selected for detailed analysis following a preliminary screening process were the following

- Alternative 0 No Action
- Alternative 1 Institutional Controls without the French Drain
- Alternative 2 Institutional Controls with the French Drain
- Alternative 3 Modified French Drain with Additional Extraction Wells

- Alternative 4 Groundwater Pumping and Soil Vapor Extraction (SVE)
- Alternative 5 Groundwater Pumping and SVE with Thermal Enhancement
- Alternative 6 Hot Air Injection with Mechanical Mixing
- Alternative 7 Soil Excavation and Groundwater Removal with Sump Pumps

These alternatives were subjected to detailed analysis as required by CERCLA and the NCP [40 Code of Federal Regulations (CFR) 300.430]. The criteria used to analyze the alternatives are the following:

- Overall protection of human health and the environment
- Compliance with ARARs
- Long-term effectiveness and permanence
- Reduction of toxicity, mobility, or volume
- Short term effectiveness
- Implementability
- Cost
- State acceptance
- Community acceptance

The two threshold criteria—overall protection of human health and the environment and compliance with ARARs—are statutory requirements that must be satisfied by any alternative in order for it to be eligible for selection as the preferred remedial action alternative. The five primary balancing criteria—long term effectiveness and permanence, reduction in toxicity, mobility and volume, short term effectiveness, implementability, and cost—are used to evaluate major performance objectives for each alternative. The performance of each alternative in addressing each primary balancing criterion is evaluated and then compared across alternatives to assist in the selection of a preferred alternative.

The two modifying criteria—state acceptance and community acceptance—evaluate the potential acceptance of the preferred alternative by regulatory agencies and the community. These last two criteria are not evaluated until after formal public comment on the CMS/FS and Proposed Remedial Action Plan/Proposed Plan (PRAP/PP) and are addressed in the final Corrective Action Decision/Record of Decision (CAD/ROD).

For OU 1 the detailed analysis of alternatives demonstrates that Alternative 1 Institutional Controls without the French Drain is the preferred alternative for groundwater remediation. This alternative consists of institutional controls to prevent unauthorized access to the 881 Hillside area and discontinuing use of the existing french drain system. Groundwater modeling conducted to support the CMS/FS indicates that under this alternative the PRGs (MCLs) will not be exceeded at Woman Creek for OU 1 contaminants of concern (COCs). This alternative results in one of the lowest overall costs, while still achieving a residual risk level of 1.99×10^{-6} at this location. The associated peak concentration predicted for PCE (the selected indicator chemical) is 3.60×10^{-3} mg/l. This is below its respective MCL of 5×10^{-3} mg/l.

These values are considered extremely conservative based on the assumptions used in the groundwater model (as discussed in Appendix B). Several significant loss mechanisms are currently not included in the model which tends to overestimate actual future predicted concentrations. In particular, volatilization, a significant loss mechanism for the volatile organic compounds (VOCs) identified as COCs for OU 1, would reduce the concentrations of these contaminants prior to reaching Woman Creek. The retardation and biodegradation factors used in the model are also extremely conservative.

Alternative 1 meets both of the threshold criteria discussed above, as well as providing long term effectiveness and permanence through natural attenuation and degradation. The toxicity, mobility, and volume of OU 1 groundwater COCs would be reduced through dispersion, biological degradation, and volatilization. In terms of short term effectiveness and implementability, this alternative is one of the most implementable alternatives proposed, which results in the lowest short term risks to workers, the public, and the environment. If at any time during the monitoring period COC concentrations appear higher than predicted, the french drain sumps would be pumped to the Building 891 water treatment plant to provide additional protection. This alternative results in a very low total present worth cost because institutional controls are currently in place at the RFETS. Monitoring would be continued under this alternative throughout the institutional control period.

State and community acceptance of this alternative will be evaluated after comments are received on the CMS/FS report and the PRAP/PP. At this time the results of this CMS/FS indicate that Alternative 1 Institutional Controls without the French Drain is the preferred remedial action alternative for OU 1 groundwater.

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LIST OF ACRONYMS

ALARA	as low as reasonably achievable
ARAR	applicable or relevant and appropriate requirement
BRA	Baseline Risk Assessment
CAD/ROD	Corrective Action Decision/Record of Decision
CAMU	Corrective Action Management Unit
CCl ₄	carbon tetrachloride
CCR	Colorado Code of Regulations
CDPHE	Colorado Department of Public Health and the Environment
CERCLA	Comprehensive Environmental Response Compensation and Liability Act
CFR	Code of Federal Regulations
CMS/FS	Corrective Measures Study/Feasibility Study
COC	contaminant of concern
CWQCC	Colorado Water Quality Control Commission
DCE	dichloroethene
DNAPL	dense non aqueous phase liquid
DOE	U S Department of Energy
EDE	effective dose equivalent
EE	Environmental Evaluation
EPA	U S Environmental Protection Agency
GAC	granular activated carbon
GRAs	general response actions
IAG	Inter Agency Agreement
IHSSs	Individual Hazardous Substance Sites
IM/IRA	Interim Measure/Interim Remedial Action
LHSU	lower hydrostratigraphic unit
MCL	Maximum Contaminant Level
MCLG	Maximum Contaminant Level Goal
NCP	National Oil and Hazardous Substances Pollution Contingency Plan
NESHAP	National Emission Standards for Hazardous Air Pollutants
O&M	operation and maintenance
OSWER	Office of Solid Waste and Emergency Response
OU 1	Operable Unit 1
OU 2	Operable Unit 2

PAHs	polynuclear aromatic hydrocarbons
PAM	Proposed Action Memorandum
PCBs	polychlorinated biphenyls
PCE	perchloroethene (or tetrachloroethene)
PHE	Public Health Evaluation
PRAP/PP	Proposed Remedial Action Plan/Proposed Plan
PRG	preliminary remediation goal
RF	radio frequency
RAOs	remedial action objectives
RCRA	Resource Conservation and Recovery Act
RFEDS	Rocky Flats Environmental Database System
RFETS	Rocky Flats Environmental Technology Site
RFI/RI	RCRA Facility Investigation/Remedial Investigation
SDWA	Safe Drinking Water Act
SID	South Interceptor Ditch
SWMU	Solid Waste Management Unit
SVE	soil vapor extraction
SVOC	semivolatile organic compound
TBC	to-be-considered
TCA	trichloroethane
TCE	trichloroethene
TM	Technical Memorandum
TMV	toxicity mobility or volume
TSCA	Toxic Substances Control Act
UHSU	upper hydrostratigraphic unit
USC	United States Code
VOC	volatile organic compound

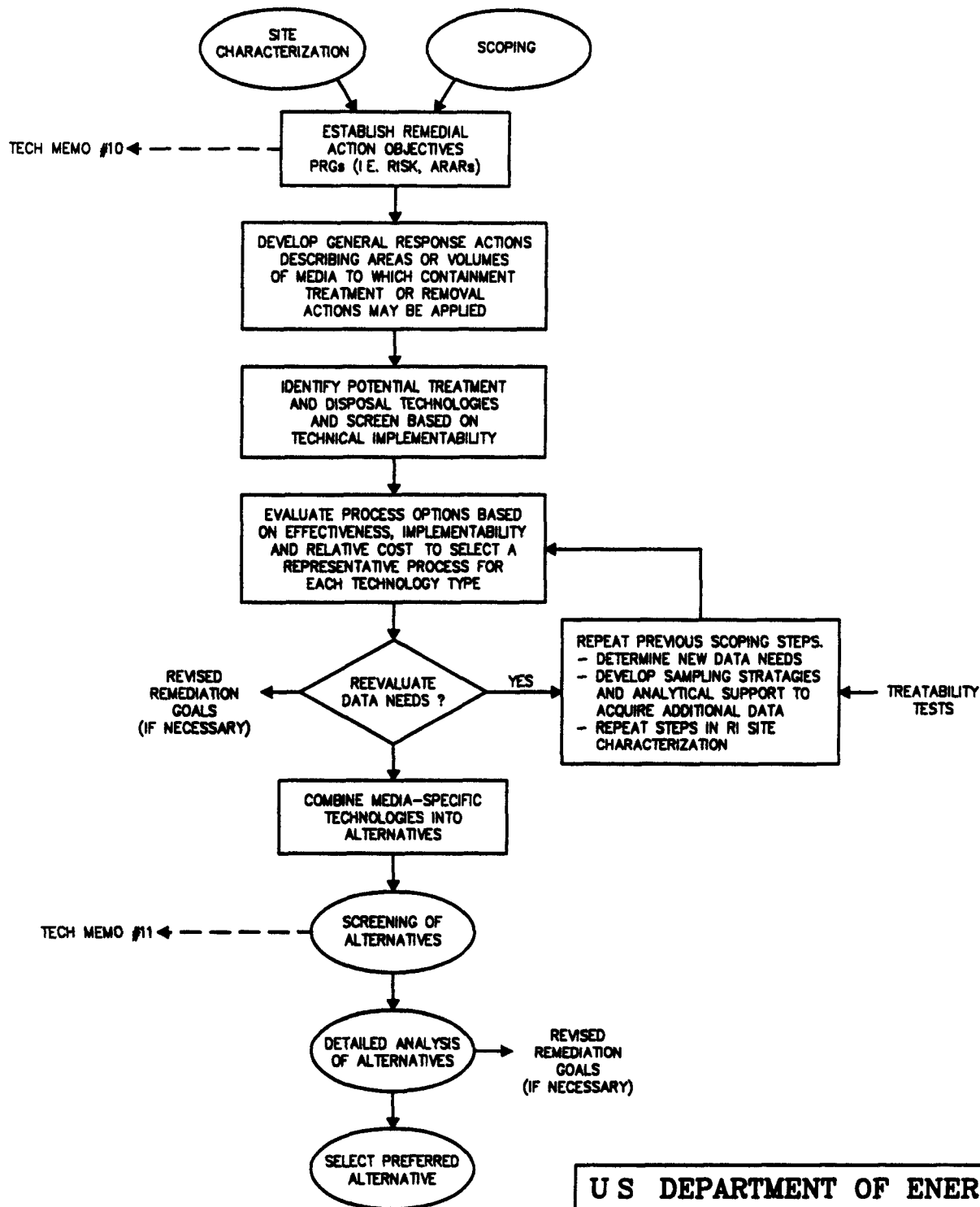
Section 1

1 0 INTRODUCTION

This Corrective Measures Study/Feasibility Study (CMS/FS) report is part of a comprehensive program developed by the U S Department of Energy (DOE) the U S Environmental Protection Agency (EPA) and the Colorado Department of Public Health and the Environment (CDPHE) This program is authorized pursuant to the Rocky Flats Interagency Agreement (IAG) of January 1991 In accordance with the requirements of the IAG this CMS/FS report addresses provisions of both the Comprehensive Environmental Response Compensation and Liability Act (CERCLA), and the Resource Conservation and Recovery Act (RCRA) Background information on Operable Unit 1 (OU 1) of the Rocky Flats Environmental Technology Site (RFETS) was obtained primarily from the *Phase III RCRA Facility Investigation/Remedial Investigation (RFI/RI) Report* (DOE 1994a) However wherever appropriate more recent data were used to develop figures and contour maps presented herein These data were obtained directly from the Rocky Flats Environmental Database System (RFEDS) and were used to supplement the information presented in the Phase III RFI/RI report

1 1 Purpose and Organization of Report

This CMS/FS is based on the CERCLA RI/FS process developed by EPA for the Superfund program (EPA 1988a EPA 1990b) Essentially the process is designed to provide decision makers with a tool by which they can make an informed decision regarding the preferred remediation alternative for a suspected hazardous waste site The methodology that EPA has established for this type of study is outlined in the *Guidance for Conducting Remedial Investigations and Feasibility Studies Under CERCLA* (EPA 1988a) and is shown graphically in Figure 1 1 Pursuant to the IAG, two technical memorandums (TMs) were prepared to present the general approaches proposed for the CMS/FS to the regulatory agencies involved prior to submitting the draft CMS/FS report Technical Memorandum #10 *Development of Remedial Action Objectives* (DOE 1994b), and Technical Memorandum #11 *Development and Screening*



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**881 HILLSIDE AREA
 OPERABLE UNIT NO 1**

**CMS/FS Logic Flow
 Diagram**

Figure 1-1

for review and will not be finalized since they do not require formal approval Comments received on the documents are incorporated herein where appropriate

The method proposed by EPA can be viewed as occurring in three distinct phases These phases are (1) the development of alternatives (2) the screening of alternatives and (3) the detailed analysis of alternatives

The first phase of the CMS/FS involves determining which technologies will be used in the development of alternatives and then combining these technologies to form a range of remedial alternative options for the operable unit This determination is based on the following items and is documented in TM #10 and #11 results are incorporated in Section 2 0 of this report

- Development of media specific remedial action objectives (RAOs)
- Development of media specific general response actions (GRAs)
- Identification of volumes and/or areas of the media which require GRAs
- Identification and screening of technologies and process options for each GRA
- Evaluation of process options within each technology type to select a representative process option for the development of remedial action alternatives

The second phase of the CMS/FS is an optional step depending on the number of alternatives developed during the first phase If numerous waste management options were developed after the screening of technologies then these alternatives can be screened to reduce the number of alternatives that are carried forward for detailed analysis This screening is conducted on the basis of effectiveness implementability and cost and is documented in TM #11 Results of the screening are incorporated in Section 3 0 of this report

The final phase in the CMS/FS process is documented in Section 4 0 and consists of the detailed analysis of alternatives which were carried forward from the screening phase described above In this phase the alternatives are further refined and analyzed in detail with respect to nine

criteria as required by the National Oil and Hazardous Substances Pollution Contingency Plan (NCP) The criteria are listed below

- Overall protection of human health and the environment
- Compliance with Applicable or Relevant and Appropriate Requirements (ARARs)
- Long term effectiveness and permanence
- Reduction of toxicity mobility or volume
- Short term effectiveness
- Implementability
- Cost
- State acceptance
- Community acceptance

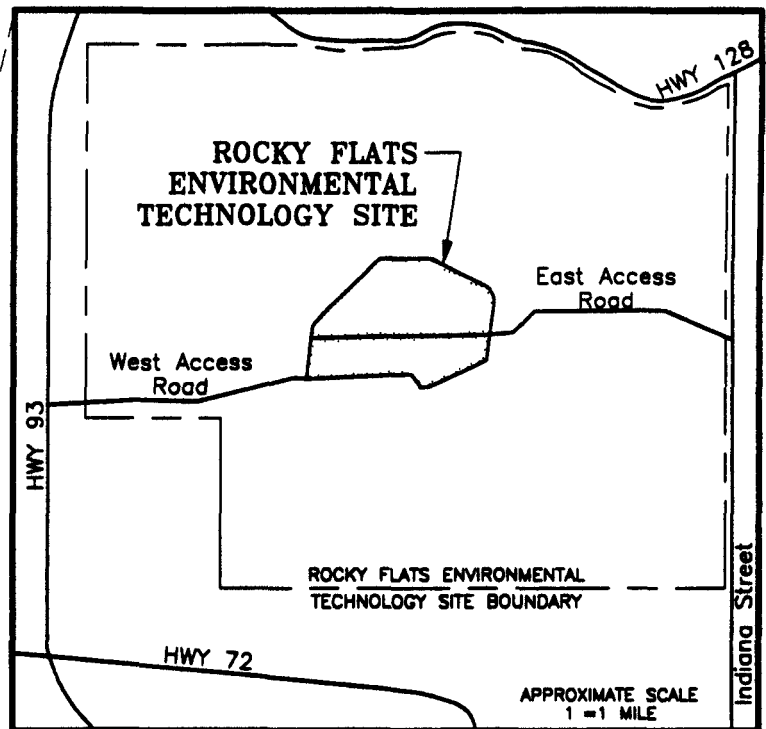
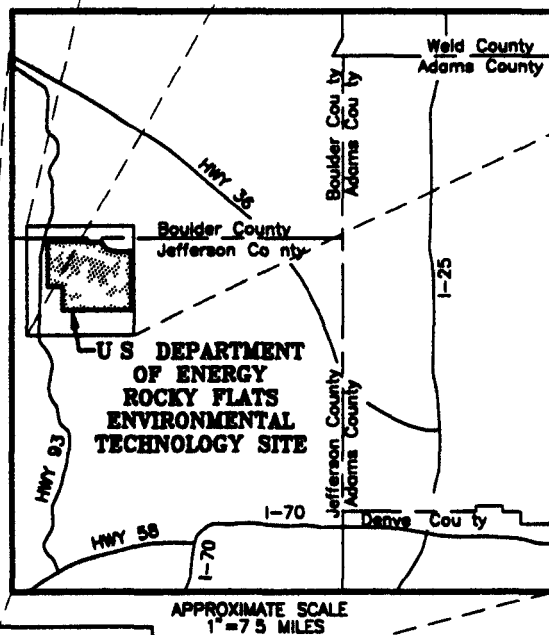
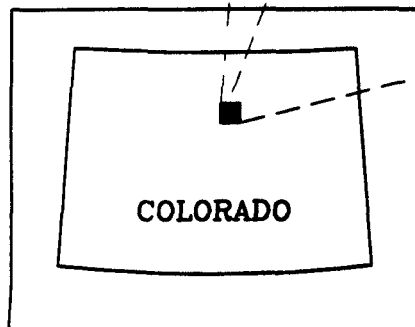
In the detailed analysis these criteria are evaluated in two ways First each alternative is evaluated individually on its ability to satisfy the nine criteria, and second the alternatives are subjected to a comparative analysis (against each other) to assess the relative performance of each alternative against the criteria

These chapters document the CMS/FS process as it was applied to OU 1 of the RFETS Sections 2 0 through 4 0 contain the technical basis for the selection of a preferred alternative while Section 1 0 presents the background for the report A site description and history along with a summary of the extent of contamination and the results of the baseline risk assessment are included in this section Several appendices are included to support the information presented in the CMS/FS

1 2 Background Information

The RFETS is a DOE owned facility and is located approximately 16 miles northwest of downtown Denver Colorado (see Figure 1 2) RFETS occupies approximately 6,550 acres of federally-owned land in northern Jefferson County Colorado The majority of the RFETS plant buildings are located within a 400-acre area referred to as the RFETS security area The 6 150

OUT-LOC DWG



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881 HILLSIDE AREA
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General Location of
Rocky Flats Environmental
Technology Site

Figure 1-2

acres surrounding the plant buildings provide a buffer zone around the RFETS security area RFETS is operated and managed by EG&G Rocky Flats Inc for DOE

In July 1994 the plant was renamed to the RFETS to better reflect its new mission of environmental restoration and the advancement of new and innovative technologies for waste management characterization, and remediation Until 1992 RFETS fabricated nuclear weapon components from plutonium, uranium beryllium and stainless steel Parts made at the plant were shipped elsewhere for assembly Support activities included chemical recovery and purification of recyclable transuranic radionuclides and research and development in metallurgy machining nondestructive testing coatings remote engineering chemistry and physics The production process at RFETS resulted in the generation of radioactive and non radioactive wastes On site storage and disposal of these wastes has contributed to hazardous and radioactive contamination in soils, surface water and groundwater

1 2 1 881 Hillside Site Background and Description

Previously Building 881 was used for enriched uranium operations and stainless steel manufacturing The laboratories in Building 881 also performed analyses of the materials generated in production The building is located south of the plant on a south facing hillside which then slopes down to Woman Creek Topographically the highest point near OU 1 is Building 881, approximately 6 000 feet above mean sea level and the lowest point is in Woman Creek about 5 830 feet above mean sea level Two surface drainages occur in the vicinity of OU 1 Woman Creek flows along the base of 881 Hillside south of OU-1 and the South Interceptor Ditch (SID) crosses OU 1 between the plant and Woman Creek A french drain was constructed in 1993 across a significant portion of OU 1 above the SID to collect alluvial groundwater as an Interim Measure/Interim Remedial Action (IM/IRA)

OU 1 includes 11 units previously identified as Individual Hazardous Substance Sites (IHSSs) where past operational practices have resulted in the contamination of soils surface water, and/or groundwater Of the 11 IHSSs, two have been identified as having potentially the most

contamination. These two areas are identified as IHSSs 119.1 and 119.2 and are referred to as Multiple Solvent Spill Sites. The areas are located east of Building 881 and along the southern perimeter road. Beginning in 1967, these two areas were used as drum storage areas. The drums contained unknown types and quantities of solvents and solvent wastes and possibly some radionuclides. Drum storage operations were suspended at IHSSs 119.1 and 119.2 in 1972. OU 1 includes nine additional IHSSs (see identification numbers in Figure 1.3). Brief descriptions of all OU 1 IHSSs are included in Table 1.1.

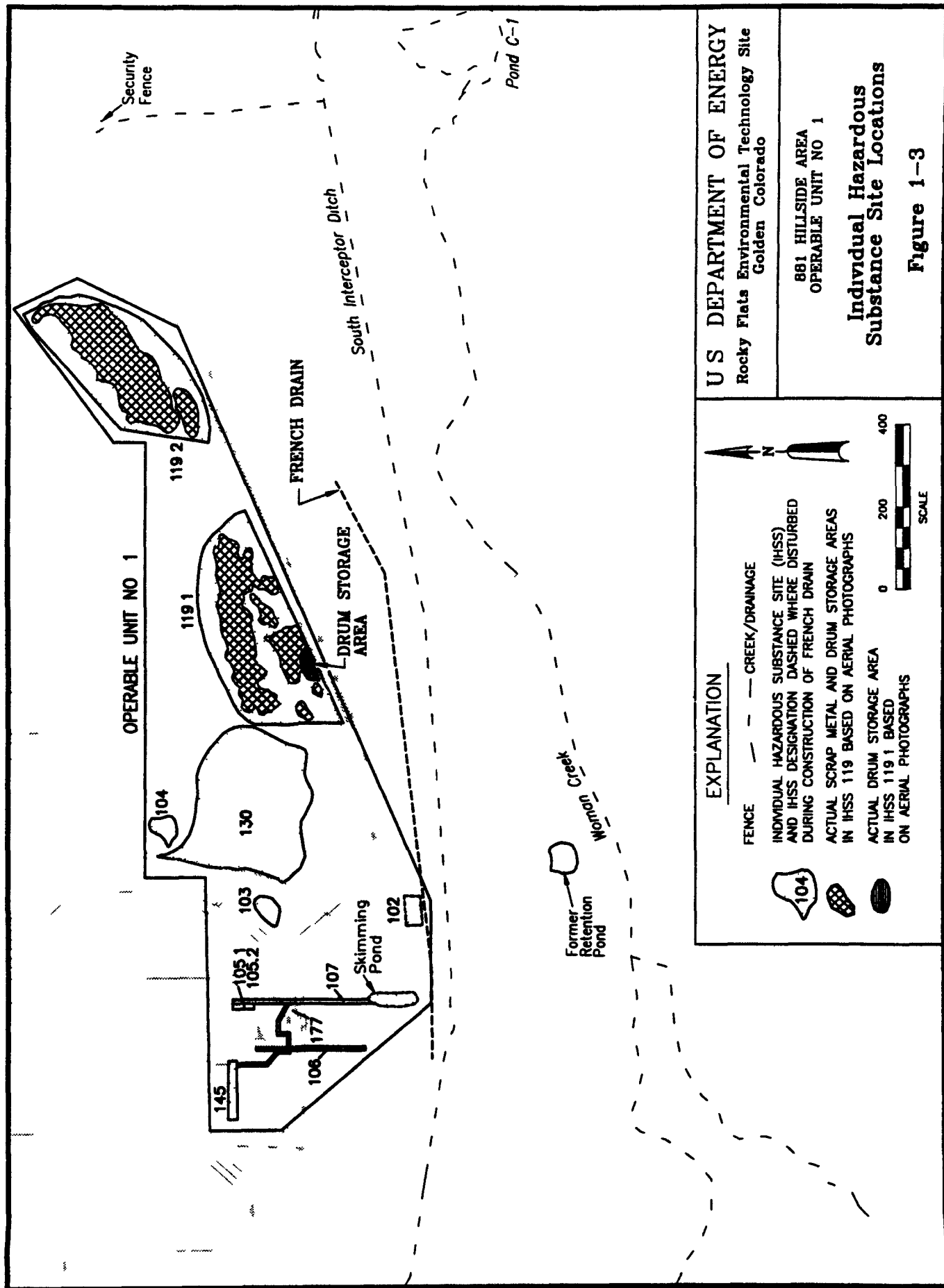
1.2.2 Geomorphology

The geomorphology of a site can influence potential contaminant transport pathways, including surface water and groundwater flow. The geomorphology at OU 1 reflects the interaction of several erosional and depositional processes on the bedrock and surficial materials underlying the site and accounts for the gently rolling to moderately steep slopes developed on 881 Hillside. Subsequent to the initial siting of the plant, the terrain has been recontoured in several areas at various times. These include the construction of Building 881, the placement of fill and waste materials in several areas including the contractor yard and several IHSSs, the grading of roads at the site, the construction of the SID, and most recently, the construction of the french drain.

The steepness of the hillside, combined with various construction and excavation activities at OU 1, has resulted in mechanical failure manifested in widespread slumping of material. The number of damaged wells on the hillside testifies to the prevalence of earth movement. Previous studies have also delineated slumps in the 881 Hillside area. One study map shows the entire hillside as being susceptible to landslides (DOE 1994a).

1.2.3 Hydrogeology

Groundwater hydrogeology has been a central component of three phases of study at OU 1. The most recent interpretations in the Phase III RFI/RI report represent a comprehensive evaluation



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881 HILLSIDE AREA
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**Individual Hazardous
 Substance Site Locations**

Figure 1-3

Table 1 1
Individual Hazardous Substance Site Descriptions

IHSS Number	IHSS Name	Description
102	Oil Sludge Pit Site	Approximately 40 x 70 ft ² area located approximately 180 feet south of Building 881 where 30 to 50 drums of non-radioactive only sludge were emptied in the late 1950s. The sludge was from the cleaning of two No. 6 fuel oil tanks designated as IHSSs 105 1 and 105 2 and was backfilled when disposal operations ceased.
103	Chemical Burial Site	Approximately 50 feet in diameter (2,000 ft. ²) the pit is circular in shape and is located approximately 150 feet southeast of Building 881 on 1963 aerial photographs. Area was reportedly used to bury unknown chemicals.
104	Liquid Dumping Site	Reportedly a former (pre-1969) liquid waste disposal pond in area east of Building 881. No exact location or dimensions of pit location is uncertain due to poor quality of 1965 aerial photograph. Approximate dimensions are 50 x 50 ft ² .
105 1 105 2	Out-of-Service Fuel Oil Tank Sites	Located immediately south of Building 881 these were storage tanks for No. 6 fuel oil. Suspected leaks in 1972. Tanks closed in place through filling with asbestos-containing material and cement. IHSS 107 the Hillside Oil Leak Site may have been caused by leakage from these tanks.
106	Outfall Site	Overflow line from the sanitary sewer sump in Building 887. The outfall was used for discharge of untreated sanitary wastes in the 1950s and 1960s. Due to concern about discharges from the outfall entering Woman Creek, several small retention ponds and an interceptor ditch were built in 1955 and 1979 respectively to divert the outfall water to Pond C 2.
107	Hillside Oil Leak Site	Site of 1972 fuel oil spill from Building 881 foundation drain outfall. A concrete skimming pond was built below the foundation drain outfall to contain the oil flowing from the foundation drain, and an interceptor ditch was constructed to prevent oil contaminated water from reaching Woman Creek.
119 1 119 2	Multiple Solvent Spill Sites	Former drum storage areas east of Building 881 along the southern perimeter road. IHSS 119 1 is the larger western drum and scrap metal storage area and appears to have contained mostly drums in the southern part of the IHSS and mostly scrap metal in the northern part, although material was moved around frequently as documented by aerial photographs. IHSS 119 2 is the smaller eastern drum and scrap metal storage area and appears to have contained mostly scrap metal. The drums contained unknown quantities and types of solvents and wastes. The scrap metal may have been coated with residual oils and/or hydraulic coolants.
130	Radioactive Site 800 Area #1	Area east of Building 881. Used between 1969 and 1972 to dispose of soil and asphalt contaminated with low levels of plutonium and uranium. IHSS 130 is referred to as the Contaminated Soil Disposal Area East of Building 881 in the HRR to better match the history of waste disposal. The site is included in the discussion of the 900 area at RFETS in that report. IHSS 130 contains approximately 320 tons or 250 cubic yards which came from three sources: 1) plutonium-contaminated soil and asphalt placed in September of 1969; 2) road asphalt and soil rad contaminated by leaking drum in transit; and 3) 60 cu yds of plutonium-contaminated soil removed from around the Building 774 process waste tanks in 1972.
145	Sanitary Waste Line Leak	Six inch cast iron sanitary sewer line that originates at the Building 887 lift station and that leaked on the hillside south of Building 881. The line had conveyed sanitary wastes and low-level radioactive laundry effluent to the sanitary treatment plant from about 1969 to 1973.

of the OU hydrogeology based on eight years of investigation and monitoring. Groundwater at OU 1 is present in the unconsolidated surficial material consisting of the Rocky Flats Alluvium colluvial material and the Valley Fill Alluvium. Groundwater is also inferred to occur locally in the upper portion (i.e. 0 to 25 feet) of the Laramie claystone bedrock. These units contain unconfined groundwater and comprise the upper hydrostratigraphic unit (UHSU). Groundwater also occurs in deeper (>25 feet) bedrock sandstones and claystones of the upper Laramie Formation. This bedrock unit is labeled the lower hydrostratigraphic unit (LHSU) and groundwater here is confined in places. The Phase III RFI/RI results for OU 1 indicate the presence of these units, based on their exhibiting different hydrogeological characteristics although defining the boundary is difficult.

Over most of the site, UHSU groundwater flow occurs in disconnected northwest southeast trending channels that have been scoured into the bedrock surface. Bedrock highs and lithologic variability, notably the presence of clay lenses, act to retard the rate of groundwater flow. Flow has been observed in glide planes bounding the slump blocks. Parts of OU 1 particularly in the eastern portion are only seasonally wet and contain groundwater only in the spring months when there is high precipitation. Groundwater levels across OU 1 are higher in spring than in the remainder of the year.

Recharge to the UHSU is minimal and is primarily through infiltration of precipitation, which ranges from 2 inches per hour for initial infiltration to 0.5 inches per hour for final (saturated) infiltration. Localized sources of recharge include seepage from the Rocky Flats Alluvium to colluvial materials and former recharge from the Building 881 footing drain which has since been rerouted to the french drain collection system. Flow from this drain averages 3.5 gallons per minute (gpm). Discharge occurs largely through evapotranspiration and discharge at boundaries such as seeps, Woman Creek, the South Interceptor Ditch (SID) and the french drain (DOE 1994a).

From aquifer test data, the average linear flow velocity was estimated at 70 feet per year in the vicinity of IHSS 119, 18 feet per year in the vicinity of Building 881, and 180 feet per year

within the Valley Fill Alluvium. The volume of UHSU groundwater at OU 1 was estimated at 5.8 acre feet in January 1992 to 5 acre feet in April 1992 (DOE 1994a).

Measured hydraulic conductivities vary widely because geological characteristics that control permeabilities vary widely in the materials that comprise the unconsolidated material of the UHSU. The overall range of hydraulic conductivity values estimated for UHSU materials was 3×10^{-3} to 2×10^{-6} cm/sec. The hydrologic data show that a wide range of hydraulic conductivity values characterize the surficial materials at OU 1. Also, the horizontal hydraulic conductivity values in bedrock appear to be 10 to 1,000 times greater than values in the vertical direction, which range from 1.2×10^{-3} to 2.5×10^{-9} . Water levels screened in the UHSU rise annually in response to spring recharge and decline during the remainder of the year (DOE 1994a).

Based on groundwater level data from the french drain monitoring wells, the recently installed french drain system appears to be effective in capturing UHSU groundwater and aqueous phase contaminants migrating from OU 1. Data from April 1993, a month of high precipitation, showed that most of the UHSU monitoring wells were dry (DOE 1994a).

1.3 Nature and Extent of Contamination

This section summarizes the results of the nature and extent of contamination at OU 1 as presented in the Phase III RFI/RI report. This information is presented by contaminant group with specific areas identified as impacted. Table 1.2 summarizes the contaminants originally identified in the Phase III RFI/RI report for the media of groundwater, surface soils, subsurface soils, surface water, and sediments.

Table 1 2
Contaminants Identified in the RFI/RI by Media

Contaminant	Ground Water	Surface Soil	Subsurface Soil ^a	Surface Water ^a	Sediment ^a
Volatile Organic Compounds					
Carbon Tetrachloride	X		X		
Chloroform	X		X		
1 1 Dichloroethane	X		X	X	
1 2 Dichloroethane	X		X	X	
1 1 Dichloroethene	X		X	X	
1 2 Dichloroethene	X		X	X	
cis 1 2 Dichloroethene	X		X	X	
Tetrachloroethene	X		X	X	
Toluene	X		X	X	X
Total Xylenes	X		X	X	
1 1 1 Trichloroethane	X		X	X	X
1 1 2 Trichloroethane	X		X	X	
Trichloroethene	X		X	X	
Metals					
Selenium	X				
Vanadium	X				
Radionuclides					
Americium		X	X	X	X
Uranium		X	X	X	
Plutonium		X	X	X	X
Polychlorinated Biphenyls (PCBs)					
AROCLOR 1248		X			
AROCLOR 1254		X			X

**Table 1-2
(Continued)**

Contaminant	Ground Water	Surface Soil	Subsurface Soil ^a	Surface Water ^a	Sediment ^a
Polynuclear Aromatic Hydrocarbons (PAHs)					
Acenaphthene		X	X		
Acenaphthylene		X			
Anthracene		X	X		
Benzo(a)anthracene		X	X		
Benzo(a)pyrene		X	X		
Benzo(b)fluoranthene		X	X		X
Benzo(ghi)perylene		X	X		
Benzo(k)fluoranthene		X	X		X
Chrysene		X	X		X
Dibenzo(a,h)anthracene		X	X		
Fluoranthene		X	X		X
Fluorene		X	X		
Indeno(1,2,3-cd)pyrene		X	X		
2-Methylnaphthalene			X		
Naphthalene		X	X		
Phenanthrene		X	X		X
Pyrene		X	X		X

X Contaminant is a COC which has been detected in the medium

^a Contaminants in shaded media did not result in a cancer risk greater than 10^{-6} nor a hazard index greater than one

1 3 1 Volatile Organic Compounds

Volatile organic compounds (VOCs) are present in subsurface soils and occur in some locations at high concentrations in groundwater ($> 10 \text{ mg/l}$) Chlorinated solvents occur sporadically and at low concentrations ($< 10 \text{ mg/kg}$) in subsurface soils throughout the IHSSs Sources for the high concentrations of these VOCs in groundwater have not been sampled but there is adequate circumstantial evidence to conclude that subsurface soils with high chlorinated solvent concentrations (sources) exist Toluene occurs throughout OU 1 in subsurface soils at relatively low concentrations however the source of the toluene is unknown The occurrence of toluene in the OU 1 samples may be a result of laboratory or field introduced contamination

Three general source areas for VOCs in groundwater at OU 1 have been identified (see Figure 1-4) Within these three general areas multiple release points appear likely based on concentration gradients and chemical fingerprints The three general source areas include

- South of Building 881
- IHSS 119 1 area
- Southeast of IHSS 119 2

These areas are discussed further in the following paragraphs For the sake of consistency the terms used to define these areas are used throughout the report The terms are used in place of IHSS designations because for the most part actual sources have not been conclusively determined for all of these contaminated groundwater areas In addition remediation measures would need to be evaluated for the "plumes" directly and could not be limited to individual IHSSs

Area South of Building 881

Groundwater in this area contains generally low concentrations of chlorinated solvents ranging from non-detects to $130 \text{ } \mu\text{g/l}$ as a maximum However, the spatial distribution of the detections is inconsistent and does not clearly indicate a discrete source In addition the description of the

historical activities at IHSSs 145 107, and 106 does not clearly indicate use or disposal of chlorinated compounds

The maximum detection of 130 $\mu\text{g}/\ell$ of 1 1 1 trichloroethane (1 1 1 TCA) in samples collected from well 1 87 may indicate IHSS 145 is a source. However the results of a soil gas survey presented in the previous Phase I RI Report revealed no 1 1 1 TCA in the soil gas sample collected closest to well 1 87 and thus do not corroborate the source

Soil gas survey results reveal a high concentration of tetrachloroethene (PCE) in soil gas approximately 30 feet southwest of well 5287 and is shown on Figure 1-4 as a suspected source area. This detection is the second highest out of several hundred soil gas samples collected at OU 1 suggesting a source for PCE in subsurface soils and the possible existence of residual dense non aqueous phase liquid (DNAPL) (DOE 1994a). The lack of PCE detections in groundwater samples collected from wells south of the soil gas detection (well 5487/5387) suggest that either the solvent release did not reach the water table (as a free phase wetting front) or that groundwater is not present at the location of the release

IHSS 119.1 Area

Documented waste storage practices at this IHSS resulted in the release of chlorinated solvents which now pose a continuing source for VOCs in groundwater. VOC concentrations are highest in the southwest portion of the IHSS. This fact coupled with the apparent presence of drummed waste as seen in historical aerial photographs permits approximate definition of the source area in the southwest portion of the IHSS. Within this source area, individual releases from drums cannot be resolved due to their apparent small areal extent. However the results of the Phase I soil gas survey suggest several locations which may represent the actual release points thus permitting an estimation of where DNAPL contaminants may have originated. A comparison of the chemical suite detected in groundwater at several locations within the drum storage area revealed at least two distinct chemical mixtures. One is dominated by trichloroethene (TCE) and

1 1 1 TCA (well 0974) and the other is dominated by carbon tetrachloride (CCl_4) (well 1074) which supports the multiple release point concept

Given the assumed release mechanism namely leaking drums on the ground surface it is reasonable to assume that gravity driven wetting fronts of chlorinated solvents may have advanced through the vadose zone and at least portions of the saturated zone. If consistent with solvent migration theories as the wetting front passed it may have left behind residual free phase chlorinated solvents in both the vadose and saturated zone. This residual would thus constitute a continuing source for VOCs in groundwater at this location.

A major finding of the Phase III RFI/RI was that chlorinated solvents released at IHSS 119 1 have likely entered the UHSU as DNAPLs. Phase III RFI/RI results suggest that the DNAPL zone is limited to the UHSU directly beneath IHSS 119 1. An aqueous plume of TCE, TCA, and several other VOCs has been defined as emanating from the DNAPL zone and has migrated along the preferential groundwater flow pathway. This plume is currently being intercepted by the french drain.

The migration of VOC contamination in groundwater downgradient of the western portion of IHSS 119 1 is currently limited by the french drain which acts as a hydraulic barrier in the UHSU. The historical maximum concentration of VOCs in groundwater is defined by well 4787 although in general well 4787 has had sporadic low level detections of VOC contaminants.

The extent of contamination originating from the unidentified source is unknown. Well 0687 was destroyed during the construction of the french drain. The french drain now captures all UHSU groundwater that once flowed through the area occupied by well 0687.

Area Southeast of IHSS 119.2

Concentrations of chlorinated solvents detected in two closely spaced monitoring wells downgradient of IHSS 119 2 (wells 6286 and 6386) are attributed to potential VOC release areas.

at both IHSS 119 2 and outside the operable unit The occurrences of these VOCs in groundwater within the IHSS include one time detections of 9.3 $\mu\text{g}/\ell$ in UHSU well 34791 and 0.1 $\mu\text{g}/\ell$ LHSU well 4587 Chloroform detections occurred three times in well 4587 with a maximum detection of 18 $\mu\text{g}/\ell$

Wells 6286 and 6386 contain contaminated groundwater and are located in a drainage hydraulically downgradient from IHSS 119 2 Therefore a VOC release point is suspected and is shown on Figure 1-4 based on the location of suspected waste disposal features depicted on aerial photographs The location and size of this suspected VOC release point is uncertain It is possible that contamination from the 903 Pad is also responsible for the VOCs detected in monitoring wells on the Hillside The 903 Pad is upgradient of the impacted wells and is known to be a source for CCL_4 and other dissolved chlorinated solvents in groundwater

The occurrence of chlorinated solvents in subsurface soils is limited to a maximum detection of 140 $\mu\text{g}/\text{kg}$ in one borehole (BH5887) The occurrence of VOCs in soil gas is limited to low levels of PCE and 1,1,1 TCA at one location within the IHSS However the magnitude of the soil gas detections is several orders of magnitude less than those noted at Building 881 and IHSS 119 1 and are more representative of the local background around IHSS 119 2 Nevertheless as was the case at IHSS 119 1 the presence of a VOC release point within IHSS 119 2 boundaries is suspected based on the downgradient groundwater chemistry

1.3.2 Metals

Metal contaminants include vanadium and selenium, both of which are significantly elevated in groundwater These elements are not elevated in surface or subsurface soils Although these substances were not reported to have been associated with wastes stored or disposed of at OU 1 they appear to be elevated primarily in areas where VOC wastes were stored at OU 1 It is postulated that these metals are undocumented constituents of wastes stored at IHSS 119 1 It is unlikely that they were leached from the soil by organic wastes disposed of at OU 1 since hydraulic oil and chlorinated solvents have poor chelation properties and are not strongly acidic

or basic. Nevertheless, the potential for leaching of these metals exists. Alternatively, these constituents may be naturally occurring; however, there is insufficient data to support either conclusion. Four areas have been identified at OU 1 with elevated selenium and/or vanadium as discussed below.

IHSS 119.1 Area

Multiple detections of selenium and vanadium were noted in monitoring wells located in the southwestern portion of the IHSS (Figure 1.5). Typically, the elevated metals were seen in association with VOCs. In particular, the highest metal concentration (2200 µg/l of Se) was detected in a well with one of the highest VOC concentrations anywhere at OU 1 (Well 1074). The maximum downgradient extent of selenium in groundwater at IHSS 119.1 appears to be in the vicinity of well 0487. The occurrence of vanadium is similar to selenium except that vanadium only occurs above background in UHSU wells.

Area South of Building 881

One detection of vanadium was noted at well 5387 at approximately six times the background level of 30 mg/l. This well exhibits concentrations of various chlorinated compounds in the 1 to 25 µg/l range. Several potential VOC source areas have been identified in the area south of Building 881, however, well 5387 is not particularly close to the suspected source areas. Nevertheless, it is conceivable that the vanadium present in groundwater at 5387 represents a plume originating from one of the VOC source areas previously discussed. The extent of vanadium concentrations above background near Building 881 appears to be limited to the immediate vicinity around well 5387.

Area East of IHSS 102

One detection of vanadium and three detections of selenium were noted above the background level in well 6986. No detections of VOCs have been noted at this well. It is unclear whether

these detections represent contamination or naturally occurring levels as the maximum vanadium and selenium concentrations represent 126 percent and 194 percent of background respectively. Based on these relatively low levels, a contaminant source is not suspected in this area.

Southeast Corner of IHSS 130

Vanadium is the only contaminant detected at this location over background levels. A maximum of 403 $\mu\text{g}/\ell$ was detected at well 37191, which represents approximately five times the background level. Only exceedingly low levels of VOC contamination ($< 0.5 \mu\text{g}/\ell$) were found in association with the vanadium. The extent of vanadium and selenium contamination in the southeast corner of IHSS 130 appears to be limited to the immediate vicinity around well 37191.

1.3.3 Semivolatile Organic Compounds

The only semivolatile organic compounds (SVOCs) that are identified contaminants at OU 1 are PAHs and PCBs. Although PAHs are considered to be OU 1 contaminants in the Phase III RFI/RI, they are not considered to be of OU 1 origin. PAHs occur over most of OU 1 in surface soils and tend to decrease in concentration with depth. PAHs have also been detected in sediments. Several areas of OU 1 have been identified where PAHs appear more concentrated relative to the surrounding area. The areas do not coincide with IHSS locations (see Figure 1.6). The sources for the PAHs at OU 1 are presumed to be general urban fallout including asphalt dust and larger particles, vehicle exhaust, and furnace exhaust.

1.3.4 Polychlorinated Biphenyls

PCB occurrence is restricted to IHSS 119.1 and 119.2 surface and subsurface soils (Figure 1.7). One PCB detection has been also noted in sediments. However, sediments are not addressed as part of the OU 1 CMS/FS, and in addition, this detection was at the western OU 1 boundary and is not considered of OU 1 origin. The contaminant release mechanism for PCBs is unknown.

1 3 5 Radionuclides

Americium, plutonium, and uranium have been identified as OU 1 contaminants and are elevated in surface and subsurface soil. In addition, plutonium and americium are evaluated in surface water and sediment. The widespread plutonium and americium contamination appears to be a result of deposition of wind-disseminated plutonium/americium-contaminated dust originating from the 903 Pad Area. A general decrease in activities is noted from east to west ranging from a maximum of 22.7 pCi/g to 0.0076 pCi/g of plutonium and 4.15 pCi/g to 0.0129 pCi/g of americium (see Figure 1.8).

In contrast to the wide-spread plutonium/americium contamination, localized hotspots are present at OU 1 that are markedly contaminated with either plutonium/americium or uranium. These "hotspots" are postulated to have arisen from releases of radionuclide-contaminated liquids stored in drums at OU 1 and are being addressed through an early removal action discussed later.

Unlike plutonium and americium, uranium contamination is not wide-spread, although it is significantly elevated at discrete locations in surface and subsurface soils at OU-1. Uranium was below background levels at SS100393, slightly above background at SS100493 and significantly above background at SS100193 and SS100293. The low levels at SS100493 coupled with uranium 233/uranium 238 ratios of approximately 1 to 2 suggest the uranium may be naturally occurring. The highest activities of uranium at SS100193 and SS100293 occur just beneath the surface as the deeper composites have the higher activities. The maximum total uranium activity at SS100193 is approximately 550 pCi/g with a uranium 233/uranium 238 activity ratio of 3.5. This suggests contamination with enriched uranium. The maximum total uranium activity at SS100293 is approximately 240 pCi/g with an activity ratio as high as 160. This suggests contamination with uranium-233 as the activity ratio far exceeds that for enriched uranium.

1 4 Fate and Transport of Contaminants

This section discusses potential mechanisms by which contaminants identified in the Phase III RFI/RI can migrate. Although several mechanisms are identified in the following sections, the groundwater medium is the most significant pathway. Figure 1 9 depicts potential groundwater migration pathways. Note that this figure does not include the volume of groundwater available for transport. Many areas of OU 1 are currently dry and remain dry throughout the year. The migration pathways presented in the figure merely present potential pathways assuming adequate groundwater is present.

1 4 1 Volatile Organic Compounds

The release mechanisms for VOCs at OU 1 are varied including pure product leakage from stored drums, possible leakage of dilute aqueous solutions of VOCs from pipelines, and seepage of aqueous VOC solutions or pure product from impoundments and disposal pits. In the area south of Building 881, the release mechanisms likely to have occurred include leaking pipelines and leakage from impoundments and disposal pits. In the western portion of OU 1 (IHSS 119 1), the release mechanism is most likely leakage from drums stored on the land surface.

Once the contaminant has entered the subsurface, the pathways for VOC migration include gravity driven wetting fronts of aqueous solutions and/or small volumes of pure product through the vadose zone to the water table. In the case of pure product, the density of the pure chlorinated solvent would allow the contaminant to migrate vertically through the saturated zone. The migration as pure product would be arrested once the wetting front of contamination became depleted by the process in which residual product is retained by soils and rock. Alternatively, the migration would stop once the pure product came to rest in a topographic low on an impermeable surface (possibly the Laramie claystone). At this point, migration would continue in the form of an aqueous phase hydrocarbon plume (if groundwater is present). Precipitation and infiltration would also contribute to VOC migration as pure chlorinated solvents are dissolved and transported downward by infiltrating snowmelt and rainwater.

The dissolved phase plume would migrate with the groundwater being retarded to varying degrees as a function of the physical and chemical properties of the contaminant geologic materials and groundwater. In the case of OU 1 the organic contaminants identified in the Phase III RFI/RI report are primarily retarded by the clayey materials in the subsurface environment. This is due to the relatively low organic carbon content of the soils found in OU 1. Retardation is particularly significant for OU 1 contaminants with high K_{ow} values like CCL_4 (DOE 1994a).

At OU 1 the shallow groundwater, which carries most of the contamination, is controlled to a large degree by the topography of the bedrock surface. Active channels in the bedrock are covered by unconsolidated material of varying thickness that is variably saturated. Typically groundwater will flow towards the axis of the bedrock channel and continue downgradient along the axis of the channel potentially to the Woman Creek Alluvium. Therefore at OU 1 an aqueous phase hydrocarbon plume in groundwater has the potential to discharge to Woman Creek although this is not likely due to the low initial volume of contaminants of concern (COCs) available for transport. However the existing french drain acts as a hydraulic barrier preventing the discharge of contaminated groundwater in the western and central portions of OU 1 to Woman Creek. In the eastern portion of OU 1 the potential exists for continuous contaminant migration pathways in groundwater from the suspect source areas to Woman Creek. However conclusive evidence of this occurrence has not been found, and the COC concentrations found to date limit the amount of contamination available for transport.

VOC-contaminated groundwater may also discharge to surface water through seeps which have historically been observed at OU 1 (DOE 1994a). While VOCs in surface water have been previously detected in the SID the recent construction of the french drain has intercepted this pathway.

Other migration pathways for VOCs include volatilization of pure product into soil gas and subsequent migration of soil gas laterally and vertically away from the source area. VOCs can

also partition out of contaminated groundwater into soil gas or desorb from organic matter into the soil gas

VOCs would not be expected to migrate in significant quantities through surface water or wind transport of VOC contaminated surface soil. This is based on the assumption that VOCs would quickly volatilize from the respective media. One apparent exception to this is the occurrence of toluene in OU 1 surface soils. Although there is no evidence to suggest that toluene is migrating through surface water or wind, it apparently is persistent in near surface soils despite its relative high volatility.

1.4.2 Metals

The mechanism for the release of metal contaminants into the environment is less clear than for VOCs. It is presumed that selenium and vanadium are undocumented RFETS wastes that were associated with the VOC wastes stored and disposed of at OU 1. It is unlikely that they were leached from the soil by organic wastes disposed of at OU 1 since hydraulic oil and chlorinated solvents have poor chelation properties and are not strongly acidic or basic. Nevertheless, the potential for leaching of these metals exists. Alternatively, these constituents may be naturally occurring, however, there is insufficient data to support either conclusion. In either case, the primary migration pathway is as a dissolved phase contaminant plume in groundwater. This migration pathway was previously presented for VOCs.

1.4.3 Semivolatile Organic Compounds

It is presumed that PAHs were deposited at OU 1 as fallout of combustion products or wind blown asphalt dust. Asphalt dust and larger particles may also have been transported and deposited by vehicles traversing OU 1 or by disposal of asphalt waste at OU 1.

Once in place, the dispersion mechanisms include vertical migration by infiltrating surface water carrying small particles composed of PAHs. The low solubility and high K_{oc} values of PAHs

precludes mobilization of significant quantities in the dissolved form therefore transport via groundwater is not significant Other transport mechanisms include surface water and wind transport of particulate

1 4 4 Polychlorinated Biphenyls

Transport mechanisms relevant for PCBs are similar to those for PAHs however the source areas for PCBs are more discrete than for PAHs PCBs are expected to be very immobile given the high k_{oc} values and the high carbon and clay content in surface soils at OU 1 Adsorption of PCBs at OU 1 is expected to be substantial on soils and clay particles (DOE 1994a)

1 4 5 Radionuclides

Transport mechanisms relevant to radionuclides are similar to PAHs In particular plutonium has a strong affinity for the solid phase and will not be readily mobilized by precipitation and infiltration Plutonium is strongly adsorbed to clay particles and is expected to undergo strong cation-exchange reactions due to its strong positive charge (DOE 1994a) The primary transport mechanism for plutonium is wind dispersion

1 5 Baseline Risk Assessment

The OU 1 Baseline Risk Assessment (BRA) consists of both a public health evaluation and an environmental evaluation The primary purpose of each evaluation is to examine the current and future risks associated with contaminants identified during the analysis of the nature and extent of contamination The following subsections summarize each evaluation and provide an overall summary of the risks associated with OU 1

1 5 1 Public Health Evaluation

During the course of the Public Health Evaluation (PHE) site population and land use data were analyzed in order to devise several representative exposure scenarios (potentially exposed receptors) for assessing the risk to current and future human health from identified contaminants at the 881 Hillside Area. For each of these scenarios pathways were traced which represented exposure routes from the source to potential receptors.

Pathway elements were examined relative to the results of the Phase III field investigation which indicated that contamination exists in groundwater, surface soils, subsurface soils, sediments and surface waters. The contaminants identified in these areas included VOCs, PAHs, PCBs, inorganic contaminants, and radionuclides. The contaminant release mechanisms evaluated were leaching, volatilization, resuspension of particulates by wind, etc. Potential transport media identified were surface water, groundwater, air, soil, and biota. The exposure route (the route of entry into the human body) for these media included ingestion, inhalation, and dermal contact. In accordance with the *Risk Assessment Guidance for Superfund Volume I Human Health Evaluation Manual (Part A)* (EPA 1989a), if any of the above-mentioned pathway elements is missing, the projected receptor will not receive a chemical or radionuclide dosage and no excess risk will exist from that contaminant.

The OU 1 physical environment, including the french drain and treatment system, was used with information about the potentially exposed population, land use scenarios, and exposure pathways to form the conceptual site model. This was evaluated to identify complete pathways for credible and plausible exposure scenarios. The following describes the specific land use scenarios and pathways selected with the conceptual site model for quantitative assessment:

- Current Off Site Resident
 - Inhalation of airborne particulates
 - Soil ingestion (following deposition of particulates on residential soil)

- Dermal contact with soil (following airborne deposition of particulates)
- Ingestion of homegrown vegetables/fruit (following surface disposition and uptake of particulates)
- **Current On Site Worker**
 - Inhalation of airborne particulates
 - Soil ingestion
 - Dermal contact with soil
 - Sediment ingestion
 - Dermal contact with sediment
 - Surface water ingestion
 - Dermal contact with surface water
- **Future On Site Worker**
 - Inhalation of VOCs in endower air (office worker only) and outdoor air (construction worker only)
 - Inhalation of airborne particulates
 - Soil ingestion
 - Dermal contact with soil
 - Sediment ingestion (office worker only)
 - Dermal contact with sediment (office worker only)
 - Surface water ingestion (office worker only)
 - Dermal contact with surface water (office worker only)
- **Future On Site Ecological Researcher**
 - Inhalation of airborne particulates
 - Soil ingestion
 - Dermal contact with soil
 - Sediment ingestion
 - Dermal contact with sediment
 - Surface water ingestion
 - Dermal contact with surface water
- **Future On-Site Resident**
 - Inhalation of indoor VOCs from basement vapor
 - Inhalation of particulates
 - Soil ingestion
 - Dermal contact with soil
 - Sediment ingestion
 - Dermal contact with sediment
 - Surface water ingestion

- Dermal contact with surface water
- Ingestion of homegrown vegetables/fruit (following surface deposition of particulates and uptake)

In addition four special cases of the on site residential scenarios were included to show the impact of the use of groundwater and to evaluate risk at the source. It should be noted that groundwater yield was examined through several UHSU well simulations as part of the BRA. These simulations indicated that the yield of contaminated groundwater in IHSS 119 1 is inadequate to support a household of four people (see Appendix F).

The results of the BRA indicate that only the media of groundwater and surface soils present a risk greater than the acceptable risk range of 10^{-4} to 10^{-6} . The risk to a human receptor from exposure to groundwater contaminants of concern (COCs) is driven primarily by the exposure routes of ingestion, inhalation of volatiles, and dermal contact. For a future on site resident this risk is on the order of 10^{-3} to 10^{-2} but applies only to exposures occurring directly at IHSS 119 1. Risk results excluding this source location are much lower for groundwater.

The risk to a human receptor from exposure to surface soil COCs is driven primarily by the exposure routes of ingestion of vegetables, ingestion of soil, inhalation of particulates, and dermal contact. For a future on site resident this risk is on the order of 10^{-3} . It should be noted, however, that this risk is based on OU 1 sitewide average radionuclide concentrations. These average radionuclide concentrations include a few areas of high contaminant concentrations (i.e., hotspots) that are limited in extent and only exist within the boundaries of IHSSs 119 1 and 119 2. These hotspots are currently scheduled for remediation under an early removal action for OU 1 and will be remediated to measured background concentrations. The risk to a future on site resident, excluding the hotspots, is on the order of 10^{-5} . Risk results are summarized in Tables 1-3 and 1-4.

Table 1 3
Summary of OU 1 Point Estimates of Carcinogenic Risk

Scenario	Total Excess Cancer Risk	Dominant COC*	Dominant Pathway
Current			
On Site Worker (Security Specialist)	1×10^{-4}	Plutonium 239 240	Inhalation of dust
Off Site Resident (Adult)	2×10^{-6}	Plutonium 239 240	Inhalation of dust
Standard Future			
Future On Site Worker (Office)	2×10^{-3}	Plutonium-239 240	Inhalation of dust
Future On Site Worker (Construction)	4×10^{-7}	1 1 Dichloroethene	Inhalation of volatiles
On Site Ecological Researcher	2×10^{-3}	Plutonium-239 240	Inhalation of dust
On Site Resident (Adult)	3×10^{-3}	Plutonium-239 240	Inhalation of dust
Other Future			
On Site Resident (Adult) (Sitewide With Groundwater)	6×10^{-3}	1 1 Dichloroethene	Ingestion of groundwater
On Site Resident (Adult) (Assuming Adequate Groundwater At Source)	7×10^{-2}	1 1 Dichloroethene	Ingestion of groundwater
On Site Resident (Adult) (Groundwater At Source With Public Water)	4×10^{-2}	Plutonium 239 240	Inhalation of dust
On Site Resident (Adult) (Without Source/Without Groundwater)	5×10^{-5}	Dibenzo(a h)anthracene	Ingestion of vegetables

Plutonium concentrations are biased high by the presence of several hotspots which are currently being evaluated for removal. Upon removal of the hotspots, the dominant surface soil COC is no longer plutonium for those areas where the radionuclide hotspots drive the risk. In these cases risks from surface soils will be approximately 5×10^{-5} as calculated for the on-site resident scenario without the source.

Table 1-4
Summary of OU-1 Point Estimates of Noncarcinogenic Risk

Scenario	Total Hazard Index		Dominant COC	Dominant Pathway
	Child	Adult		
Current				
On Site Worker (Security Specialist)	N/A	8 x 10 ⁻⁵	Pyrene	Dermal contact with soil
Off Site Resident	1 x 10 ⁻⁷	6 x 10 ⁻⁸	Fluorene	Ingestion of vegetables
Standard Future				
Future On-Site Worker (Office)	N/A	3 x 10 ⁻³	1 1 1 Trichloroethane	Inhalation of volatiles through foundation
Future On Site Worker (Construction)	N/A	1 x 10 ⁻⁴	1 1 1 Trichloroethane	Inhalation of volatiles during excavation
On Site Ecological Researcher	N/A	2 x 10 ⁻³	Pyrene	Dermal contact with soil
On Site Resident	2 x 10 ⁻²	5 x 10 ⁻³	1 1 1 Trichloroethane	Inhalation of volatiles through foundation
Other Future				
On Site Resident (Sitewide With Groundwater)	2 x 10 ⁺¹	9 x 10 ⁰	Carbon Tetrachloride	Ingestion of groundwater
On-Site Resident (Assuming Adequate Groundwater At Source)	3 x 10 ⁺²	1 x 10 ⁺²	Carbon Tetrachloride	Ingestion of groundwater
On Site Resident (Groundwater At Source With Public Water)	3 x 10 ⁺¹	1 x 10 ⁺¹	Carbon Tetrachloride	Ingestion of groundwater
On Site Resident (Without Source/Without Groundwater)	7 x 10 ⁻³	3 x 10 ⁻³	Fluorene	Ingestion of vegetables

1 5 2 Environmental Evaluation

As part of the overall BRA an environmental evaluation (EE) conducted to ascertain whether contamination resulting from RFETS activities in OU 1 may have impacted or could adversely impact ecological receptors in the vicinity where ecological receptors are operationally defined as plants and animals other than humans and domesticated species

COCs were selected for the EE based on a comparison of maximum concentrations of OU 1 contaminants to benchmark values COCs identified in the EE include VOCs PAHs PCB radionuclides and selenium The EE evaluated the impact that these COCs had on the following endpoints

- Vegetative Community
- Small Mammal Community
- Mule Deer Population
- Toxic Exposure to Top Predators

The results of the EE indicate that the concentrations of VOCs in groundwater and PAHs and PCBs in soils are potentially toxic to ecological receptors however the restricted distribution of these contaminants limits the duration and frequency of contact with receptors and therefore limits exposures

1 5 3 Risk Summary

As indicated by the PHE portion of the BRA risks to human receptors at OU 1 are primarily associated with exposure to groundwater COCs Although this medium is not available for current residential use this scenario presents the highest and only, unacceptable risk per the NCP guideline of 10^{-4} to 10^{-6} Environmental risks currently have not been identified by the Phase III RFI/RI and therefore do not warrant further examination

OU 1 risks are a result of widespread contamination found in low concentrations and in various media throughout the site. The Phase III RFI/RI results indicate that for the most part individual IHSSs cannot be associated directly with any one contaminant group or area. Table 1.5 lists the primary contaminants present at each IHSS and summarizes how these contaminants will be addressed in the CMS/FS.

1.6 Interim Measures/Interim Remedial Actions

The IM/IRA that was completed for OU 1 consists of a french drain designed to collect contaminated alluvial groundwater from the operable unit and to prevent further downgradient migration of contaminants. The IM/IRA included a geotechnical investigation that was performed in order to evaluate the site characteristics along the proposed french drain alignment (EG&G 1990). Construction of the french drain began in November 1991 and was completed in April 1992. The water treatment plant located in Building 891 is part of the IM/IRA and will be converted to sitewide uses. Hereinafter this plant is referred to as the Building 891 water treatment plant.

The french drain was constructed by excavating a trench approximately 1,435 feet in length (DOE 1994a). The trench was keyed into bedrock material that exhibited a hydraulic conductivity on the order of 1×10^{-6} cm/sec. A permeable membrane was placed on the upgradient side of the drain and an impermeable polyvinyl chloride membrane was placed on the downgradient side of the drain. A perforated pipe was placed along the drain to collect groundwater, and the drain was backfilled with gravel and then soil. Currently, groundwater collected from the drain is fed into an ultraviolet and hydrogen peroxide (UV/H₂O₂) treatment unit for treatment of organic compounds. Inorganic contaminants are removed via a series of ion exchange columns.

An additional removal action is planned for OU 1 to remove surface soil radionuclide hotspots identified during the Phase III surface soil investigation. The action is documented in a Proposed Action Memorandum (PAM) and is intended to eliminate the plutonium and uranium

**Table 1-5
Summary of Primary IHSS Contaminants**

IHSS Number	Primary Contaminants^a	Disposition
102	Groundwater contaminated with PCE and TCE	Considered in Building 881 Area
103	Possible groundwater and subsurface soils contaminated with low levels of PCE and TCE	Considered in Building 881 Area
104	Potential toluene in subsurface and groundwater wide array of PAHs	Not identified as a source no action required
105 1 & 105 2	Low levels of VOCs in groundwater PCE detected below detection limit, potential solvent contamination in soils at north end	Considered in Building 881 Area although not identified as a source
106	Groundwater contaminated with chlorinated solvents potential solvent contamination in soils at north end	Considered in Building 881 Area although not identified as a source
107	Groundwater contaminated with chlorinated solvents	Considered in Building 881 Area although not identified as a source
119 1 & 119 2	Groundwater contaminated with chlorinated solvents and selenium possible DNAPL sources in subsurface radionuclide hotspots	Considered under IHSS 119 1 and Area East of 119 2
130	Radionuclide-contaminated soil and asphalt PAHs in subsurface soils	No risk pathway for rads and PAHs in subsurface soils no action required Not identified as a source of VOCs
145	Groundwater contaminated with chlorinated solvents potential low level rad contamination	Considered in Building 881 Area although not identified as a source

^a Radionuclide and PAH contamination in near surface soils is not identified in this table due to the widespread and consistent nature of the contamination, indicating a source outside of OU 1 and unrelated to OU 1 disposal activities

hotspots that are currently evaluated as part of the sitewide risk. This removal action is scheduled to be completed prior to completion of the proposed plan for OU 1. For the purposes of alternative development it will be assumed that the hotspots are not present. This assumption lowers the sitewide risk from surface soil contaminants to 10^{-5} and below for all exposure scenarios.

2 0 IDENTIFICATION AND SELECTION OF TECHNOLOGIES AND REPRESENTATIVE PROCESS OPTIONS

This section summarizes the results of the identification screening evaluation and selection of technologies and representative process options used in the development of remedial action alternatives for OU 1. Information on how these activities are conducted is included in the CERCLA RI/FS guidance. In general, the guidance identifies the following steps for selecting representative process options:

- Develop media specific RAOs
- Develop media-specific GRAs
- Identify volumes and/or areas of the media which require GRAs
- Identify and screen technologies and process options applicable to each GRA
- Evaluate process options within each technology type to select a representative option for developing remedial action alternatives

This section summarizes how these steps were applied to OU 1. Originally two technical memoranda were prepared to seek input from the regulatory agencies on RAOs, preliminary remediation goals (PRGs), and the alternative development process. A final version of TM #10 and a draft final version of TM #11 were submitted to the agencies in April 1994. Comments were received on both documents and were incorporated in this report where appropriate. The technical memoranda will not be resubmitted but are available for review in the administrative record.

Elements from both technical memoranda are included in this section and in Section 3.0, particularly where specific comments are being addressed. However, the detailed calculations involved in estimating PRGs and the screening and evaluation of technologies and process options are not included in this report. This information is presented in TMs #10 and #11 and is summarized in this section and Section 3.0.

2 1 Contaminants of Concern

The list of contaminants originally identified in the Phase III RFI/RI is presented in Section 1 0 of this report. Potential contaminants identified early in the RFI/RI process were subjected to a multi level screening process that identified COCs for inclusion in the PHE and EE. The screening process shortened the list of potential contaminants to consider further as risk contributors. The process is presented in detail in the Phase III RFI/RI report. Contaminants that survived the risk based screening process were designated as COCs in the BRA.

The PHE and EE present the results of COC screenings that were performed to identify potential risk contributors to human and ecological receptors respectively. The COCs identified in the EE were the following:

- carbon tetrachloride
- 1 1 dichloroethene
- tetrachloroethene
- 1 1 1 trichloroethane
- trichloroethene
- toluene
- selenium
- PAHs
- PCBs
- americium
- plutonium
- uranium

Because these contaminants do not contribute a significant risk to ecological receptors and no adverse impacts are currently identified in the EE, they are not evaluated separately in this report.

However, groundwater COCs identified in the PHE are a potential concern at OU 1. Risks associated with exposure to these COCs exceed 10^{-4} at IHSS 119.1. The following COCs were identified for groundwater:

- carbon tetrachloride
- 1,1-dichloroethene
- tetrachloroethene
- 1,1,1-trichloroethane
- selenium

Surface soil COCs were also identified in the PHE including PAHs, PCBs, and radionuclides. However, radionuclide contaminants associated with windblown dispersion of OU 2 contaminants (from the 903 Pad) are not addressed in this report. The radionuclide contaminants consist primarily of low concentrations of americium and plutonium spread across several operable units. These radionuclides must be addressed as a whole through the medium and source where they originate. Therefore, in order to develop appropriate remedial action alternatives, the administrative transfer of radionuclide contaminants to OU 2 is currently in progress.

In addition, several radionuclide hotspots were identified in OU 1 as risk contributors. These hotspots are currently scheduled for remediation under a removal action and are therefore not considered for alternative development.

PAHs and PCBs are the only COCs remaining for evaluation in surface soils. Excluding radionuclides, the highest risk associated with PAHs and PCBs in OU 1 surface soils is on the order of 10^{-5} . This is similar to background risks from PAHs in urban areas and is within the acceptable risk range specified in the NCP (10^{-4} to 10^{-6}). Because PAHs and PCBs in surface soils do not present an unacceptable risk and cannot be physically isolated from the radionuclides in surface soils being addressed under OU 2, these contaminants are not included in the development of remedial action alternatives.

As discussed in Section 1.0, the BRA indicates that no media other than groundwater and surface soil result in a risk greater than 10^{-6} , nor do they result in adverse impacts to environmental receptors. In addition, surface water and sediments evaluated in the BRA are being addressed through OU 5. For these reasons, no COCs are identified for any media except groundwater.

and surface soils. Subsurface soil COCs are identified as a possible source of groundwater contamination and are addressed through the groundwater COCs, RAOs, and GRAs.

2.2 Summary of Remedial Action Objectives

RAOs were developed using appropriate regulatory guidelines (i.e., CERCLA RI/FS guidance and the NCP) and by examining the COCs identified in Section 2.1 and their associated exposure pathways. Briefly, the RAOs for OU 1 are the following:

- 1) Prevent the inhalation of, ingestion of, and/or dermal contact with VOCs and inorganic contaminants in groundwater that would result in a total excess cancer risk greater than 10^{-4} to 10^{-6} for carcinogens and/or a hazard index greater than or equal to one for non-carcinogens.
- 2) Minimize further degradation of groundwater beneath OU 1 by eliminating and/or containing residual subsurface soil DNAPLs to the maximum extent practicable.
- 3) Prevent the inhalation of, ingestion of, and/or dermal contact with PAHs, PCBs, and radionuclides in surface soils that would result in a total excess cancer risk greater than 10^{-4} to 10^{-6} for carcinogens and/or a hazard index greater than or equal to one for non-carcinogens.
- 4) Prevent exposure to carcinogenic radionuclides in surface soil hotspots that would result in an excessive short-term risk to a human receptor.

These RAOs are used to determine what area or areas of OU 1 require remedial action evaluation and are quantified through the use of PRGs. The third and fourth RAOs listed above are already being addressed through the OU 1 surface soil hotspot removal action. This action, along with the administrative transfer of other radionuclides in surface soils to OU 2, will result in a residual risk level within the acceptable risk range of 10^{-4} to 10^{-6} . Therefore, the focus of this report is on meeting the first and second RAO, which are concerned with groundwater.

2 3 Development of Preliminary Remediation Goals

PRGs are generally identified through use of readily available information such as chemical specific ARARs or other reliable information (EPA 1990a) Where ARARs or to-be considered (TBC) criteria are not available PRGs are developed on the basis of a 10^{-6} point of departure risk for each chemical within a given medium This also applies when ARARs are not considered sufficiently protective because of the presence of multiple contaminants or multiple pathways of exposure Note that PRGs developed at this stage are considered initial goals which may be revised through the course of the CMS/FS The following sections present the sources of information used for identifying appropriate PRGs for OU 1 both chemical specific ARARs and risk-based cleanup goals Existing potential OU 1 chemical specific ARARs are currently the basis for alternative development

2 3 1 Potential Chemical Specific Applicable or Relevant and Appropriate Requirements

CERCLA Section 121(d)(2) provides a statutory basis for determining ARARs in a remedial action context concerning hazardous substances, pollutants, or contaminants that will remain on site

If any standard requirement criteria or limitation under any federal environmental law or any [more stringent] promulgated standard requirement criteria or limitation under a state environmental or facility siting law is legally applicable to the hazardous substance concerned or is relevant and appropriate under the circumstances of the release or threatened release of such hazardous substance pollutant or contaminant the remedial action shall require at the completion of the remedial action a level or standard of control for such hazardous substance pollutant or contaminant which at least attains such legally applicable or relevant and appropriate standard requirement criteria or limitation [42 United States Code (USC) ----§ 9621(d)(2)]

where applicable requirements" are those

cleanup standards standards of control or other substantive environmental protection requirements criteria or limitations promulgated under federal environmental or state environmental or facility siting laws that specifically address a

hazardous substance pollutant contaminant remedial action location or other circumstance found at a CERCLA site Only those state standards that are identified by a state in a timely manner and that are more stringent than federal requirements may be applicable

According to the NCP and the *CERCLA Compliance with Other Laws Manual* (EPA 1988b)

Relevant and appropriate requirements are those cleanup standards standards of control, and other substantive requirements criteria, or limitations promulgated under federal environmental or state environmental or facility siting laws that while not applicable to a hazardous substance pollutant contaminant remedial action location or other circumstance at a CERCLA site address problems or situations sufficiently similar to those encountered at the CERCLA site so that their use is well suited to the particular site Only those state standards that are identified in a timely manner and are more stringent than federal requirements may be relevant and appropriate

Chemical specific ARARs were identified in accordance with CERCLA guidance and the requirements of the NCP [see 40 Code of Federal Regulations (CFR) Part 300 Subsection 430(e)(2)(i)] Chemical specific requirements under a variety of Federal and state laws were reviewed to identify potential groundwater and surface soil chemical specific ARARs

Current Groundwater Classification

The Colorado Water Quality Control Commission (CWQCC) designated the Quaternary and Rocky Flats Aquifers beneath the RFETS as domestic use quality agricultural use quality and surface water protection according to 3 12 7 of 5 Colorado Code of Regulations (CCR) 1002 8 Subsection 3 12 7 The intent of these classifications is to protect specified groundwater from uncontrolled degradation and thereby protect existing and future uses of groundwater (5 CCR 1002 8 Subsection 3 11 9) Furthermore groundwater is classified domestic use or agricultural use quality if

the groundwater is either used or reasonably likely to be used for domestic or agricultural purposes within the specified area or

the most recent State Engineer s well records or applicable court decrees reveal that

groundwater is permitted or decreed for such uses within the specified area (5 CCR 1002
8 Subsection 3 11 4)

The Phase III RFI/RI report does not support the CWQCC conclusion that there is groundwater beneath OU 1 which could be used as a drinking water supply. Included in the Phase III RFI/RI report are water production capability simulations and well production tests which conclude that neither the Rocky Flats Alluvium nor the Arapahoe Aquifers beneath OU 1 is capable of producing sufficient water for even domestic purposes. In addition, a letter from the Office of the Colorado State Engineer confirms that the conclusion that neither aquifer (referring to the Rocky Flats Alluvium and the Arapahoe Aquifer) is a potential source for domestic water supplies in the 881 Hillside area is valid when considering future land use (see Appendix F). The Colorado State Engineer's letter dated March 12, 1992, is in reference to the water production capability simulations and well production tests that were included in the Phase III RFI/RI report (see Appendix F).

DOE may petition the CWQCC when appropriate to consider changing the water quality classification beneath OU 1. Documentation of potential water use and quality of water in the Quaternary and Rocky Flats Aquifers beneath the site will be presented to the Commission for reconsideration of the current use classifications.

Potential Groundwater ARARs

The groundwater beneath the RFETS is currently classified for domestic use quality. TM #10 listed Federal maximum contaminant levels (MCLs) as the chemical specific ARARs for OU 1. CDPHE commented on TM #10 that the State's MCL standards should be ARARs. The State implements the Federal Safe Drinking Water Act (SDWA) through its drinking water program; therefore, state drinking water standards are presented in Table 2.1 as potential chemical specific ARARs for OU 1. The Superfund Amendments and Reauthorization Act (SARA) and the NCP rules require classification of non-zero maximum contaminant level goals (MCLGs) as potential chemical specific ARARs. Federal non-zero MCLGs for OU 1 contaminants are the same as the State MCLs listed in Table 2.1.

Table 2-1
Potential Groundwater Chemical Specific ARARs
State Drinking Water Standards^a
(µg/l)

Chemical ^b	State MCL
Volatile Organic Compounds	
Carbon Tetrachloride	5
Chloroform (total trihalomethanes)	<100
1 1 Dichloroethane	N/A
1 2 Dichloroethane	5
1 1 Dichloroethene	7
1 2 Dichloroethene	N/A
cis 1 2 Dichloroethene	70
Tetrachloroethene	5
Toluene	1 000
Total Xylenes	10 000
1 1 1 Trichloroethane	200
1 1 2 Trichloroethane	5
Trichloroethene	5
Metals	
Selenium	50
Vanadium	N/A

^a From CRS 25-1 107 25-1 108 25-1 109 and 25-1 114

^b All contaminants originally identified by the Phase III RFI/RJ are listed

^c Federal MCLs are numerically equivalent to these State MCLs

State groundwater standards are identified in Table 2.2 and are TBCs for OU 1. The standards were evaluated against the definition of ARARs in the NCP (40 CFR 300.5). The state groundwater standards are not assessed ARARs because the classifications requiring those standards have not been applied consistently throughout the state and thus fail the NCP criteria of general applicability in 40 CFR 300.400(g)(4).

Potential Surface Soil ARARs

Soil chemical specific ARARs requirements under State and Federal laws do not exist for the contaminants identified in OU 1 (i.e., there are no established protective levels for surface soil contamination based on risks to human health and/or the environment) with the exception of PCBs under Toxic Substances Control Act (TSCA). However, TSCA requirements for pre-1987 cleanups are determined on a case by case basis depending on the potential for contamination. Spills after 1987 define clean soil as containing less than 1 mg/kg PCBs (40 CFR 761.120 and 761.125). The concentrations of PCBs found in OU 1 are below this concentration except in one instance which is at a 1.2 mg/kg concentration. Therefore, it was considered to be within an acceptable range of determining a cleanup level. Accordingly, no chemical specific ARARs are identified for this medium.

2.3.2 Preliminary Remediation Goals Based on 10^{-6} as the Point-of Departure

In TM #10, both groundwater and surface soil PRGs were estimated for OU 1. Other media involving surface water and sediments were not considered for PRG development as part of the OU 1 CMS/FS since they, as well as subsurface soils, do not present a direct risk greater than 10^{-6} nor a hazard index greater than one and therefore do not warrant risk based PRGs.

Groundwater and surface soil PRGs were estimated in TM #10 for the following exposure scenarios:

- Future On Site Resident
- Commercial/Industrial Workers

Table 2 2
Statewide and Basin-Specific Groundwater Standards^a
($\mu\text{g}/\ell$)

Chemical ¹	State Standard	Practical Quantitation Limit
Volatile Organic Compounds		
Carbon Tetrachloride	0.3	1.0
Chloroform	6/0 19 ^b	1.0
1,1-Dichloroethane	—	—
1,2-Dichloroethane	0.4	1.0
1,1-Dichloroethene	7	1.0
1,2-Dichloroethene	—	—
cis-1,2-Dichloroethene	70 ^c	1.0
Tetrachloroethene	5/0 8b ^b	1.0
Toluene	1,000	1.0
Total Xylenes	10,000 ^c	1
1,1,1-Trichloroethane	200	1.0
1,1,2-Trichloroethane	3/0 6 ^b	1.0
Trichloroethene	5	1.0
Metals		
Selenium	10 ^d /20 ^e	—
Vanadium	100 ^e	—

- ¹ All contaminants originally identified by the Phase III RFI/RI are listed
- ^a CDPHE/Water Quality Control Commission Basic Standards for Groundwater 3 11 0 effective 3/30/94
- ^b CDPHE/Water Quality Control Commission Classification and Water Quality Standards for Groundwater 3 12 0 effective 1/31/94
- ^c Listed as drinking water MCL in state groundwater standards Table A
- ^d Measured as a dissolved concentration
- ^e Agricultural standard

The ecological reserve scenario was not used for estimating PRGs because the scenario does not apply to groundwater and is identical to the commercial/industrial scenario in terms of exposures for surface soils. Tables 2.3 and 2.4 present the risk based PRGs estimated for groundwater and surface soils respectively (for the COCs identified in the BRA). Table 2.5 compares the groundwater PRGs identified using chemical specific and risk based PRGs. For the purposes of OU 1, state MCLs are currently used to evaluate remedial action. Risk based PRGs are presented for information only. In addition, the geometric mean concentrations are presented for both sitewide and IHSS 119.1 only.

In addition to establishing PRGs that comply with ARARs and protect human health and the environment, DOE plans to reduce exposures and the risk associated with residual contamination during remedial actions at OU 1 to levels that are as low as reasonably achievable (ALARA) considering appropriate technical, economic, and social constraints. In applying the ALARA process at OU 1, PRGs are combined with technical and economic considerations to identify the levels of risk reduction that might reasonably be achieved. These criteria are only applicable to surface soils which may be disturbed during implementation of remediation activities at OU 1.

The ALARA process includes both planning and field components. The discussions presented in this section are consistent with the planning component of ALARA, in which PRGs are estimated for residual contamination based on hypothetical exposures. This initial analysis will be used to support implementation of ALARA in the field where, based on specific field conditions, additional contamination might be reduced to below levels determined in the planning phase.

As a general standard for radiological exposures, DOE also requires compliance with all Federal requirements for limiting doses from specific exposure modes. DOE Order 5400.5 establishes standards for nonspecific radiological exposures. These standards require that the effective dose equivalent (EDE) to a member of the public not exceed 100 mrem/year above background from all non-occupational exposure routes and that these exposures be reduced to ALARA levels.

Table 2 3
Groundwater
Risk-Based Preliminary Remediation Goals
($\mu\text{g}/\ell$)

Chemical	Preliminary Remediation Goal by Scenario ^a	
	Future On-Site Resident	Commercial/ Industrial Worker
Volatile Organic Compounds		
Carbon Tetrachloride	0 658	13 8
1 1 Dichloroethene	0 150	1 99
Tetrachloroethene	1 85	683
1 1 1 Trichloroethane	3120	293 684
Metals		
Selenium	183	N/A

^a The ecological reserve researcher scenario does not apply to this medium

Table 2-4
Surface Soils
Risk-Based Preliminary Remediation Goals
(mg/kg)

Chemical	Preliminary Remediation Goal by Scenario ^a	
	Future On-Site Resident	Commercial/ Industrial Worker
Polynuclear Aromatic Hydrocarbons		
Acenaphthene	326	2 658
Benzo(a)anthracene	0 168	1 378
Benzo(a)pyrene	0 156	0 137
Benzo(b)fluoranthene	0 307	0 070
Benzo(k)fluoranthene	1 98	1 33
Dibenzo(a h)anthracene	0 017	0 134
Fluoranthene	1 010	1 771
Fluorene	251	1 745
Pyrene	634	1 342
Polychlorinated Biphenyls		
AROCLOR 1254	0 050	0 125
Radionuclides^b		
Americium 241	1 80	4 12
Uranium 233 234	4 34	6 81
Uranium 238	2 29	3 55
Plutonium 239 240	1 25	3 68

^a The ecological reserve researcher scenario results in the same PRGs as the commercial/industrial worker scenario

^b Radionuclides are reported in pCi/g

Table 2 5
Comparison of Risk-Based PRGs, ARARs, TBCs, and Existing Concentrations
(µg/l)

Chemical ¹	Existing Concentration (grand mean) ²	IHSS 119 1 Concentration (grand mean) ²	Risk Based PRG ³	State MCL ⁴	State Groundwater Standard	PQL ⁵	
						RFETS	CDPHE
Volatile Organic Compounds							
Carbon Tetrachloride	81 20	360 6	0 658	5	0 3	5	1 0
Chloroform	4.68	16	2.33	<100	6/0.19 ^a	5	1 0
1 1-Dichloroethane	2.10	4.94	2.670	—	—	5	—
1,2-Dichloroethane	6.18	3.7	0.780	5	0.4	5	1 0
1 1 Dichloroethene	283 23	1 270	0 150	7	7	5	1 0
1,2-Dichloroethene	NA	NA	711	—	—	5	1 0
cis-1,2-Dichloroethene	0.52	2.62	225	70	70 ^b	—	1 0
Tetrachloroethene	103 48	459 5	1 85	5	5/0 8	10	1 0
Toluene	4.68	16.48	1,570	1,000	1,000	5	1 0
Total Xylenes	3.23	6.09	1,210	10,000	10,000 ^c	5	1
1 1 1 Trichloroethane	363 29	1 630 1	3 120	200	200	5	1 0
1,1,2-Trichloroethane	2.58	7.67	123	5	3/0.8 ^a	5	1 0
Trichloroethene	371.63	1,667	7.43	5	5	5	1.0
Metals							
Selenium	283 4	503 2	183	50	10 ^d /20 ^a	5	—
Vanadium	4.68	43.3	236	—	100 ^b	50	—
Semi-volatile Organic Compounds							
Naphthalene	NA	NA	885	—	—	10	—

¹ Shaded contaminants were not designated as COCs by the BRA portion of the Phase III RFI/RI

² From Phase III sampling results in Section 4 Final Phase III RFI/RI Report, June 1994

³ Based on the Future On-Site residential scenario

⁴ From CRS 25-1 107 25-1 108 25-1 109 and 25-1 114

⁵ Practical Quantitation Limits (PQLs) are reported for both the RFETS and the CDPHE.

CDPHE/Water Quality Control Commission, Basic Standards for Groundwater 3 11 0 effective 3/30/94

^b Listed as drinking water MCL in State groundwater standards Table A

CDPHE/Water Quality Control Commission Classification and Water Quality Standards for Groundwater 3 12 0 effective 1/31/94

^d Measured as a dissolved concentration
Agricultural standard.

Both ALARA and DOE Order 5400.5 requirements are appropriate for handling surface soils in OU 1 and will be addressed further under OU 2 remedial actions

2.4 General Response Actions

GRAs are general waste management strategies that are designed to satisfy remedial action objectives. Examples of GRAs include treatment, containment, excavation, and extraction. GRAs are medium specific and therefore a list of GRAs will be developed for each medium of concern. For OU 1, GRAs were only identified for the groundwater medium. Because subsurface soils are a potential continual source of groundwater contamination, they are included in discussions involving groundwater GRAs and remedial action alternatives.

2.4.1 Surface Soil General Response Actions

As indicated in the Phase III RFI/RI report and summarized in Section 1.0 of this report, surface soil PAHs, PCBs, and radionuclides will result in an acceptable risk of 10^{-4} to 10^{-6} when the hotspots are removed. Because the hotspots will be removed before implementation of any remedial actions, and because surface soil radionuclides will be addressed specifically under OU 2, the surface soil medium has been eliminated from further consideration in this report.

2.4.2 Groundwater General Response Actions

The GRAs identified for the OU 1 groundwater medium are: no action, institutional controls, containment, removal, in situ treatment of chlorinated solvents, ex situ treatment of chlorinated solvents, in situ treatment of inorganics, and ex situ treatment of inorganics. These GRAs target the contaminant groups discussed in the RAO for groundwater. A brief description of each GRA is provided below:

- *No Action* Required by CERCLA as a benchmark for comparison against other remedial action alternatives. This implies that no direct action will be taken to alter the existing situation other than short- and long-term monitoring of site conditions.

- ***Institutional Controls*** - Refers to controls based on legal and/or management policies which minimize the public's exposure to potential contaminants. Examples include controlling site access, restricting land use, and restricting access to groundwater.
- ***Containment*** For groundwater, containment would consist of actions which minimize the flux of vapor phase VOCs to the surface and/or minimize the migration of groundwater contaminants across site boundaries.
- ***Removal*** For OU 1, removal implies extraction of contaminated groundwater for treatment in the existing Building 891 water treatment system or other facilities. This also includes the excavation of soils to locate potential subsurface soil DNAPL zones and to extract contaminated groundwater.
- ***In Situ Treatment of Chlorinated Solvents*** In general, in situ treatment technologies seek to treat contaminants in place without extraction or removal of large volumes of groundwater or soil. Treatment would seek to remove, destroy, and/or immobilize contaminants through biological, chemical, or physical means. This category includes extraction technologies such as soil vapor extraction and in situ heating and includes aboveground treatment of off gas and address as both groundwater and subsurface soil contamination.
- ***Ex Situ Treatment of Chlorinated Solvents*** This GRA is similar to in situ treatment except that contaminants would be extracted/removed before treatment. Treated groundwater would be discharged through existing channels (i.e., the existing Building 891 water treatment system).
- ***In Situ Treatment of Inorganics*** This GRA is similar to the in situ treatment of chlorinated solvents. In this case, treatment would seek to immobilize contaminants through chemical or physical means.
- ***Ex Situ Treatment of Inorganics*** This GRA is similar to in situ treatment of inorganics. In this case, treatment would seek to extract and/or immobilize contaminants through chemical or physical means. Treated groundwater would be discharged through existing channels (i.e., the existing Building 891 water treatment system).

2.4.3 Volume and Area Estimates

Based on the results of the OU 1 Phase III RFI/RI report and the BRA, in particular, contaminated groundwater in OU 1 was found to contribute a significantly higher risk to those receptors exposed to groundwater beneath a specific portion of IHSS 119.1 than to receptors exposed to groundwater beneath other locations in OU 1. IHSS 119.1 was designated a source

location in the PHE for this reason. Other areas of the operable unit contain groundwater contaminant concentrations above detection limits; however, the concentrations are greatest at this IHSS.

The quantity of groundwater requiring remedial action in the IHSS 119.1 source area cannot be calculated directly because of seasonal variations in the water table. Instead, a lower bound was estimated using computer codes that compared the bedrock topography beneath the IHSS to the water level data from wells located in this area. The wells used to identify and delineate this area were 0487, 0974, 1074, 4387, 32591, and 37991.

The Phase III RFI/RI report contains several saturated thickness maps for OU 1 during a typical dry period. These maps were used to estimate the volume of contaminated groundwater in the source location when groundwater levels were at their lowest. Using an average porosity of 0.10 (DOE 1994a), the volume of groundwater estimated to be present in the southwest corner of IHSS 119.1 during the dry season is 80,000 gallons. This volume represents a single pore volume, although more than one pore volume would likely have to be removed to achieve RAOs.

In addition, the Phase III RFI/RI report estimated that the volume of available groundwater in OU 1 is between 5.0 and 5.8 acre-feet (1.6 and 1.9 million gallons). The volume of groundwater estimated to be beneath IHSS 119.1 and the volume of groundwater beneath OU 1 are used to estimate remediation requirements; however, because groundwater elevations in OU 1 are highly dependent on seasonal variations in precipitation, these values are engineering estimates only.

To address the potential residual DNAPL sources assumed to be present in IHSS 119.1 subsurface soils, a volume calculation was conducted for this medium at IHSS 119.1. The amount of soil requiring remediation was estimated by visually inspecting the potential source areas described in the Phase III RFI/RI report and presented in Figure 1.3 of this report, and by assuming that subsurface soil remediation activities would attempt to remediate saturated zone

soils to a depth of five feet into bedrock. Figure 2.1 depicts the potential soil excavation area identified for IHSS 119.1. The amount of contaminated subsurface soils cannot be calculated due to the limited data available for this medium. This limited data is typical of sites contaminated with residual DNAPLs. The excavation area, however, is estimated to contain approximately 20,000 cubic yards of soil.

2.5 Identification and Screening of Technologies and Process Options

Appendix A contains the tables and figures originally presented in TM #11 for the identification and screening of technologies and process options. This section summarizes the technologies and process options that were identified for remediation of OU 1 and also describes options that were maintained for further evaluation. Note that a process option chosen to develop an alternative is considered a representative process option only. The alternative might not be implemented with the specific process option. The selected process option represents a class of options that might be implemented.

The technologies and process options originally examined for use at OU 1 include the following:

No Action

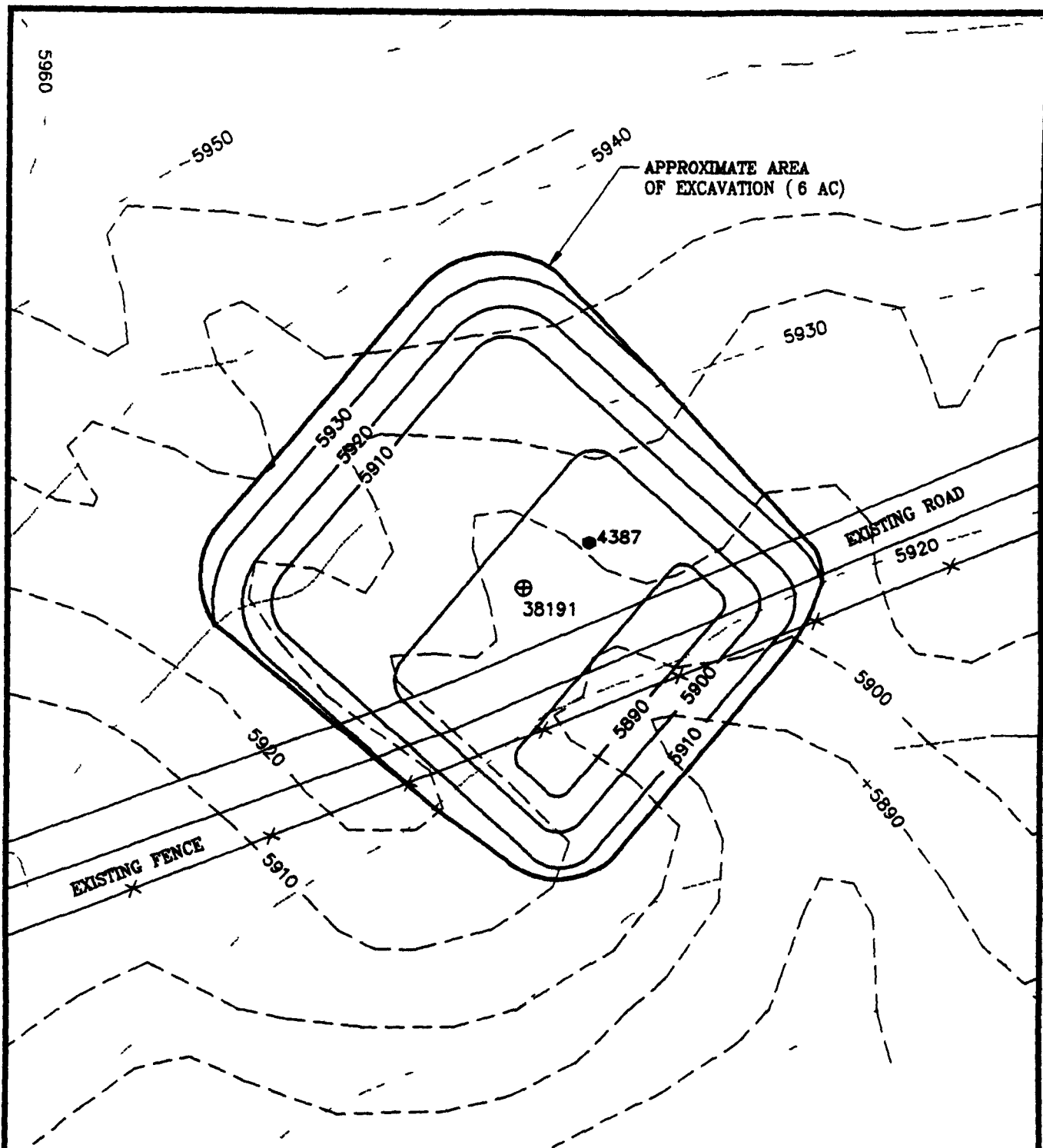
- **Monitoring**
 - **Groundwater monitoring**

Institutional Controls

- **Access restrictions**
 - **Legal restrictions on well placement**
 - **Legal restrictions on land use**

Containment

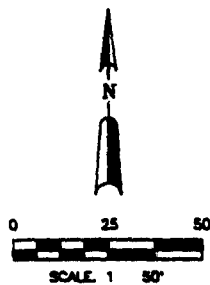
- **Vertical subsurface flow control**
 - **Subsurface drains**
 - **Grout curtains**
 - **Slurry walls**
 - **Sheet piling**



EXPLANATION

- SURFACE CONTOUR
- - - BEDROCK CONTOUR
- EXCAVATION CONTOUR

- 4387 ALLUVIAL WELL
- ⊕ 38191 PIEZOMETER



U S DEPARTMENT OF ENERGY
Rocky Flats Environmental Technology Site
Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

Potential
Soil Excavation Area

Figure 2-1

OU1-EXC2 DWG

- Cryogenic barrier
- Horizontal subsurface flow control
 - Grout injection
 - Block displacement
- Vapor containment
 - Surface cap
 - Environmental isolation enclosure

Removal

- Passive removal
 - Subsurface drains
- Active removal
 - Horizontal and/or vertical extraction wells or sumps
- Excavation
 - Loader/excavator/dozer

In—situ Treatment of Chlorinated Solvents

- Biological
 - Bioremediation
- Chemical
 - Polymerization
 - Chemical oxidation
- Physical
 - Hot air/steam stripping with mechanical mixing
 - Air sparging
 - Soil vapor extraction
 - Permeable treatment beds
 - In situ adsorption with wells (proprietary process)
 - Radio frequency/ohmic heating

Ex situ Treatment of Chlorinated Solvents

- Biological
 - Bioremediation
- Chemical
 - Solvent extraction

- Ultraviolet photolysis with chemical oxidation
- Physical
 - Gamma irradiation
 - Activated carbon or carbonaceous adsorbents
 - Air stripping
 - Membrane processes
 - Hot air/steam stripping
 - Evaporation
 - Freeze crystallization
- Thermal
 - Incineration
 - Plasma arc discharge
 - Catalytic oxidation

In situ Treatment of Inorganics

- Physical
 - Electrokinesis

Ex situ Treatment of Inorganics

- Physical
 - TRU clear (proprietary process)
 - Oxidation/reduction
 - Ferrite process
 - Magnetic separation
- Chemical
 - Freeze crystallization
 - Ion exchange
 - Evaporation
 - Membrane processes
 - Electrocoagulation
 - Precipitation

As described in TM #11 these technologies and process options were systematically screened to reduce the number to a smaller and more representative number appropriate for the preparation of remedial alternatives. The screening was accomplished by examining the technical implementability of each technology and/or process option at OU 1 (see Appendix A). Technologies and/or process options that were maintained for further evaluation are listed below.

No Action

- **Monitoring**
 - **Groundwater monitoring**

Institutional Controls

- **Access restrictions**
 - **Legal restrictions on well placement**
 - **Legal restrictions on land use**

Containment

- **Vertical subsurface flow control**
 - **Subsurface drains**
- **Vapor containment**
 - **Surface cap**
 - **Environmental isolation enclosure**

Removal

- **Passive removal**
 - **Subsurface drains**
- **Active removal**
 - **Horizontal and/or vertical extraction wells or sumps**
- **Excavation**
 - **Loader/excavator/dozer**

In situ Treatment of Chlorinated Solvents

- **Biological**
 - **Bioremediation**
- **Physical**
 - **Hot air/steam stripping with mechanical mixing**
 - **Air sparging**
 - **Soil vapor extraction**
 - **Radio frequency/ohmic heating**

Ex situ Treatment of Chlorinated Solvents

- **Biological**

- Bioremediation
- Chemical
 - Ultraviolet photolysis with chemical oxidation
- Physical
 - Activated carbon or carbonaceous adsorbents
 - Air stripping
 - Hot air/steam stripping
- Thermal
 - Plasma arc discharge
 - Catalytic oxidation

In situ Treatment of Inorganics

- Physical
 - Electrokinesis

Ex situ Treatment of Inorganics

- Physical
 - TRU clear (proprietary process)
 - Oxidation/reduction
 - Ferrite process
- Chemical
 - Ion exchange
 - Membrane processes
 - Electrocoagulation
 - Precipitation

2.6 Evaluation and Selection of Representative Process Options

Technologies and process options determined to be implementable and applicable for remediation of OU 1 were subjected to a more detailed evaluation to determine which process options should be used to develop alternatives. The evaluation was performed by comparing the ability of each process option to satisfy the given criteria under the same technology type and GRA. The criteria used to evaluate process options were effectiveness, implementability and cost (see Appendix A). Any process option that survived the initial screening could be incorporated into

an established remedial action alternative in the future

Based on the evaluation of process options the following technologies and process options were selected for alternative development

No Action

- **Monitoring**
 - **Groundwater monitoring**

Institutional Controls

- **Access restrictions**
 - **Legal restrictions on well placement**
 - **Legal restrictions on land use**

Containment

- **Vertical subsurface flow control**
 - **Subsurface drains**
- **Vapor Containment**
 - **Surface cap**
 - **Environmental isolation enclosure**

Removal

- **Passive removal**
 - **Subsurface drains**
- **Active removal**
 - **Horizontal and/or vertical extraction wells or sumps**
- **Excavation**
 - **Loader/excavator/dozer**

In situ Treatment of Chlorinated Solvents

- **Physical**
 - **Hot air/steam stripping with mechanical mixing**
 - **Soil vapor extraction**
 - **Radio Frequency/ohmic heating**

Ex situ Treatment of Chlorinated Solvents

- **Chemical**
 - **Ultraviolet photolysis with chemical oxidation**

Ex situ Treatment of Inorganics

- **Chemical**
 - **Ion exchange**

The evaluation of process options to treat extracted groundwater favored the selection of the existing Building 891 water treatment system. Since the system has been proven to effectively treat the COCs present in OU 1 groundwater at their current concentrations and the capital costs have already been incurred for designing and constructing this system, this process option is the most favorable for aboveground treatment of groundwater. If necessary, the system may also be used for other operable units with minor modifications.

In addition, the limited availability of groundwater and the complex nature of the bedrock system beneath OU 1 favored treatment by process options that would extract residual sources (e.g., DNAPL zones) to the greatest extent possible while minimizing the potential for forcing contaminants further into the bedrock system. Therefore, process options that required the injection of additional fluids into the subsurface (e.g., bioremediation and soil flushing), were not favorable. Standard and thermally-enhanced vapor extraction process options were selected for alternative development and will be used in conjunction with limited groundwater pumping to remove contaminated groundwater and potential residual DNAPLs from OU 1 subsurface soils.

Other options originally retained for alternative development included excavation and capping and were retained to provide conceptual variety to the alternatives presented for remediation of OU 1. Excavation could be used to remove subsurface soils to locate pools of contaminated groundwater and to ensure that any residual DNAPL zones are removed. Capping, on the other hand, would attempt to limit the mobility of vapor phase contaminants, thereby minimizing the risk from one of the primary risk pathways, inhalation of groundwater volatiles. These options

are further described in the discussion of alternatives in Section 3 0

Process options were also retained that would result in the assembly of limited or minimal action alternatives and include groundwater monitoring use of the existing french drain system and institutional controls These options are also discussed further in Section 3 0

Although it is currently undergoing treatability studies at RFETS bioremediation was not included in the development of remedial action alternatives for the following reasons

- The effectiveness of bioremediation at OU 1 is limited by the nature of the contaminants identified Although laboratory studies have shown up to 90 percent reduction of TCA and TCE concentrations under ideal conditions researchers are skeptical as to the full scale applicability of bioremediation under field conditions stating that implementation of biodegradation of chlorinated hydrocarbons in field situations may be limited by the toxicity of high concentrations of these compounds to microorganisms and by the slow rate of degradation possible (Baker et al 1994)
- PCE, a major OU 1 contaminant is a highly refractory compound (resistant to decay) for which there is no established field method for degradation at rates which make treatment practical
- Bioremediation is not effective in treating inorganics such as selenium An aboveground treatment system could be used to remove selenium from extracted groundwater however this would most likely limit the effectiveness of reinjection systems that recycle nutrients or non indigenous bacteria
- Site conditions at OU 1 particularly fluid circulation, limit the technical implementability of bioremediation at OU 1 The Phase III RFI/RI demonstrates the lack of a consistent, defined water source beneath IHSS 119 1 Well and borehole data in the area have indicated varying water table levels and depths of saturated zones Implementation of bioremediation at OU 1 would require injection of large volumes of water to provide nutrients and/or non indigenous bacteria to treatment zones This might mobilize and spread contamination and accelerate slumping at OU 1 Experience with installation of the french drain system has indicated that slumping is a serious concern for unsaturated conditions, and would be more serious for the highly saturated conditions that would be required to implement bioremediation

2.7 Existing IM/IRA Treatment System

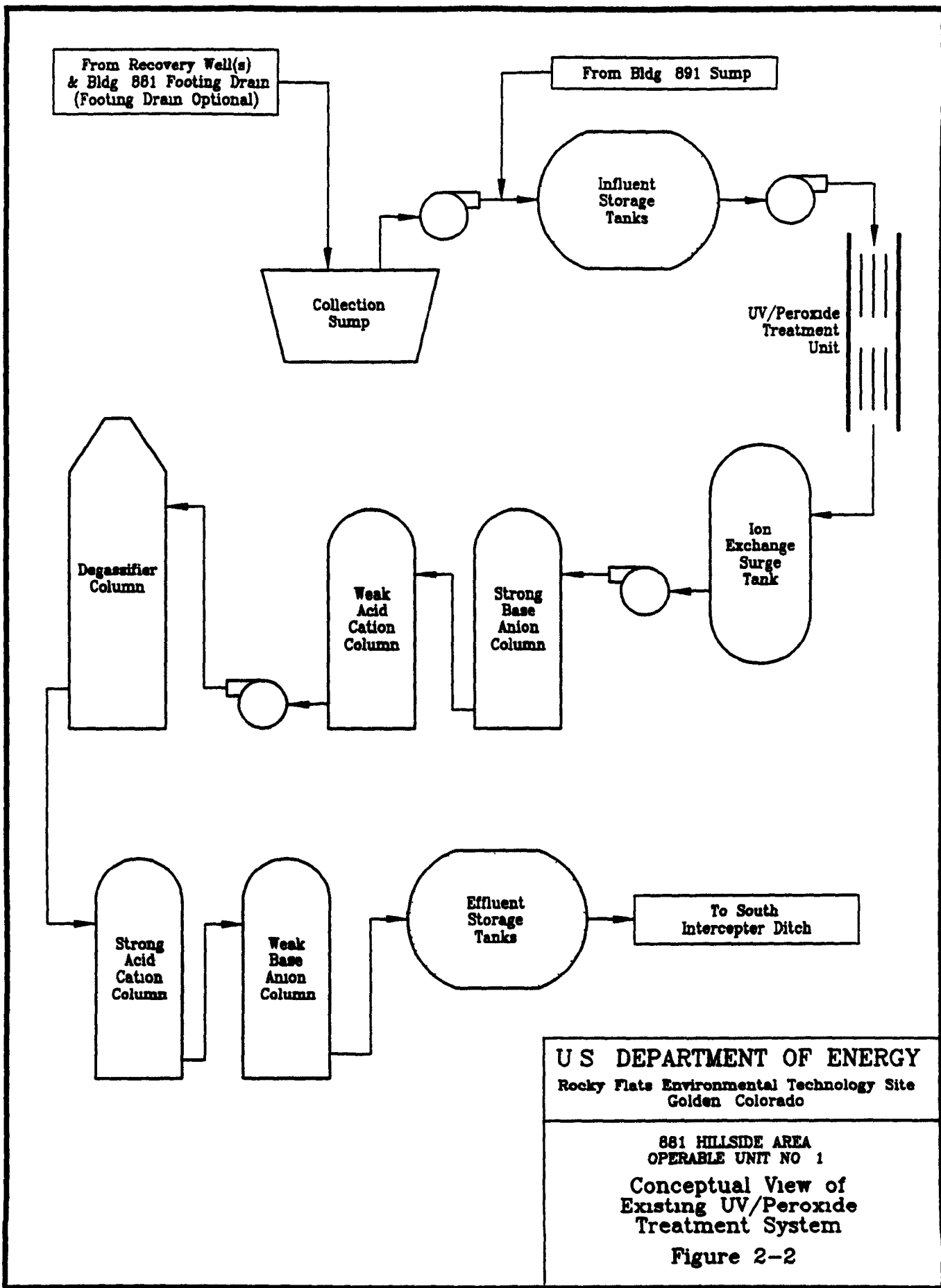
The existing Building 891 water treatment system (UV/H₂O₂ and ion exchange) will be essential for proposed remedial action alternatives for OU 1 and other operable units that require aboveground groundwater treatment. The system constitutes a comprehensive process treatment train for treating water contaminated with organic and inorganic (including radionuclide) contaminants (see Figure 2.2). The system consists of a collection and pumping system to supply the treatment facility, an influent storage and transfer system, separate treatment systems for organic and inorganics contaminants, and an effluent storage and discharge system. The system is designed for a 30 gpm flow rate capacity and has equalization tanks to normalize treatment rates.

The french drain collection and pumping system includes the recovery well pump located in IHSS 119.1, two french drain sump pumps, and the Building 891 sump pumps which may be discontinued under a proposed modification. These pumps are normally controlled by level switches in the well or sump that determine whether the pumps operate. The collection system connects to the influent transfer system, which includes two influent equalization tanks and two influent transfer pumps. The influent transfer pumps supply water from the influent equalization tanks to a UV/H₂O₂ treatment unit at a constant rate. The UV/H₂O₂ unit is designed to destroy organic contaminants in the influent stream.

Treatment efficiency depends on flow rate (residence time), H₂O₂ concentration, and UV wavelength intensity. The system has a design throughput of 30 gpm or 14,400 gallons per day (gpd) with an 8-hour operating shift. It uses 50 mg/l of H₂O₂, with sixteen 15-kW UV lamps providing an equivalent power of 240 kW for breaking down organics.

When the water leaves the UV/H₂O₂ system, it enters the ion exchange system, which consists of the ion exchange surge tank, four columns containing beds of ion exchange resins, and a degassing tower. The ion exchange system processes the water in the following sequence:

OU1-PTS DWG



- 1 The water enters the ion exchange surge tank and is pumped at a constant rate into the first ion exchange column. This column contains 28 cubic feet of Ionac A 440 a strong base anion resin for removing uranium.
- 2 The water then flows directly to the second column which contains 32 cubic feet of Ionac CC a weak acid cation resin for removing heavy metals.
- 3 The water then enters the degassing tower to allow carbon dioxide and other gases produced during the UV/H₂O₂ process to escape. Excessive gas content in the ion exchange columns could cause short circuiting of the resins thereby reducing the efficiency of the system.
- 4 The water is then pumped to the third ion exchange column which contains 56 cubic feet of Ionac C 240H a strong acid resin for removing hardness and metals.
- 5 The water then enters the fourth and final column which contains 56 cubic feet of Ionac AFP 329 a weak base anion resin, for removing anions.
- 6 The water which is now treated, is stored in one of three effluent storage tanks and discharged by gravity feed.

In terms of proposed remedial action alternatives the system can handle most contaminants identified in OU 1 groundwater at their current concentrations and the proposed treatment rates. If unusually high concentrations of specific contaminants are encountered the system may require modification to maintain effluent requirements. If other operable units require the use of this system the system may require modifications to remove contaminants if their concentrations differ significantly from OU 1.

3 0 DEVELOPMENT AND SCREENING OF REMEDIAL ACTION ALTERNATIVES

This section presents the alternatives that were assembled for remediating the groundwater medium at OU 1. These alternatives were assembled using the technologies identified both in Section 2 0 and evaluated in detail in TM #11. Appendix A summarizes the evaluation of technologies and process options. TM #11 also contains the screening of remedial action alternatives the results of which led to the alternatives presented in this section. Note that the alternatives presented herein differ slightly from those discussed in TM #11. Alterations were made where necessary in response to agency comments. However, the conceptual approaches originally proposed are still maintained in this document. In general, most of the material presented in TM #11 is presented herein. The only alternative screened from further consideration in TM #11 involved capping of the site. Capping would require institutional controls to maintain the integrity of the cap, but would not provide any additional protection beyond the controls themselves.

An integral component of most of the alternatives presented in this section involves utilizing the existing Building 891 water treatment system. This system may also be used for treating contaminated water from other areas of the RFETS. If required, the system could be modified to treat higher concentrations of specific contaminants; however, currently the system is capable of treating the COCs identified at OU-1 at their current concentrations. Decommissioning of the french drain may be appropriate once remedial actions are completed. This subject is discussed further under each alternative.

3 1 Development of Remedial Action Alternatives

Remedial action alternatives were developed by combining process options which were selected as being representative options based on the results of the evaluation of process options and technologies. Process options were combined in such a way as to permit alternatives to be developed that would range from treatment alternatives that eliminate or minimize the need for long term management to limited or no action alternatives. This range of alternatives includes

containment options that involve little or no treatment but achieve RAOs by preventing exposures or by reducing the mobility of contaminants. The no action alternative was developed to provide a baseline alternative against which other alternatives could be compared. In all cases the alternatives were developed with the goal of achieving the groundwater RAO presented in Section 2.0 by combining appropriate GRAs to form site-specific remediation strategies.

The alternatives that were developed for remediation of OU 1 are the following:

- Alternative 0 No Action
- Alternative 1 Institutional Controls without the French Drain
- Alternative 2 Institutional Controls with the French Drain
- Alternative 3 Modified French Drain with Additional Extraction Wells
- Alternative 4 Groundwater Pumping and Soil Vapor Extraction
- Alternative 5 Groundwater Pumping and SVE with Thermal Enhancement
- Alternative 6 Hot Air Injection with Mechanical Mixing
- Alternative 7 Soil Excavation and Groundwater Removal with Sump Pumps
- Alternative 8 Capping with Institutional Controls

Table 3.1 depicts a summary of the development of remedial action alternatives. The table presents the GRAs and process options that were combined to form the various alternatives. After developing alternatives for remediation of OU-1, the alternatives were screened on the basis of effectiveness, implementability, and cost, as described in TM #11. Alternatives that were dropped from further consideration are also indicated in Table 3.1 by shaded areas.

3.2 Groundwater Remedial Action Alternatives

Groundwater remedial action alternatives were developed that could potentially achieve the RAOs described in Section 2.0. The primary risk pathways that determined which GRAs would be used to develop alternatives were based on the OU 1 BRA, which indicated that ingestion of groundwater and inhalation of vapors rising up through unsaturated soils were the largest concerns. The following groundwater alternatives were designed to achieve RAOs by removing and destroying the contaminants in groundwater by restricting access to wells.

positioned within the boundaries of OU 1 and/or by limiting access to the site completely. These alternatives assume that surface soil hotspots would be removed prior to commencing remedial activities and would be put into temporary storage for treatment with similar wastes from another OU or shipped off site for immediate treatment and/or disposal.

3.2.1 Alternative 0. No Action

The No Action alternative for groundwater was developed to meet the requirements of the NCP which state that a No Action alternative should be developed regardless of site-specific conditions (EPA 1990a). The alternative provides a baseline against which other alternatives can be compared during the detailed analysis of alternatives. The No Action alternative uses the results of the BRA to define what the exposure levels would be to receptors under this alternative, and does not seek to actively remediate any portions of OU 1.

This alternative includes monitoring only to determine if any changes occur in contaminant concentrations or in contaminant migration patterns. Groundwater monitoring would begin immediately and would take place for as long as institutional controls are active at the site or until it is determined that monitoring is no longer required. Wells no longer deemed necessary for monitoring would be abandoned as appropriate.

This alternative assumes that the site would eventually be abandoned, and that no remedial actions would be initiated to reduce the risk from groundwater contaminants or to remediate potential residual DNAPL zones believed to be present in the subsurface soils beneath IHSS 119.1. The alternative assumes that the treatment portion of the existing french drain system would be non-operational. Groundwater reaching the drain would begin to flow around the drain at a slow rate assuming the existing sumps were not pumped regularly. This would result in a saturated region directly upgradient of the drain, and a less saturated region downgradient of the drain. However, no adverse impacts are expected. If desired, the drain could be decommissioned by excavating portions of the impermeable layer downhill from the drain with a standard backhoe to increase its effective permeability. For the purposes of detailed analysis

it is assumed that the drain would be decommissioned as suggested under this alternative

Since no remedial actions would be conducted under this alternative there is no remediation time frame involved. Decommissioning the french drain would be accomplished using RFETS equipment and would require minimal effort. This alternative would also not involve any packaging or transportation of waste nor any permitting actions.

3.2.2 Alternative 1. Institutional Controls without the French Drain

This alternative is intended to minimize the risk from contaminated groundwater by restricting access to any wells impacted by OU 1 contaminants and by eliminating the possibility of building construction above areas known to be contaminated with VOCs. This alternative would attempt to meet RAOs by applying institutional controls to the boundary of the RFETS at Woman Creek.

The alternative assumes that the existing french drain system would not be operational as in the No Action alternative. Groundwater reaching the drain would flow around the drain at a slow rate assuming the existing sumps were not pumped regularly. This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain however no adverse impacts are expected. If desired the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability. For the purposes of detailed analysis it is assumed that the drain would be decommissioned as suggested under this alternative.

Groundwater monitoring would be required for this alternative to determine when institutional controls could be discontinued. Once acceptable groundwater contaminant concentrations were achieved through natural degradation and dispersion of contaminants, the area would be released from institutional controls. Groundwater monitoring would take place for as long as required to meet this criterion. Wells no longer deemed necessary for monitoring would be abandoned.

This alternative assumes the site would not be abandoned during the institutional control period but that no remedial actions would be taken to actively reduce the contaminant concentrations in groundwater, or to remediate potential residual DNAPL zones believed to be present in the subsurface soils beneath IHSS 119 1

As in the No Action alternative there is no remediation time frame associated with this alternative since the site would not be released until acceptable groundwater concentrations are achieved. Decommissioning the french drain would be accomplished using RFETS equipment and would require minimal effort. For the purposes of detailed analysis a 30-year institutional control period is assumed for groundwater monitoring.

This alternative would not involve any packaging or transportation of waste, nor any permitting actions other than the administrative requirements associated with maintaining the security of the site.

3 2 3 Alternative 2. Institutional Controls with the French Drain

This alternative is intended to minimize the risk from contaminated groundwater by restricting access to any wells impacted by OU 1 contaminants while continuing to treat groundwater collected by the existing french drain. This alternative is similar to Alternative 1 with the exception that the french drain would not be decommissioned. This alternative applies to the area south of building 881 and to the IHSS 119 1 source area. Dilute concentrations of contaminated groundwater to the east of the operable unit would not be actively remediated by this alternative although institutional controls would prevent unauthorized construction and uses of groundwater in all areas of OU 1. Suspect areas of subsurface soil DNAPL contamination are not addressed under this alternative other than through containment of groundwater.

Groundwater monitoring would take place for as long as required to verify that contaminant concentrations in groundwater have been permanently reduced below appropriate limits. For this alternative the existing extraction well located in IHSS 119 1 would continue to be used as

a groundwater collection source. Wells no longer deemed necessary for monitoring would be abandoned as appropriate.

Although remedial actions would be conducted under this alternative in the form of operating the french drain system, there is no remediation time frame defined since the system is currently operational and would continue operating until acceptable contaminant concentrations are achieved. Based on operations to date of the existing french drain system, however, it is reasonable to assume that the slow groundwater collection rate would require its operation for an extensive period of time. Monitoring of groundwater would also begin immediately. This alternative could involve packaging and transportation of spent ion exchange resin.

3.2.4 Alternative 3. Modified French Drain with Additional Extraction Wells

Alternative 3 differs from Alternative 2 in that additional extraction wells would be added to the existing french drain system to enhance its effectiveness. This alternative would seek to provide protection of human health and the environment by removing contaminants from all areas of OU 1 groundwater, and by entirely containing groundwater upgradient of the french drain. As in Alternative 2, suspect areas of subsurface soil DNAPL contamination are not addressed under this alternative other than through containment of groundwater.

Wells could be added to the southeastern corner of the operable unit to capture any contaminated groundwater potentially flowing around the french drain, to the IHSS 119-1 source area to assist the existing recovery well, and/or in front of the french drain in any suspected sandstone lenses which could form conduits for groundwater transport beneath the drain. These wells could also be used to monitor COC concentrations in the area. In addition, under this alternative wells could be installed in the area south of Building 881 to enhance the recovery of contaminated groundwater in that area. Figure 3-1 shows possible locations of additional extraction wells. These locations are identified solely for defining the conceptual approach suggested for this alternative. Prior to designating exact locations, a thorough review of the impact the french drain is having at the site, potentially including computer modeling, is necessary.

Modifications required to the french drain itself would be to eliminate the flow of the footing drain water from the 881 Building to limit the amount of clean water that is sent through the treatment system

Groundwater recovered from the extraction wells would be routed to the french drain sump then transferred to the influent storage tanks of the existing Building 891 water treatment system. Recovered groundwater would therefore have to be pumped at a flow rate compatible with the system's 30 gpm capacity. This system was constructed to treat groundwater from the 881 Hillside area to achieve the treatment goals presented in the *Systems Operation and Optimization Test Report* (DOE 1992).

The institutional control of groundwater monitoring would be employed to determine when contaminant concentrations fall below acceptable levels (assumed to be 30 years for costing purposes in the detailed analysis of alternatives). The existing french drain system would provide containment of contaminants during remedial actions while also assisting in the collection of groundwater. After remedial actions are completed, however, the drain could be left in place or decommissioned. If left in place, groundwater reaching the drain would begin to flow around the drain at a slow rate, assuming the existing sumps were not pumped regularly. This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain, however, no adverse impacts are expected. If desired, the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability. For the purposes of detailed analysis, it is assumed that the drain would be decommissioned as suggested under this alternative.

Four to six additional extraction wells would be installed under this alternative and would require approximately six inch diameter casing. Because of the low hydraulic conductivity and small saturated thickness of 881 Hillside colluvial materials, cyclical operation with pumping rates below one gpm per well would be required to remove groundwater.

Computer simulations of domestic water production capabilities from OU 1 were completed and

presented in the report entitled *OU 1 Domestic Water Supply Simulations* (EG&G 1992) and are included in Appendix F of this report. Results of these simulations showed that with a hydraulic conductivity of 1×10^{-4} cm/sec pumping rates exceeding 0.14 gpm would desaturate the modeled well cell in under 365 days. The model assumed a 12 hour pumping period. With a hydraulic conductivity of 1×10^{-5} cm/sec pumping rates exceeding 0.013 gpm would desaturate the modeled well cell in under 365 days. Based on the Phase III RFI/RI report, the hydraulic conductivity at IHSS 119.1 and the area south of IHSS 119.2 is estimated at 9.4×10^{-5} ft/min while the area south of Building 881 has an estimated hydraulic conductivity of 1.5×10^{-5} ft/min. These hydraulic conductivities suggest that extremely low pumping rates would be required to remove contaminated groundwater without desaturating the modeled well cells.

The overall remediation time frame based on using this alternative would be extensive considering the low groundwater pumping rates achievable at OU 1. The potential exists for an extensive extraction time required for removal of residual DNAPLs potentially present in saturated soils. Recent EPA guidance recognizes that complete remediation of DNAPL contaminants using conventional groundwater extraction techniques is not technically practicable (EPA 1992b). Again, for the purposes of detailed analysis it is assumed that this alternative would be implemented for at least 30 years.

3.2.5 Alternative 4. Groundwater Pumping and Soil Vapor Extraction (SVE)

This alternative seeks to achieve groundwater RAOs by dewatering the IHSS 119.1 source area using conventional pumping techniques, and then following this action with implementation of a localized SVE system. The combined technologies proposed under this alternative are considered emerging technologies which may be more effective than when applied individually. In general, this alternative targets only the IHSS 119.1 source area, although additional vapor extraction wells could be installed in other areas to treat suspected DNAPL sources.

SVE would assist the vaporization and subsequent recovery of contaminants present in the

saturated soils unsaturated soils and groundwater at OU 1 The technology targets contaminants that have partitioned either to the aqueous phase in the subsurface adsorbed onto subsurface soils exist as pools of DNAPL or occupy soil pore spaces as vapor As discussed above groundwater residing in shallow pools throughout IHSS 119 1 would be extracted via the existing well the existing french drain and one or two additional recovery wells Collected groundwater would be treated by the existing Building 891 water treatment system or another appropriate facility with any modifications required to treat unusually high contaminant concentrations if encountered These same areas would be subjected to SVE once desaturated to enhance the removal of any residual contaminants

In general soil vapor extraction is an in situ physical treatment technology that has been used primarily to remediate soil and groundwater contaminated with VOCs A typical SVE system consists of either a single or if necessary a network of vapor extraction wells screened at depths consistent with the contaminated soils If multiple vapor extraction wells are used they are usually joined together by a common header pipe Makeup or clean air replacing the contaminated soil gas removed through SVE enters the soil either passively via the ground surface and/or inlet wells or actively via air injection wells Also the application of surface seals may redirect makeup air to desired treatment zones

The basic principle behind SVE involves inducing vapor flow through the unsaturated zone towards an extraction well by applying a vacuum to that well Contaminants volatilized from the soil matrix and those that are already in the vapor phase are swept by the carrier gas flow (primarily air) to the extraction well(s) The carrier gas also tends to increase the volatilization of any aqueous phase or free phase DNAPL contaminants in the vicinity Many complex processes occur on the microscale, however there are three main factors that control the performance of an SVE operation (a) the vapor flow rate through the unsaturated zone (b) the flow path of carrier vapors relative to the location of the contaminants and (c) the chemical composition of the contaminants (Johnson et al 1989)

To successfully design and operate an SVE system site geology and contaminant properties must

be considered. Site geology can have a significant influence on a vapor extraction well's radius of influence. Geological factors include depth to groundwater, subsurface soil/rock type, and subsurface permeability, which must be great enough to allow carrier vapors to strip VOCs from the subsurface matrix and carry them to an extraction well. Soil vapor extraction performance is also dependent on the characteristics of the contaminants targeted for extraction. A compound is a likely candidate for SVE if it has a vapor pressure of 1.0 mm or more of mercury at 20°C and a dimensionless Henry's Law constant greater than 0.01 (Danko 1989). Table 3.2 presents these values for the COCs under consideration at OU 1 as well as other general physical and chemical data. The data shown indicate that all of the COCs under consideration are amenable to recovery by SVE. A cross-sectional view of the proposed conceptual configuration of an SVE system is presented in Figure 3.2.

For this alternative it is assumed that approximately 10 to 30 vapor extraction wells would be installed in IHSS 119.1 and in other areas if deemed appropriate. A detailed soil gas survey would have to be conducted prior to installing these wells in order to pinpoint exact well locations. Wells would be installed to a depth of approximately 20 feet and would be 4 to 6 inches in diameter. These wells would be operated cyclically to enhance recovery and would be used in combination with a granular activated carbon (GAC) unit to treat extracted vapors. This unit would most likely require a National Emission Standards for Hazardous Air Pollutants (NESHAPs) permit to operate; however, this would not present any unusual administrative constraints.

The remediation time frame associated with this alternative is assumed to be approximately five years. Groundwater monitoring would continue once the alternative was implemented to ensure that all residual phase DNAPL zones are remediated. Monitoring would be active for 30 years after remediation unless it was determined that acceptable contaminant concentrations have been reached.

Also, once the SVE system was decommissioned, operation of the French drain would be suspended. In general, groundwater reaching the drain would flow around the drain at a slow

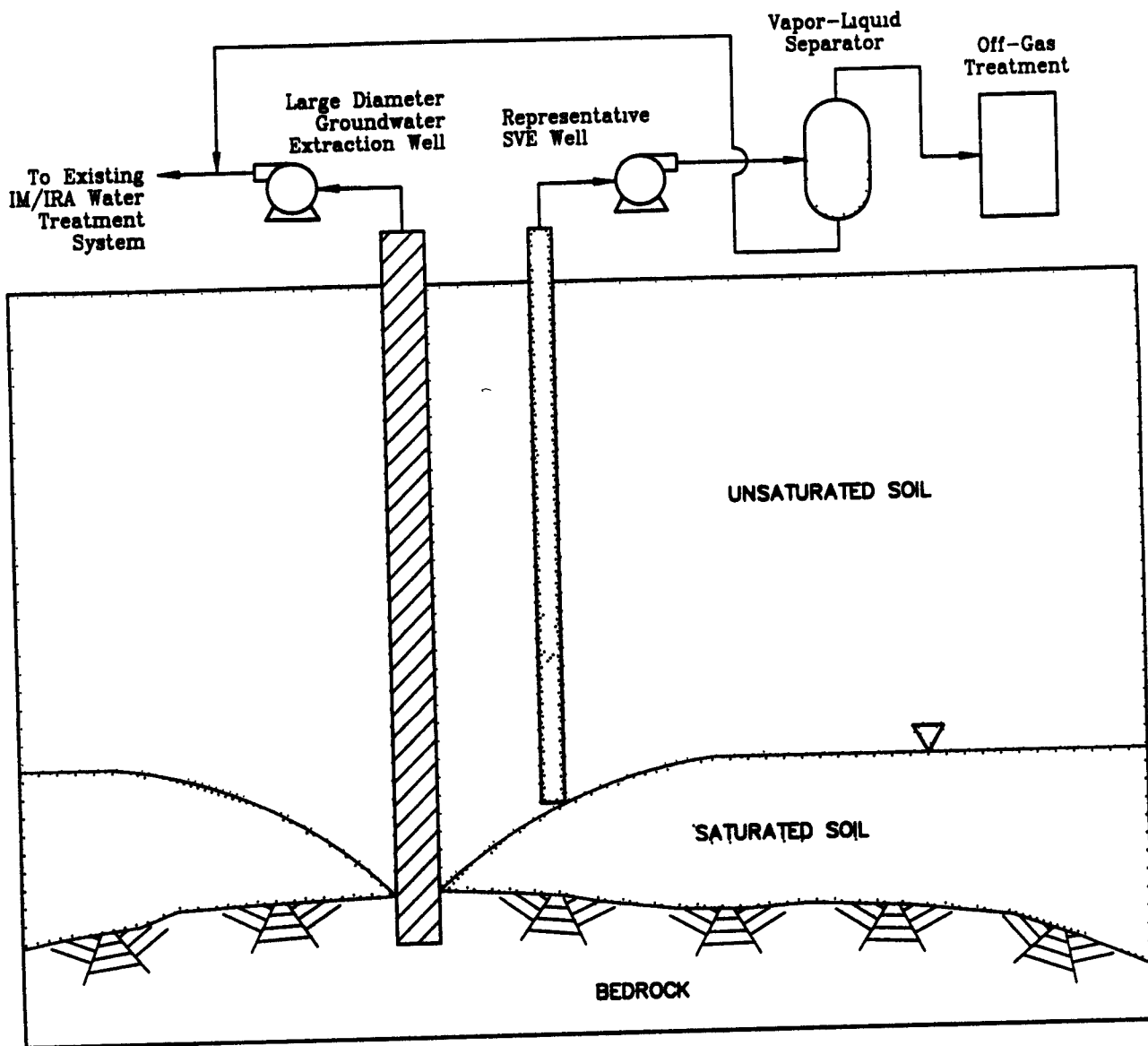
Table 3-2
Physical and Chemical Properties of the Primary VOCs in Groundwater

Chemical	Formula ^a	Molecular Weight ^a	Specific Gravity ^b	Boiling Point (°C) ^b	Aqueous Solubility (mg/l) ^a	Vapor Pressure (mm Hg)	Henry's Law Constant (Dimensionless)
Carbon Tetrachloride	CCl ₄	153.82	1.59	76.5	757	90	1.001
1,1-Dichloroethene	C ₂ H ₂ Cl ₂	96.94	1.22	37.0	5,500	182	0.179
Tetrachloroethene	C ₂ Cl ₄	165.83	1.62	121	150	17.8	1.076
1,1,1-Trichloroethane	C ₂ H ₃ Cl ₃	133.39	1.34	75.1	950	100	0.170

^a from *Hazardous Waste Management* LaGrega, Buckingham and Evans, McGraw Hill, New York, 1994

^b from *Selecting Process Equipment, vol. 1* Woods, McMaster University, Canada, 1990

^c from *Hazardous Waste Management* *ibid.* at 20 °C



Not To Scale

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**Conceptual View of
 SVE System
 Cross Section**
Figure 3-2

OU1-CS1 DWG

rate assuming the existing sumps were not pumped regularly This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain however no adverse impacts are expected If desired the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability For the purposes of detailed analysis it is assumed that the drain would be decommissioned as suggested under this alternative

3 2 6 Alternative 5. Groundwater Pumping and SVE with Thermal Enhancement

This alternative seeks to enhance the vaporization and subsequent recovery of contaminants present in the subsurface soils and groundwater at OU 1 As with the previous alternative this alternative targets contaminants that have partitioned to aqueous, and vapor phases or are residuals in the subsurface This alternative considers technologies that enhance vaporization through the elevation of subsurface temperature in areas where target contaminants are concentrated Groundwater residing in shallow pools throughout IHSS 119 1 would be extracted via existing wells the existing french drain, and one or two new recovery wells Collected groundwater would be treated by the existing Building 891 water treatment system These same areas would be subjected to thermal enhancement techniques once desaturated to enhance the removal of any residual contaminants Thermal enhancement is expected to be especially well suited for sites with tight formations such as is the case with OU 1 and is considered an emerging technology by EPA

As soil gas contaminated with contaminant vapors is recovered through a standard vapor extraction system and replaced with clean soil gas aqueous phase, DNAPL phase and adsorbed contaminants vaporize until they return to equilibrium saturation levels in the clean soil gas thus increasing both the vaporization rate of these contaminants and the equilibrium air saturation levels by temperature elevation subsequently increases recovery by vapor extraction Although increased vaporization rate and increased equilibrium saturation levels would increase the effectiveness of the vapor extraction system the primary increase in total contaminant recovery would result from an increase in the number of open pore spaces available for vapor

transport Any vaporization enhancement techniques used with vapor extraction would decrease the moisture content of the surrounding media Pore spaces that were initially filled with water would be opened once the water was vaporized and driven off The open pore spaces would allow for a greater diffusion rate of vapor phase contaminants thereby increasing their extraction rate and possibly the radius of influence of a vapor extraction system

This alternative considers two viable treatment technologies that can effect an increase in subsurface soil temperatures — radio frequency heating and electrical resistance (ohmic) heating Both technologies are discussed below although for the purposes of detailed analysis radio frequency heating is analyzed further, whereas ohmic heating is merely assumed to be potentially applicable at OU 1 and is not included in the detailed analysis of alternatives

Radio Frequency Heating

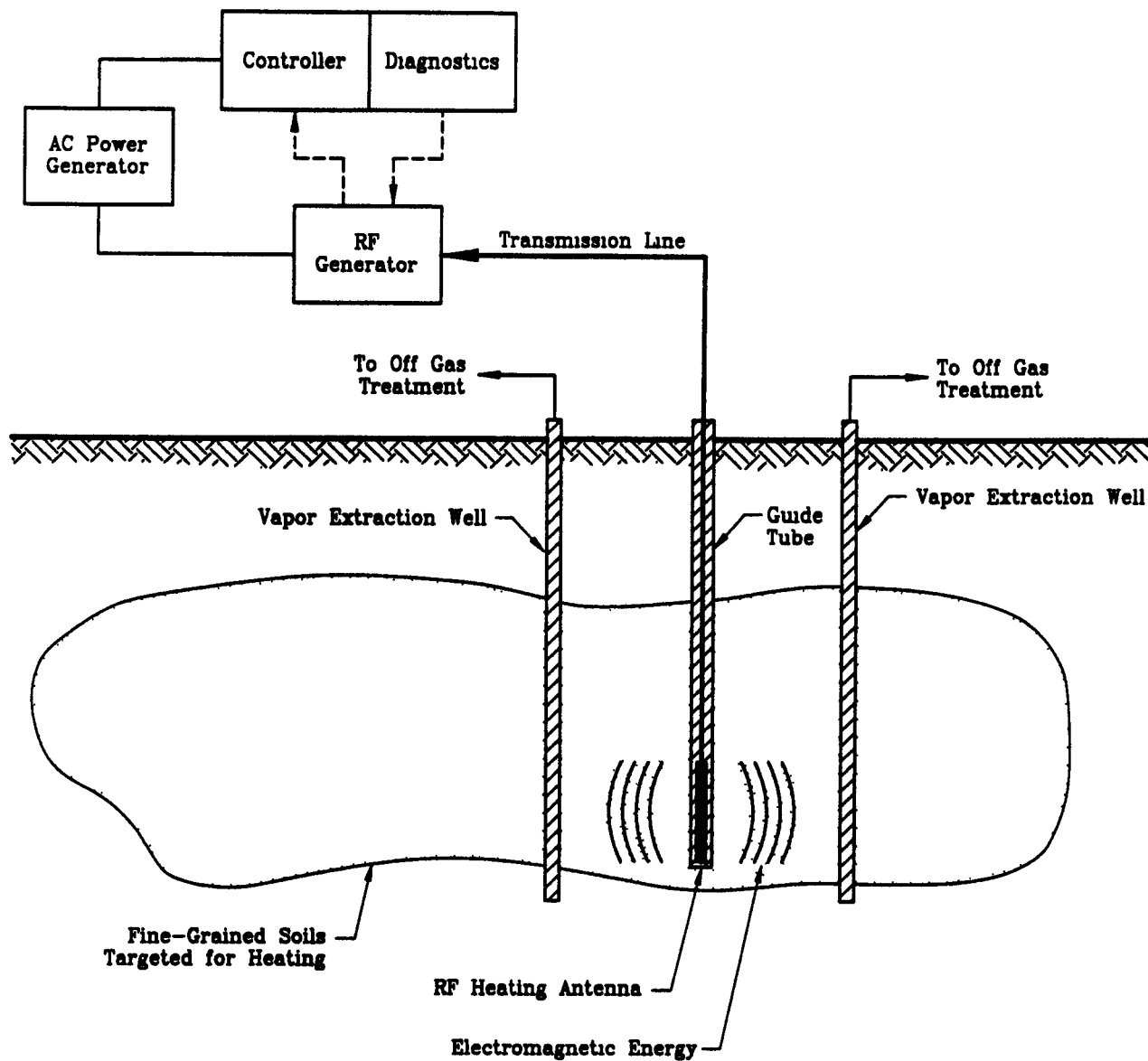
Radio frequency (RF) heating was selected as one of the two representative process options to effect an elevation in temperature of the subsurface materials at OU-1 that are contaminated with those contaminants that are VOCs RF heating is an innovative in situ technology for volatilizing organic constituents in soil and water as well as vaporizing pore space moisture The technology is desirable since additional chemicals are not introduced into the subsurface and no special arrangement (e g grids) are necessary as in conventional ohmic heating

The in situ RF heating process requires minimal intrusion using 3 to 6-inch diameter boreholes containing strategically placed antennae in the desired treatment area Through a combined mechanism of ohmic and dielectric heating the temperature in the media is raised and the volatile and semivolatile organic constituents are volatilized (Kasevich 1992) Volatilized organics are then collected with the vapor extraction system and subjected to further treatment RF heating is expected to supplement vapor extraction in a manner that allows for quicker recovery of VOCs from certain areas of the subsurface Specifically heating VOC source areas can expedite VOC recovery in the vapor form (i e , hotspots are likely to contain aqueous DNAPL and adsorbed phase VOCs which would be driven to vapor under elevated temperature

conditions) Figure 3 3 illustrates a simple application of RF heating combined with vapor extraction for this alternative

The dielectric loss of a material (i.e. the amount of energy a material dissipates as heat when placed in a varying electric field) contributes to the heating of the contaminated media. An indicator of a material's ability to successfully absorb electromagnetic energy is its dielectric constant. Most soils have suitable dielectric constants that allow for effective treatment. Water and/or soil moisture is vaporized by RF energy; however, steam is transparent to RF energy and does not continue to absorb radiation energy. While the steam may become superheated, this occurs only by energy conduction from the solid media and not from direct electromagnetic energy absorption. The steam in turn serves to heat surrounding materials, enhancing additional vaporization. Thus, water and/or soil moisture does not present a hindrance to the treatment process. Fractures and voids within the contaminated matrix also do not present treatment problems since thermal conduction is not the primary heat transfer mechanism. Densely packed soils are well suited to this treatment as are other consolidated geologic materials. A variety of heating profiles can be generated by manipulating the subsurface placement of RF antennae, their operating frequencies, and the phase output of the different antennae. Virtually uniform heating within a specified volume can be achieved with minimal heating of surrounding material using a properly designed configuration. Thus, localized treatment can be attained with proper design.

RF heating has been shown to be capable of increasing soil temperature to approximately 500°F. This temperature would be great enough to volatilize both sorbed and potentially dissolved phase contaminants (e.g., aqueous phase) in the subsurface materials as well as drive off any moisture in nearby pore spaces. The temperature of the subsurface medium would be raised gradually; therefore, vapor extraction wells would be able to extract vapor as it is generated. The heating and resulting steam/vapor generation rate could be controlled so that the capacity of the vapor recovery system would not be exceeded. Such control would prevent the spread of contamination by steam plume expansion. Also, RF heating would only be implemented in the vicinity of a vapor extraction well. Placement of an RF heating antennae in this manner would provide assurance that RF heating would not lead to a spread of contamination. A vapor



Note Figure represents information provided in part by KAI Technologies Inc

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**Conceptual View of
Radio Frequency
Heating System**

Figure 3-3

recovery system supplemented with RF heating would likely require additional air drying capacity since it is expected that the RF heating system would lead to the extraction of a greater amount of soil moisture than conventional vapor extraction

The primary piece of equipment of this alternative is the applicator antenna which is placed in a borehole. This antenna is generally a flexible component of varying length that radiates electromagnetic energy in the form of radio frequency waves. The energy originates from a generator at the surface and is transmitted to the antenna via a metal coaxial cable. Standard drilling equipment can be used to complete a borehole. The borehole is generally cased with fiberglass or a similar material that is transparent to electromagnetic radiation. The antenna can be placed in vertical or horizontal boreholes. Also, several antennae may be used concurrently in various areas with elevated contaminant concentrations.

Locations of RF antennae and vapor extraction wells for cleanup of the volatile subsurface contaminants at OU 1 are contingent on detailed design through which the optimum system design would be defined, however, it is assumed under this alternative that RF heating antennae would be installed in vapor extraction wells near the vapor extraction wells being operated. The number of vapor extraction wells required would range from 10 to 30 depending on saturation levels. The spacing between boreholes can range depending on the RF heating frequency, depth, interval of heated volume, and properties of the materials heated. An array of multiple boreholes can provide uniform heating of a given subsurface volume. Control devices monitor performance of the RF generator and adjust the outputs to optimize system performance. Soil gas monitoring wells must be in place in the vicinity of the RF heating antennae. These wells are necessary to monitor for potential increased migration of contaminant outside of the radius of influence of the vapor extraction well(s).

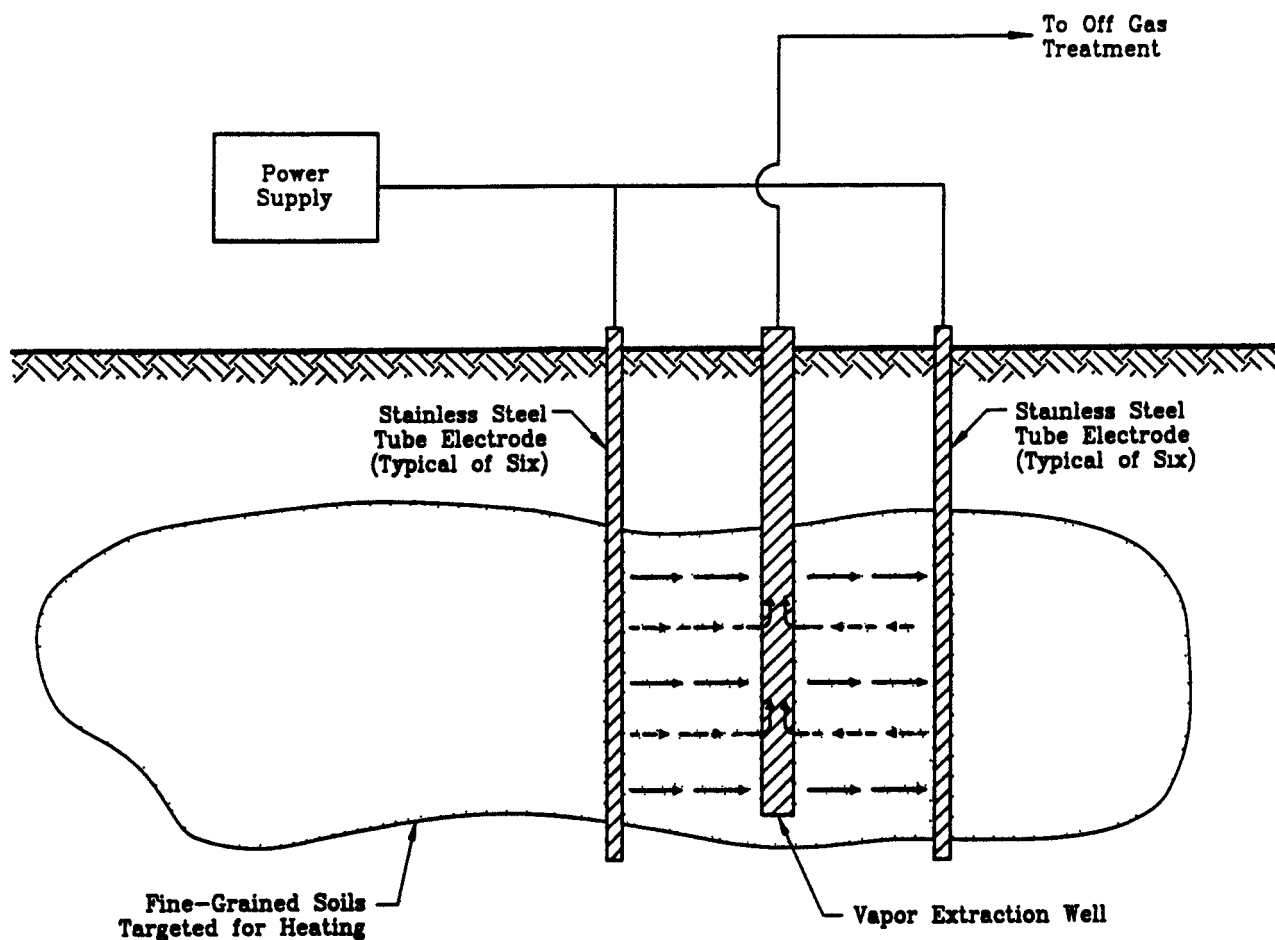
Support equipment for RF heating can be housed in one trailer. A portable power supply, such as a diesel motor generator, may provide the necessary three phase power for the RF antennae. All transmission lines connecting support equipment to the RF antennae are commercially available.

Ohmic Heating

Ohmic heating was also selected as one of the two representative process options to effect an elevation in temperature of the subsurface materials at OU 1 that are contaminated with volatile contaminants. This technology is considered an emerging technology which is currently being examined under the Operable Unit 2 (OU 2) treatability study program. Like RF heating, ohmic resistance heating is an innovative in situ technology for enhancing the performance of soil vapor extraction by volatilizing organic constituents in soils and groundwater and by vaporizing pore space moisture. Unlike RF heating, however, ohmic resistance heating results from the transmission of an electrical current through the media targeted for cleanup. As such, a prerequisite for ohmic heating is that the media must be able to conduct an electrical current. Ohmic heating requires the placement of a grid of electrodes and sometimes the addition of water in the area targeted for remediation. The process requires only minimal intrusion and has most often been implemented using six electrodes installed in a hexagonal pattern to the depth of the contaminants, with a vapor extraction well placed in the center of the pattern as shown in Figure 3-4 (Aines et al).

Six or three phase power can be used to supply current to the installed electrodes. There is some benefit with six phase power in that a more uniform heating pattern can be realized in the area being treated (Buettner et al). However, the increased uniformity comes at the expense of needing additional equipment to split normal three phase power into six phase. Electrodes are usually constructed of stainless steel tubing, which can also serve as passive air inlets.

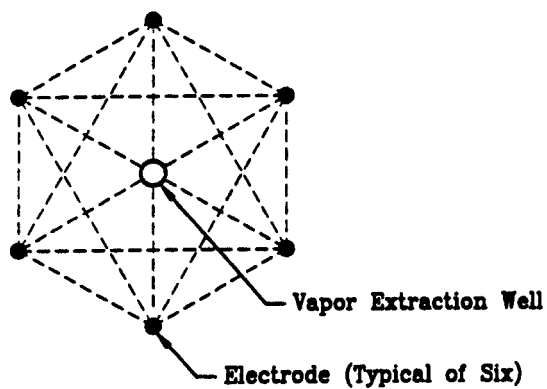
The principle of ohmic heating is simple. Basically, electrical currents are made to flow between electrodes placed in a contaminated region causing resistance heating (much the same way that passing an electrical current through an oven heating element generates resistance heating). Current flow through subsurface materials tends to be greatest in fine-grained soils such as silts and clays. These types of soils are generally less permeable than sands and gravel, thus heating the clays and silts can drive off contaminants contained therein that are not easily accessible with conventional soil vapor extraction. Once the volatile contaminants are driven



EXPLANATION

- Flow of Electrical Current Between Electrodes
- > Vapor Flow Toward Vapor Extraction Well

Plan View of Grid Arrangement



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**Conceptual View of
 Electrical Resistance
 Heating System**

Figure 3-4

out of the less permeable clays and silts into the more permeable sands and gravel they are more susceptible to recovery by vapor extraction. As with RF heating, soil moisture can be heated with ohmic heating to generate steam. Steam can provide additional stripping of adsorbed contaminants. Also, the removal of soil moisture can increase the air flow permeability of the soil being treated, thus enhancing the capability of vapor extraction to remove contaminants (but lessening the ability to continue heating the subsurface with electrical current).

The primary pieces of equipment needed to support ohmic heating include stainless steel piping (for electrodes), a 60 Hz power supply, an optional six phase transformer, thermocouples for monitoring subsurface temperature, and a vapor recovery/treatment system. Electrode grids may be placed at various locations targeted for treatment. Extracted vapors from multiple locations may be directed to a central treatment location or to individual treatment units.

The location of the electrode grid(s) and vapor extraction well(s) for cleanup of the volatile subsurface contaminants at OU 1 are contingent on treatability test results in which the optimum system design would be defined, however, for this alternative it was assumed that one grid would be installed at IHSS 119.1. This grid would have six electrodes inserted to approximately 20 feet below the surface in a hexagonal arrangement making up a circle with a diameter of approximately 20 feet. Additional grids would be required to remediate the entire site. As previously discussed, the conceptual approach presented for RF heating is carried forward for detailed analysis. The information presented here on ohmic heating may be beneficial if it is selected as the preferred technology prior to implementation of any remedial actions at OU 1.

Implementation of either technology would still require groundwater monitoring to ensure that residual DNAPL sources have in fact been remediated. In addition, operation of the french drain would be discontinued after implementation of the alternative unless the system was utilized for another operable unit. Groundwater reaching the drain would continue to flow around and beneath the drain albeit at a much slower rate than prior to its installation, assuming the existing sumps were not pumped regularly. This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain, however, no

adverse impacts are expected. If desired the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability. For the purposes of detailed analysis it is assumed that the drain would be decommissioned as suggested under this alternative.

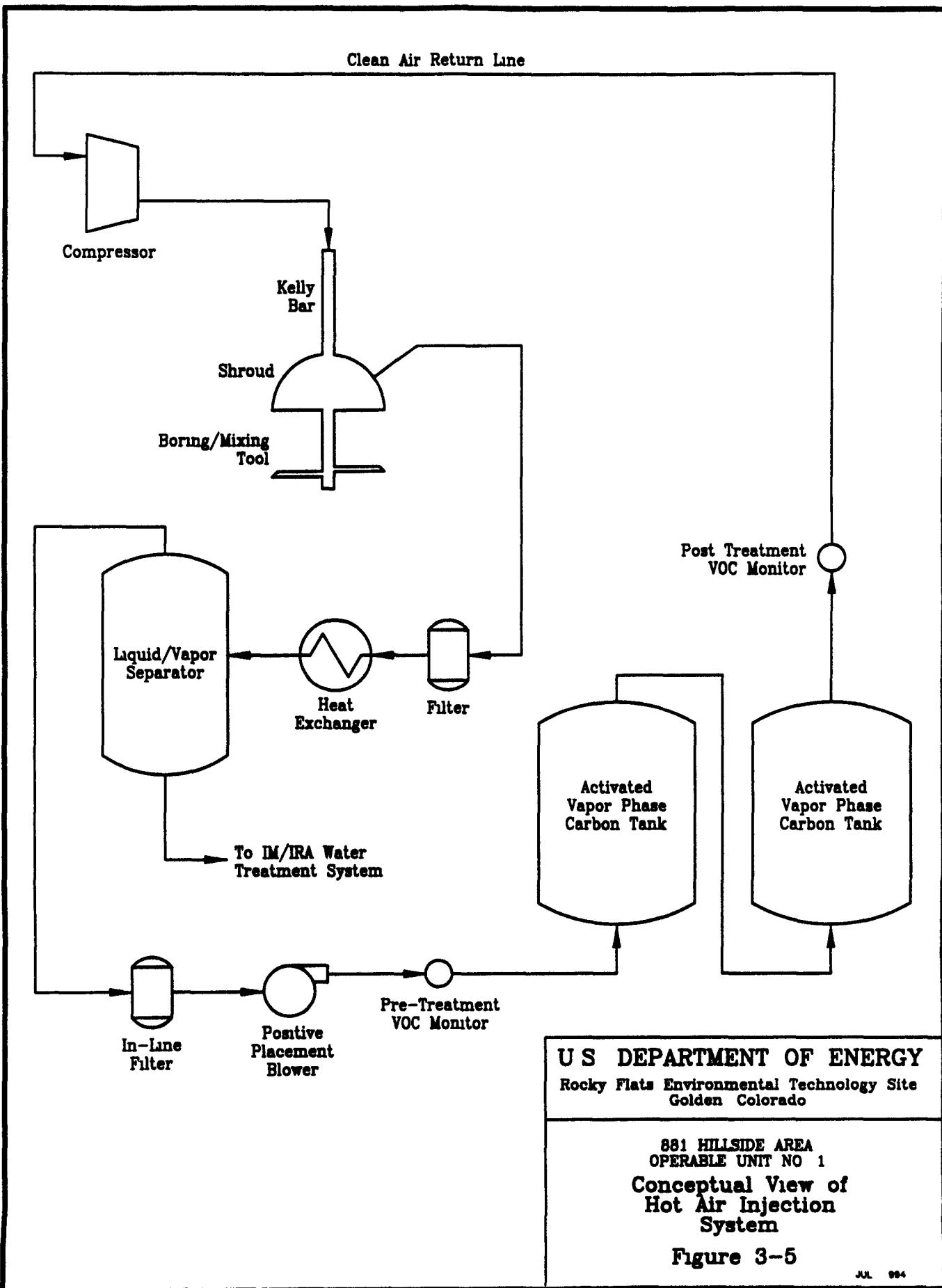
This alternative would require a remediation time frame of approximately three years. NESHAPs permits would be required for any other gas treatment systems and groundwater monitoring is assumed to be present for 30 years after remediation. This would be required to verify that all residual sources of DNAPLs in the subsurface have been remediated.

3.2.7 Alternative 6. Hot Air Injection with Mechanical Mixing

This alternative would use groundwater extraction and hot air enhanced vapor extraction with mechanical mixing to enhance recovery of contaminants present in the subsurface at IHSS 119.1. This technology is considered innovative and would have to be tested at OU 1 prior to full scale operation. Such a technology would target contaminants that have partitioned to the aqueous phase in the subsurface, those that have adsorbed onto the subsurface soils, those that exist as pools of free phase DNAPL and those that occupy soil pore spaces in the vapor phase. To maximize the efficiency of this alternative it is assumed that a detailed soil gas survey would be required to identify potential residual DNAPL sources in the subsurface.

This alternative requires the remediation of up to 20,000 cubic yards of soil in IHSS 119.1 by in situ treatment with a mobile treatment system. The treatment system selected would use hot air to enhance volatilization of adsorbed and dissolved VOCs while simultaneously increasing contact of the hot air with the VOCs by mechanical mixing. (Available groundwater would be extracted from the vicinity prior to treatment.) Heated air is both the primary means of temperature elevation, induction and of increasing subsurface vapor flow and recovery. The mixing enhances volatilization by increasing desorption surface area and eliminating barriers to contact between the contaminants and the hot air. Figure 3.5 presents a conceptual view of the hot air injection system.

OU1 - WTS DWG



The primary treatment system in this alternative consists of a caterpillar mounted drill rig with specialized drilling equipment. The drill equipment is capable of delivering treatment reagents such as hot air or steam via piping in a hollow drill bit shaft. The drill bit has a cutting/mixing blade which can vary in diameter from 4 to 12 feet. Groundwater extraction wells would be placed in previously treated soil columns. Dewatering of a small area prior to treating the initial soil column would be accomplished via an extraction well drilled with conventional drilling equipment. Extracted groundwater would be treated through the existing UV/peroxide treatment system. The drill rig can produce up to 350 000 ft lbs of torque, sufficient to provide excellent mixing of subsurface soils as the drill bit descends through the soil column. The drill bit also has multiple injection ports for hot air delivery. The multiple ports provide uniform delivery of hot air throughout the treatment zone. The caterpillar mounted drill rig is moved from one treatment zone to another sequentially until the entire site is remediated. The treatment columns or drill shafts are overlapped by 20% to ensure adequate treatment throughout the entire site. 4 to 6 columns can be treated per day, depending on site conditions.

For volatile compounds such as those at OU 1, a negative pressure shroud is placed over the entire treatment zone to capture off gases for delivery to an onboard off gas treatment system. Mats are placed under and around the rig to ensure that contaminants do not reach the atmosphere by surfacing outside the shroud. The shroud vacuum is connected to an off gas treatment system. A vapor-liquid separator removes entrained liquids for delivery to the Building 891 water treatment system. Vapors continue through the off gas treatment system. For the contaminants and concentrations at OU 1, vapor phase carbon adsorption is the preferred treatment option. Once treated, the air is recycled to a compressor and heater and reinjected to the subsurface.

Removal of groundwater by pumping will be accomplished by extraction wells placed near the treatment zone to depress the water table and recover contaminated groundwater. The wells will be placed in post treated soils due to the ease of placement in these disturbed areas. This ensures the recovery of aqueous inorganics present in the groundwater. Thus the alternative will

address inorganic as well as organic contaminants. The recovered groundwater would be pumped to the existing Building 891 water treatment system which is designed to treat all contaminants found in OU 1 groundwater.

Although this alternative involves removal of the source of contamination, monitoring of groundwater would be required once the remedial action is complete to verify that all residual DNAPL sources have been remediated.

It is assumed that after completing this alternative, the existing french drain would be decommissioned. Groundwater reaching the drain would continue to flow around and beneath the drain albeit at a much slower rate than prior to its installation, assuming the existing sumps were not pumped regularly. This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain; however, no adverse impacts are expected. If desired, the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability. For the purposes of detailed analysis, it is assumed that the drain would be decommissioned as suggested under this alternative.

This alternative would require 3 years to implement and would require permits for off-gas treatment only (assuming the existing Building 891 water treatment system is currently permitted appropriately). Groundwater monitoring would continue for 30 years or until it is determined that monitoring is no longer required.

3.2.8 Alternative 7. Soil Excavation and Groundwater Removal with Sump Pumps

This alternative is intended to reduce or eliminate the risk to a residential receptor at IHSS 119.1 through source removal of contaminated groundwater beneath a discrete portion of the IHSS. This alternative differs from the in situ groundwater treatment alternative in that a portion of unsaturated soils at the IHSS would be excavated down to the water table to allow for the removal of localized groundwater contamination. This is a worst-case scenario which would enable contaminated water to be located and subsequently removed. Such efforts may be

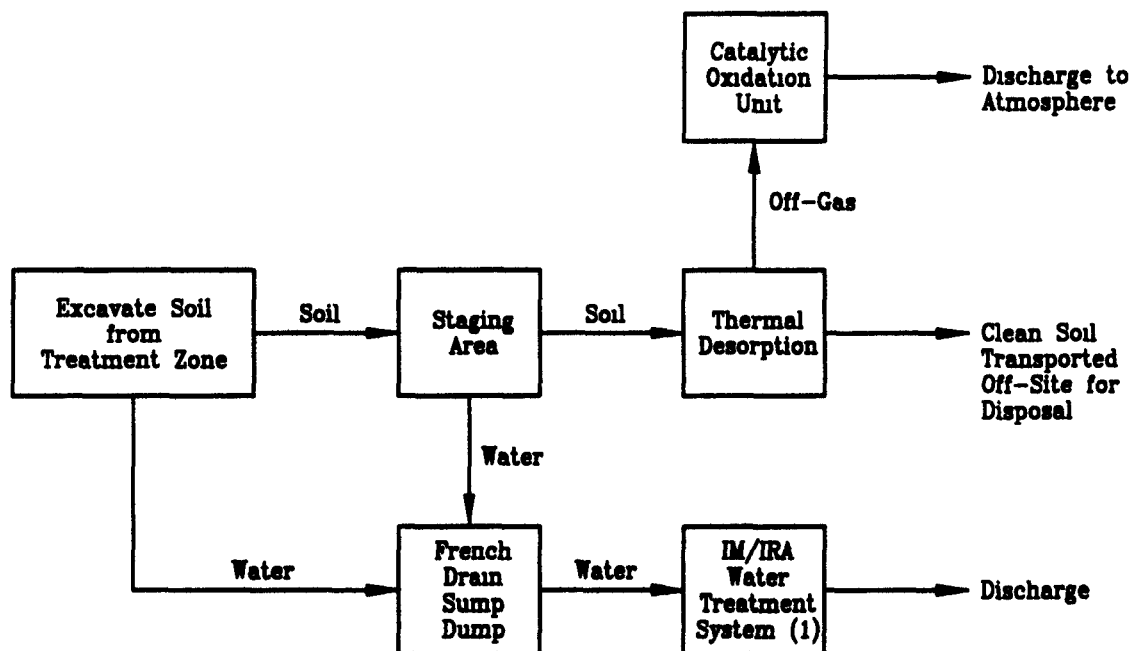
required based on the current understanding of the hydrogeologic conditions at OU 1 which suggest complex geology in the area

The volume of groundwater requiring treatment and the amount of soil which would have to be excavated for this alternative were calculated based on the results of the Phase III RFI/RI. This alternative would require excavation of approximately 20 000 cubic yards of unsaturated and potentially saturated soils in the southwest corner of IHSS 119 1 (see Figure 2 2). The amount of groundwater collected during the excavation would be approximately 80 000 gallons depending on the seasonal level of the water table. This is a rough estimate of the amount of groundwater present under low saturated conditions using the measured porosity of the soils.

Excavation would be terminated slightly below the underlying bedrock to ensure that all contaminated groundwater pools are reached. The groundwater would be collected using sumps installed within the excavation. Standard submersible pumps would be used to direct collected groundwater to the existing french drain sump pumps. The groundwater would then be transferred to the Building 891 water treatment system at OU 1 for final treatment and discharge. A piping system from the excavation to the OU-1 treatment facility would be required (see Figure 3-6). This would likely be constructed of PVC and buried to a sufficient depth to prevent freezing. A control system would also be needed to operate pumps as demand required and to minimize the need for manual oversight and operation.

The actual excavation would be accomplished using conventional construction equipment although breathing apparatus may be included as part of the machinery or may be handled separately on an individual basis. The excavated soils would be treated on-site using a skid mounted thermal desorption unit and then transported to a licensed facility for disposal.

Radiological monitoring would be conducted for the duration of the alternative due to the potential presence of plutonium in the soils. Although this alternative involves removal of the source of contamination to groundwater at IHSS 119 1, monitoring of groundwater would still be required once the remedial action is complete to verify that all sources of residual DNAPL



(1) See Figure 2-3

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Conceptual View of
Excavation and Treatment
Process

Figure 3-6

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contamination have been remediated Short term monitoring of vapor concentrations in air would also be required during the excavation and prior to its closure

The remediation time frame assumed for this alternative is less than one year Once remediation activities are completed the existing french drain would be decommissioned if appropriate If terms of the drain itself, groundwater reaching the drain would continue to flow around and beneath the drain albeit at a much slower rate than prior to its installation assuming the existing sumps were not pumped regularly This would result in a saturated region directly upgradient of the drain and a less saturated region downgradient of the drain however no adverse impacts are expected If desired the drain could be decommissioned by excavating portions at the drain with a standard backhoe to increase its effective permeability For the purposes of detailed analysis it is assumed that the drain would be decommissioned as suggested under this alternative

4 0 DETAILED ANALYSIS OF ALTERNATIVES

This section documents the detailed analysis of the following remedial action alternatives

- **Alternative 0 No Action**
- **Alternative 1 Institutional Controls Without the French Drain**
- **Alternative 2 Institutional Controls With the French Drain**
- **Alternative 3 Modified French Drain With Additional Extraction Wells**
- **Alternative 4 Groundwater Pumping and Soil Vapor Extraction**
- **Alternative 5 Groundwater Pumping and Soil Vapor Extraction With Thermal Enhancement**
- **Alternative 6 Steam Injection With Mechanical Mixing**
- **Alternative 7 Soil Excavation and Groundwater Removal With Sump Pumps**

Alternative 8, Capping With Institutional Controls is not included in the detailed analysis of alternatives because it was screened out from further analysis in Section 3

4 1 Introduction

This section analyzes the proposed remedial action alternatives using the criteria specified at 40 CFR 300.430 of the NCP. Details of the alternatives presented in Section 3.0 are used as the basis for these evaluations. The two most important criteria, the threshold criteria, are statutory requirements that must be satisfied by any alternative in order for it to be eligible for selection. The threshold criteria are overall protection of human health and the environment and compliance with ARARs.

The five primary balancing criteria of long term effectiveness and permanence, reduction in toxicity, mobility and volume, short term effectiveness, implementability, and cost are used to

evaluate major performance objectives for alternatives. The relative performance of each alternative is evaluated and compared to identify any alternatives that are clearly superior or inferior to the other alternatives under consideration.

The two modifying criteria, state acceptance and community acceptance, evaluate the feasibility of using the preferred alternative in terms of its acceptance by regulatory agencies and the community at large. These criteria are not evaluated until after the formal public comment period on the CMS/FS report, and are then addressed in the Corrective Action Decision/Record of Decision (CAD/ROD).

4.1.1 Overall Protection of Human Health and the Environment

This criterion addresses the overall protectiveness of the proposed remedy by describing how human health and environmental risks are eliminated, reduced, or controlled through treatment engineering controls or institutional controls. This evaluation criterion acts primarily as a final check on the conclusions reached in applying the other primary balancing and threshold criteria. In particular, this overall assessment of protectiveness draws on the analyses conducted under the compliance with ARARs, long term effectiveness and permanence, and short term effectiveness criteria. The evaluation of overall protectiveness examines whether an alternative results in any unacceptable short term or cross media impacts.

4.1.2 Compliance with Applicable or Relevant and Appropriate Requirements

The selection of ARARs for a remedial action alternative at a site is governed by the regulations of the NCP [40 CFR 300.400(g)] and EPA's guidance in Office of Solid Waste and Emergency Response (OSWER) Directives, such as the *Compliance with Other Laws Manual* (#9234.1). The potential list of ARARs for remedial actions at OU-1 has been presented to the regulatory agencies in TMs #10 and #11. A discussion of the selection of chemical specific ARARs for OU-1 has been presented in subsection 2.3.1. Briefly summarized, ARARs are

- applicable, that is, a requirement that under circumstances other than CERCLA apply to the contaminant action situation or location

or

- relevant and appropriate This is a requirement that is not normally applicable to the set of circumstances (contaminant activity location or situation) but because the requirement addresses an activity, location or situation similar to the circumstances proposed at the remedial action site and the requirement is well suited to the remedial action at the site it is judged relevant and appropriate It is possible for a requirement to be relevant but not appropriate for site specific circumstances

As remedial action alternatives are developed and screened through the feasibility study process so are the ARARs further analyzed and screened in the CMS/FS process

ARAR Screening Process

Action specific and location specific ARARs previously identified in the early stages of the CMS/FS process were screened again to check the jurisdictional and circumstantial prerequisites Each ARAR was noted as applicable or relevant and appropriate for each alternative at OU 1 The criteria used to evaluate applicable requirements are

- substance or contaminant addressed under statute/regulation
- time period statute/regulation is in effect
- activities/action statute/regulation requires limits or prohibits
- who is subject to statute/regulation
- exemptions under statute/regulation

The criteria used to evaluate relevant and appropriate requirements are

- similarity of substance or contaminant addressed under statute/regulation to situation at OU 1
- similarity of media affected by the requirement under statute/regulation to circumstances at OU 1

- similarity of entities affected by statute/regulation to actions/activities proposed at OU 1
- similarity of the place addressed by statute/regulation and the type of place affected by proposed action at OU 1
- similarity of structures/facility/technology addressed by statute/regulation to structure/facility/technology proposed at OU 1
- any exemptions or variances of a requirement and their availability for circumstances at the OU 1 site

Each specific remedial action alternative is assessed to determine if the proposed action will/can comply with each ARAR or TBC. Section 121(d) of CERCLA requires remedial actions to comply with or exceed the ARARs designated at a site. This is one of the primary threshold criteria designated in the NCP regulations for choosing a proposed remedial action at a site. The results of the ARAR analysis conducted at OU 1 specific to each proposed alternative is presented in Appendix D. Key ARARs selected for discussion in the detailed analysis of alternatives listed below where key ARARs are those ARARs judged to be most critical to the implementability of an alternative.

- Colorado Primary Drinking Water Standards Articles 1 14 CRS 24-4-104 105 and 25 1 107, 109, 114
- RCRA Regulations - Parts 262 264 265 and 268 and proposed changes to 261
- Colorado Solid Waste Regulations 6 CCR 1007 2 (2 1 15 2 5 5 and 2 5 7)
- Colorado Air Pollution Control Regulations 5 CCR 1001 5 Regulation 7
- Colorado Non game, Endangered or Threatened Species Conservation Act CRS 33-2-101

Compliance with an ARAR can be waived under specific circumstances as designated in CERCLA, as amended [Section 121(d)(4)] and in the NCP regulations. Any proposed waivers from compliance with ARARs is presented in the proposed decision document along with the reasons for such contemplated action.

Reasons for a waiver include the following

- a State standard has not been consistently applied in similar circumstances
- it is an interim action
- compliance will result in greater risk to human health and the environment than alternative options
- compliance is technically impracticable
- the selected remedial action will attain a standard equivalent to an applicable standard using another approach

Since the State of Colorado is authorized by EPA to implement the RCRA program the RCRA ARARs under the State program are designated as key ARARs Releases and spills at OU 1 occurred prior to the effective date of the RCRA regulations and therefore the RCRA program regulations are designated relevant and appropriate to the substances (spent solvents and contaminated media) and site circumstances at OU 1

Compliance with the RCRA program involving releases of hazardous constituents from solid waste management units (SWMUs) under Subpart F is a relevant and appropriate requirement for all alternatives In addition, the Corrective Action Management Unit (CAMU) Subpart S rule recently adopted by the State (264 552 of 6 CCR 1007 3) is a relevant and appropriate requirement for all alternatives It is a relevant and appropriate requirement because this rule allows remediation wastes to remain in place after closure of the CAMU providing certain requirements are met by the owner The definition of remediation wastes is solid and hazardous wastes and media that contain listed hazardous wastes or which exhibit a hazardous waste characteristic that are managed for the purpose of implementing corrective action requirements "

Requirements of the owner of a CAMU are

- 1) siting of the CAMU is to be in accordance with the requirements for siting

hazardous waste disposal sites under 6 CCR 1007 2 Part 2 (solid waste minimum standards 2 1) and

- 2) Subparts B C D and E of Part 264 or 265 are to be met Subpart B is focused on general inspections Subpart C is the preparedness and prevention provisions Subpart D is the contingency plan and Subpart E is the record keeping provisions Subpart E is an administrative requirement and not an ARAR

By designating the unit a CAMU CDPHE will facilitate implementation of a reliable effective protective and cost-effective remedy (criteria of an NCP/CERCLA selection) CDPHE may specify any closure post-closure and any groundwater monitoring or long term maintenance activities as part of the designation (6 CCR 1007 3, Part 264 552(e)) Waste management activities associated with CAMU cannot create unacceptable risks to humans or the environment Since the documentation of the CDPHE designation is required to be made public according to the CAMU rule it is assumed that any designation of the CAMU will appear in the Proposed Remedial Action Plan/Proposed Plan (PRAP/PP) and the CAD/ROD

4 1 3 Long Term Effectiveness and Permanence

In addition to the specific statutory requirements discussed in Section 4 1 2 CERCLA guidance emphasizes the preference for treatment to achieve long term protection and permanence for the proposed remedy Criteria for evaluating long term effectiveness and permanence include the following

- persistence, toxicity, and mobility of hazardous substances and their constituents and their tendency to bioaccumulate
- long term uncertainties associated with containment
- long term potential for adverse health effects from human exposures
- long-term cost of monitoring and maintenance
- ease of undertaking future remedial action should the proposed alternative fail

These considerations are focused on the magnitude of residual risk remaining after the response objectives have been met. The evaluation of the proposed alternative must include an analysis of the continued potential threat to human health and the environment from untreated waste or treatment residuals remaining at the site after corrective action has been taken. This analytical process includes the following elements:

- volume and concentration of contaminants in untreated media
- volume and concentration of contaminants in treated residuals
- requirements for five-year site reviews and long term monitoring
- difficulties associated with long term operations and maintenance
- confidence in the adequacy of controls
- availability of equipment used in the alternatives
- habitat value following remedial actions as compared to existing habitat

4.1.4 Reduction of Toxicity, Mobility, or Volume (TMV) Through Treatment

This criterion evaluated the ability of a remedial action alternative to reduce the risks at a site through the destruction of toxic contaminants, reduction of the total mass of toxic contaminants, reduction in contaminant mobility, or reduction in the total volume of contaminated media. The NCP states a preference for remedial alternatives that include treatment as a principal element of the remedy over those that do not. Specific considerations include the following:

- adequacy of the treatment process to address preliminary remediation goals
- specific requirements and limitations of the treatment process
- volume of the contaminated media that are treated
- extent of reduction in TMV
- irreversibility of the treatment
- quantities and toxic characteristics of the treatment residuals or by-products

4.1.5 Short Term Effectiveness

This criterion addresses the period of time during the construction and implementation of the remedy. The evaluation covers community protection and site-worker protection during the remediation period as well as any potential adverse environmental impacts that may result from

construction and implementation The consideration of environmental impacts during the remediation period also includes an evaluation of the impact of the remedial action on the quality of habitat

4 1 6 Implementability

This criterion addresses the technical and administrative feasibility of implementing a remedy including the availability of materials and services needed during its implementation Implementability is particularly important for evaluating the reliability of technologies that are less proven and when evaluating remedies that are dependent on a limited supply of equipment vendors or specialists Specific considerations include the following

- ability to construct and operate the alternative within a 10- to 30-year time frame
- availability of equipment and specialists
- availability and reliability of the components of the alternative
- ability to monitor the effectiveness of the alternative
- demonstrated performance level of the treatment components and equipment
- difficulty in implementing future remedial actions once the alternative is in place

The implementability evaluation also addresses the requirements for coordination with local state and federal offices and agencies to obtain necessary permits

4 1 7 Cost

This criterion addresses the evaluation of the capital cost for each alternative as well as the long term operation and maintenance (O&M) expenditures required to sustain it Present worth cost analysis is used to compare expenditures that occur over different time periods By discounting all costs to a common base year the cost of each alternative can be reduced to a single figure for comparative analysis To calculate the present worth of each alternative this report assumes

a discount interest rate of 5 percent (as specified in the CMS/FS guidance) and an implementation period of 30 years for long term O&M or the actual implementation period if it is less than 30 years

Cost may play a significant role in differentiating options that appear comparable with respect to long term effectiveness and permanence or when choosing among treatment options that provide similar performance. An alternative with a cost that is excessive when compared to overall effectiveness may not be feasible to implement as a final remedy. Also, an alternative with a low initial capital cost may be more costly overall when the O&M costs are considered. Higher cost may be offset by improved performance or greater long term risk reduction in the comparative analysis of alternatives. Ultimately, however, the remedial alternative that satisfies the CERCLA requirements in the most cost-effective manner will be selected as the preferred alternative.

4.1.8 State Acceptance

State acceptance refers to the state or support agency's comments on the appropriateness of the proposed remedy. The state's position and key concerns about the preferred alternative should be assessed as early in the process as practicable.

4.1.9 Community Acceptance

Community acceptance evaluates the issues and concerns raised by the general public in their response to the alternatives described in the CMS/FS report. Interested persons or groups in the community may support, have reservations about, or oppose some components of the preferred remedial alternative, and their concerns may influence the final selection process.

4.2 Detailed Analysis of Alternatives

Detailed analysis of alternatives is accomplished in this report by evaluating the two threshold

and five balancing criteria for each alternative. The analysis is conducted at a level of detail that builds on the information presented in Section 3 sufficient to provide an understanding of each alternative and any uncertainties associated with the evaluation. Key trade-offs with respect to the criteria are identified for the alternatives. According to the CMS/FS guidance, the results of the detailed analysis are designed to provide the basis for identifying a preferred alternative for remedial action.

Assumptions used in performing the detailed analysis of this CMS/FS include the following:

- DNAPLs are potentially present in the subsurface at IHSS 119.1, based on the results of the Phase III RFI/RI report. If present, it is assumed that they are primarily in residual form and in small quantities.
- Groundwater monitoring proposed under each remedial alternative would include sampling and analysis at the french drain, the existing groundwater extraction well, and potentially four new monitoring wells at OU 1. The locations would be sampled semiannually and analyzed for both organic and inorganic COCs.
- A soil gas survey would be conducted prior to initiating any of the proposed treatment actions to more accurately define the areas at IHSS 119.1 requiring treatment. For purposes of the detailed analysis, only the previously identified source area is considered.

In the comparative analysis, a qualitative sensitivity analysis is performed to assess the major assumptions which, if incorrect, could significantly impact the results of the detailed analysis of the alternatives.

Groundwater Monitoring

Groundwater monitoring is included as part of each remedial action alternative presented herein. For the purposes of the detailed analysis of alternatives, it is assumed that six monitoring points would be used for performance monitoring of each alternative. Four new wells would be installed, one deep and shallow well cluster downgradient of IHSS 119.1, and possibly two additional wells upgradient of Woman Creek. It is suggested that placement of the well cluster

be preceded by geological and geophysical support such as photographic lineament analysis, and/or three-dimensional seismic surveys. This would enable paleochannels and faulted zones to be clearly identified prior to well placement.

Samples would also be collected from the french drain sump and from the existing recovery well. Samples would be analyzed for organic and inorganic contaminants and would be collected semiannually. Analysis of individual species of inorganic contaminants is also suggested to identify individual metal species which have the potential to bioaccumulate. This additional analysis requirement should only be applied occasionally in the sampling program. PQLs would be used to determine compliance with CDPHE standards.

Groundwater Modeling

To support the detailed analysis of remedial action alternatives, groundwater modeling was performed to predict downgradient contaminant concentrations resulting from suspected sources at IHSS 119.1. Three conceptual models were identified and used to predict future COC concentrations at Woman Creek. The no action scenario was used to examine contaminant migration patterns excluding source removal and the existing french drain and extraction well (Alternatives 0 and 1). The french drain institutional control scenario was used to examine contaminant migration patterns with the french drain and extraction well in place (Alternatives 2 and 3). The remediation scenario was used to examine the effect of remediating all of the suspected sources within IHSS 119.1 to MCLs and to predict downgradient concentrations once this goal was achieved (Alternatives 4, 5, 6, and 7).

The groundwater model is described in detail in Appendix B. In general, the computer simulation code TARGET_2DU (Dames & Moore 1985) was used to simulate contaminant transport in the subsurface. TARGET_2DU is a vertically oriented finite difference model that can simulate variably saturated conditions. The model will be available for public use late in 1994. This model was selected due to the variability of the saturated zone at OU 1 and because it has been successfully applied at other Superfund sites to support final CADs and RODs.

Detailed assumptions and uncertainties associated with the model are included in Appendix B. Some of the major assumptions and/or uncertainties that identify conservative points in the model are summarized below:

- The model is two dimensional therefore dispersion (spreading) in lateral to the plane of the model is not simulated. This causes over prediction of concentrations.
- The model does not account for decay of contaminants adsorbed to soil. If desorption occurs, then concentrations are conservatively over predicted.
- The model does not account for volatilization of organic contaminants. It is likely that volatilization is an important process because of high volatilization rates for these chemicals (high Henry's Law constants) and because of the short distance from groundwater to land surface.
- The model predicts increasing concentrations at locations like Well 0487 and 4387 where observed concentrations fluctuate around a generally constant average. This is most likely due to the way in which desorption is simulated and to ignoring the effects of volatilization.

In examining the results of the modeling effort, PCE was selected as the indicator chemical for OU 1. The MCL for PCE is 5×10^{-3} mg/l. PCE concentrations at Woman Creek were at a maximum at the end of the modeled time period for the no action and french drain institutional control conceptual models where they appeared to be approaching an asymptotic value near the maximum concentration predicted. Under the remediation conceptual model peak concentrations occurred within the modeled time period. The peak concentrations predicted for PCE under each conceptual model (with alternative numbers identified) are listed below along with the year in which the peak concentration was observed:

- no action scenarios (0, 1) 3.60×10^{-3} mg/l in 2369
- french drain institutional control scenarios (2, 3) 8.62×10^{-4} mg/l in 2269
- remediation scenarios (4, 5, 6 and 7) 5.84×10^{-4} mg/l in 2152 (30-year average at peak) with 5.94×10^{-4} mg/l as the actual peak concentration

These conceptual models were used to estimate residual risk levels associated with the various

remedial action alternatives proposed in this section. Peak concentrations for other COCs were several orders of magnitude below that of PCE.

Residual Risk Assessment

The risk assessment presented in Appendix C documents the approach and calculations performed to estimate residual risks associated with the proposed alternatives. To select the most appropriate pathways and indicator chemical, the results of the OU 1 PHE were first reviewed. Groundwater modeling results were then compared to contaminant specific PRGs for OU 1. This comparison indicated that PCE is the most conservative contaminant to use in the risk assessment, that is, it contributes the highest risk to future groundwater receptors based on modeled contaminant concentrations at Woman Creek.

Groundwater modeling was performed to estimate the concentration of PCE in groundwater using three conceptual models for OU 1, as described in the modeling summary above.

Using groundwater modeling results with the most conservative exposure pathways, noncarcinogenic hazard indices and carcinogenic risk were calculated. The results of these calculations indicate that none of the calculated noncarcinogenic hazard indices approach unity and that the maximum calculated carcinogenic risk is for the scenario of no remediation of the source contaminant and discontinuing operations of the french drain and extraction well, (i.e. no action). The maximum risks are listed below for each modeled scenario with alternative numbers listed in parenthesis.

- no action scenarios (0, 1) 1.99×10^{-6}
- french drain institutional control scenarios (2, 3) 4.76×10^{-9}
- remediation scenarios (4, 5, 6, 7) 3.22×10^{-7}

4.2.1 Alternative 0. No Action

The NCP requires that the No Action Alternative be evaluated as a baseline alternative against

which other alternatives can be compared. This alternative assumes that acceptable groundwater contaminant concentrations would be achieved through natural degradation and dispersion of the groundwater COCs at OU 1, and that the site would eventually be abandoned. Therefore, no remedial actions would be initiated to reduce the risk from groundwater contaminants by actively treating the groundwater or subsurface soils. The alternative assumes that operation of the treatment portion of the existing french drain system would be discontinued. For costing purposes, it is assumed that the french drain would be decommissioned under this alternative. This would be accomplished by using a backhoe to excavate and remove sections of the drain.

Groundwater monitoring activities would continue to monitor contaminant concentrations over time. For the purposes of this detailed analysis, a 30-year institutional control period is assumed for groundwater monitoring.

The evaluation of the two threshold and five balancing criteria for Alternative 0 - No Action are summarized as follows:

Overall Protection of Human Health and the Environment

The no action alternative would be protective of human health based on exposure to OU 1 COCs at the Woman Creek location. Concentrations of contaminants in groundwater would gradually be reduced over time due to physical, biological, and chemical processes such as dispersion, volatilization, and biodegradation.

Key ARARs would be met under this alternative. In particular, the MCLs would continue to be achieved for groundwater COCs at Woman Creek. This alternative would provide long-term effectiveness in achieving the MCLs through natural processes which are essentially irreversible.

A risk level of 1.99×10^{-6} would be achieved at Woman Creek under this alternative. Therefore,

the magnitude of residual risk that would result from the implementation of the No Action alternative falls well within the acceptable risk range of 10^{-4} to 10^{-6} . Additionally, risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6} .

There would be no increase in potential risks to the public, to on-site workers, or to the environment under the No Action alternative.

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling results projected to 400 years from 1969 also indicate that there will be no exceedance of MCLs within the 400-year period. Groundwater modeling results demonstrate that the highest concentration of PCE during the 400-year period is 3.60×10^{-3} mg/l for this alternative. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

Alternative 0 will comply with RCRA regulations 6 CCR 1007.3 Parts 264 and 265 if the CDPHE designates OU 1 a CAMU in accordance with 6 CCR 1007.3, 264.552(c). Under the No Action alternative, the contaminants would remain in place subject to natural degradation, volatilization, and dispersion. Groundwater monitoring would be conducted to detect direction and movement of hazardous constituents as specified in 6 CCR 1007.3, 264.552(d)(3). The substantive requirements for disposal facilities under Parts 264 and 265 would be met through compliance with the CAMU rule.

The No Action alternative would comply with the state solid waste disposal site and facility regulations since there are no exceedances of the MCLs at Woman Creek. However, since solid

waste areas can be included in the CAMU it is believed this requirement is no longer appropriate as an ARAR

The air pollution regulations are not an ARAR for this alternative since there are no major earth moving activities or air emissions associated with this alternative

Location Specific ARARs

Alternative 0 is anticipated to comply with the laws and regulations specific to wetlands and species which use the wetlands. There is a population of Preble's meadow-jumping mouse at the RFETS which is a non game species of special concern under state wildlife policy. Once the french drain is decommissioned it is possible that wetland/riparian habitat areas would increase in size in the long term after a short term disturbance. Prior to disturbance the State Division of Wildlife would be consulted on mitigation measures to lessen impacts to this species as well as others.

Long Term Effectiveness and Permanence

The No Action alternative would involve only groundwater monitoring. This alternative would not provide any additional protection for the environment and potential downgradient receptors because operation of the french drain which currently appears to be effective in capturing groundwater migrating away from OU 1 would be discontinued under this alternative.

Residual concentrations of COCs may be acting as a continuing source at IHSS 119.1. The No Action alternative does not address treatment for residuals in either the groundwater or the soil. The existing french drain system would not be operational, potentially allowing contaminated groundwater to migrate from OU 1 and to impact groundwater and soils outside of OU 1. However modeling indicates that under this alternative MCLs for groundwater COCs are currently achieved and would continue to be met at Woman Creek. A five year review would be conducted to determine the continued effectiveness of this alternative.

Reduction of Toxicity, Mobility, or Volume Through Treatment

This alternative would not satisfy the NCP preference for treatment as a principal element of the alternative because it does not directly result in treatment of contaminated groundwater. The No Action alternative would not treat groundwater or soils and therefore would not reduce the toxicity, mobility, or volume of contaminants, although natural processes would result in a reduction of contaminant concentrations over time.

Short Term Effectiveness

Because no remedial action would be initiated, no significant additional short term risks to the local community or environment would be created as a result of the No Action alternative at OU 1.

There would be no additional potential impacts to workers as a result of this alternative. Existing safety measures used for permanent workers and visitors would offer effective and reliable protection from the COCs associated with OU 1. Adherence to appropriate health and safety measures would be required for as long as monitoring activities are continued at OU 1.

Implementability

The No Action alternative is readily implementable because it includes only the continuation of groundwater monitoring activities with installation of possibly four additional wells. The implementability of this alternative would not be limited by the availability of services and materials, nor would there be any significant technical or administrative difficulties associated with this alternative.

Cost

Capital costs associated with Alternative 0 include decommissioning the french drain. Post

closure activities for Alternative 0 include groundwater monitoring for 30 years and installation of four additional wells. The capital cost for this alternative is \$154,700. The annual O&M cost for this alternative is \$0. The cost for post-closure is \$1,740,400. The total cost for this alternative is \$1,895,100. A detailed cost estimate for this alternative is included in Appendix E.

4.2.2 Alternative 1. Institutional Controls Without the French Drain

This alternative would rely on institutional controls to restrict access to any wells impacted by OU 1 contaminants and prevent building construction above the areas known to be contaminated with VOCs. As with the No Action alternative, this alternative assumes that acceptable groundwater contaminant concentrations would be achieved through natural degradation, volatilization, and dispersion of the COCs, and that the site would not be abandoned during the institutional control period. No remedial actions would be initiated to reduce the risk from groundwater contaminants by actively treating the groundwater or subsurface soils. The alternative assumes that operation of the treatment portion of the existing French drain system would be discontinued.

This alternative presents the potential for the RFETS to be converted to a future ecological reserve. The institutional controls considered here represent sitewide control of all areas of the RFETS. Groundwater monitoring, supplemented by installation of additional wells, would be continued to determine if any changes occur in contaminant concentrations or in contaminant migration patterns. Groundwater monitoring would continue for as long as institutional controls are active at the site, or until it is determined that monitoring is no longer required. For the purposes of this detailed analysis, a 30-year institutional control period is assumed for groundwater monitoring.

The evaluation of the two threshold and five balancing criteria for Alternative 1, Institutional Controls Without the French Drain, are summarized as follows:

Overall Protection of Human Health and the Environment

This alternative would be protective of human health assuming that the institutional controls are properly implemented and that the site would not be abandoned during the institutional control period. The french drain would not be used to capture contaminated groundwater but concentrations of contaminants in downgradient groundwater would gradually be reduced over time due to physical and chemical processes, such as dispersion volatilization and biodegradation.

There would be no additional human health risk associated with this alternative because the entire RFETS site would be controlled eliminating access to OU 1 and therefore the potential for human exposure. There would also be no increase in potential risks to the public or to on site workers under this alternative.

Key ARARs would be met under this alternative. In particular the MCLs would continue to be achieved for groundwater COCs at Woman Creek. This alternative would provide long term effectiveness in achieving the MCLs, through natural processes which are essentially irreversible.

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling projected 400 years from 1969 also indicates that there will be no exceedance of MCLs within the 400-year period. Groundwater modeling results demonstrate that the highest concentration of PCE during the 400-year period is 3.60×10^{-3} mg/l for this alternative. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

The main difference between the No Action alternative and Alternative 1 is the extent of institutional controls at the RFETS. This difference would not create a difference in the alternative's compliance with the key ARARs. Alternative 1 will meet substantive requirements of the State RCRA program if CDPHE staff designates the OU 1 area as a CAMU under the recently adopted Subpart S provisions (6 CCR 1007.3 Part 264 Section 552). This alternative includes a 30-year groundwater monitoring program which complies with 6 CCR 1007.3 264.552(d)(3). Contaminants would be left in place subject to natural degradation, volatilization and dispersion.

Solid waste can be included in the CAMU and thus it is assumed that substantive portions of the State's solid waste regulations would not be appropriate to the CAMU and thus not an ARAR.

The air pollution regulations are not an ARAR for this alternative since there would not be any air emission sources or major earth moving activities.

Location Specific ARARs

Location specific ARARs associated with this alternative are focused on the protection of wetlands. Decommissioning of the french drain could cause disturbance to a small portion of wetlands for 2 to 3 days. Mitigation measures would be used to minimize the impact and to comply with DOE regulations on wetland protection as well as the State's Non Game Endangered or Threatened Species Conservation Act. Coordination with State Division of Wildlife would be done to protect the population of Preble's meadow jumping mouse, a state species of special concern.

Long Term Effectiveness and Permanence

This alternative would minimize the human health risk associated with contaminated groundwater.

by restricting access to any wells impacted by OU 1 contaminants and by eliminating the possibility of building construction above areas known to be contaminated with VOCs. This alternative would not provide any additional protection for the environment and potential downgradient receptors because operation of the french drain which currently appears to be effective in capturing groundwater migrating away from IHSS 119 1 would be discontinued under this alternative.

This alternative does not address treatment for residuals in either the groundwater or subsurface soils. The existing french drain system would not be operational, potentially allowing contaminated groundwater to migrate from OU 1, and to impact groundwater and soils outside of OU 1. However, modeling results indicate that under this alternative groundwater would continue to meet MCLs at Woman Creek. A five year review would be conducted to determine the continued effectiveness of this alternative.

Reduction of Toxicity, Mobility, or Volume Through Treatment

This alternative would not satisfy the NCP preference for treatment as a principal element of the alternative. Because it does not propose treatment for contaminated groundwater or subsurface soils, this alternative would not reduce the toxicity, mobility, or volume of contaminants, although natural processes would result in a reduction of contaminant concentrations over time.

Short Term Effectiveness

Because no remedial action would be initiated, no additional short term risks to the local community or environment would be created by implementing this alternative.

There would be no additional potential impacts to workers as a result of this alternative. Existing safety measures used for permanent workers and visitors would offer effective and reliable protection from the COCs associated with OU 1. Adherence to appropriate health and safety measures would be required for as long as monitoring activities are continued at OU 1.

Implementability

This alternative is readily implementable because it includes only institutional controls and groundwater monitoring. The implementability of this alternative would not be limited by the availability of services and materials nor would there be any significant technical difficulties associated with this alternative. Institutional controls proposed under this alternative such as deed restrictions could be implemented with no significant administrative problems.

Cost

Capital costs associated with Alternative 1 include decommissioning the french drain. Post closure activities for Alternative 1 include groundwater monitoring for 30 years and installation of four additional wells. The capital cost for this alternative is \$154,700. The annual O&M cost for this alternative is \$0, and the cost for post-closure is \$1,740,400. The total cost for this alternative is \$1,895,100. A detailed cost estimate for this alternative is included in Appendix E.

4.2.3 Alternative 2. Institutional Controls With the French Drain

This alternative is similar to Alternative 1 except that operation of the french drain and the Building 891 water treatment system would be continued. The french drain would continue to capture contaminated groundwater migrating from the IHSS 119.1 source area. Dilute concentrations of contaminated groundwater to the east of the operable unit would not be actively remediated by this alternative. As with Alternative 1, institutional controls would be utilized to restrict access to any wells impacted by OU 1 contaminants and prevent building construction above the areas known to be contaminated with COCs.

This alternative would also utilize groundwater monitoring programs to determine if any changes occur in contaminant concentrations or in contaminant migration patterns. Groundwater monitoring would continue for as long as institutional controls are active at the site or until it

is determined that monitoring is no longer required For the purposes of detailed analysis a 30 year institutional control period is assumed for groundwater monitoring The french drain would be decommissioned after monitoring confirms that COC concentrations have been reduced to acceptable levels

The evaluation of the two threshold and five balancing criteria for Alternative 2 Institutional Controls With the French Drain are summarized as follows

Overall Protection of Human Health and the Environment

Similar to Alternative 1 this alternative would be protective of human health assuming that the institutional controls are properly implemented and that the site would not be abandoned during the institutional control period In this alternative the french drain would be used to capture and treat contaminated groundwater and prevent downgradient migration of COCs

Key ARARs would be met under this alternative In particular the MCLs would continue to be achieved for groundwater COCs at Woman Creek This alternative would provide long term effectiveness in achieving the MCLs although the drain would have to be operated until all sources have been remediated in IHSS 119 1

A risk levels of 4.76×10^{-9} would be achieved at Woman Creek under this alternative Therefore the magnitude of residual risk that would result from the implementation of this alternative falls well below the acceptable risk range of 10^{-4} to 10^{-6} Additionally risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6} There would no increase in potential risks to the public or to on site workers under this alternative because no additional actions would be initiated Existing health and safety procedures would effectively protect on site workers

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling results projected 300 years from 1969 also indicate that there will be no exceedance of MCLs within the 300-year period. Groundwater modeling results demonstrate that the highest peak concentration of PCE during the 300-year period is 8.62×10^{-6} mg/l under this alternative. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

Alternative 2 will meet the substantive requirements of the state RCRA program specific to past spilled waste, if CDPHE staff designates the OU 1 area as a CAMU under 6 CCR 1007.3-264.552.

The substantive requirements include

- Generators of hazardous waste
- Releases from solid waste management units
- Closure of a disposal facility under interim status and final status
- Post-closure of a disposal unit under final status

Solid waste is allowed to be included in the designation of a CAMU. If the CAMU is adopted, it is believed the state's solid waste regulations would not be appropriate to the circumstances and thus would no longer be an ARAR.

Contaminants would be left in place at the IHSS 119.1 area subject to collection and treatment at the french drain. There is a potential for some contaminants to be left in place outside the area of the capture zone and collection system.

The State's air pollution control regulations are not an ARAR for this alternative since there would not be a source of air emissions

Location Specific ARARs

Alternative 2 will not involve disruption of the current scenario at the site for an estimated 30 years. Therefore, impacts to wetland and riparian habitat areas are not anticipated to occur within this time period. The populations of Preble's meadow jumping mouse, a state species of special concern, would continue on their current course until that time.

Decommissioning of the french drain would involve a short term disruption to some portion of wetlands. Mitigation measures would be used to minimize the impact and to comply with DOE regulations on wetland protection as well as the State's Non Game Endangered or Threatened Species Conservation Act.

Long Term Effectiveness and Permanence

Under this alternative, the existing french drain would continue to remove contaminated groundwater migrating from IHSS 119.1. This alternative would also minimize the human health risk associated with contaminated groundwater by restricting access to any wells impacted by OU 1 contaminants and by eliminating the possibility of building construction above areas known to be contaminated with VOCs. The alternative would provide long term protection for potential human receptors assuming that the institutional controls initiated are guaranteed to remain in place.

Residual concentrations of COCs may be acting as a continuing source at IHSS 119.1. This alternative does not address treatment for residuals in either the groundwater or subsurface soils. However, modeling indicates that under this alternative, groundwater would continue to meet MCLs for the COCs at Woman Creek. Concentrations of contaminants in groundwater would gradually be reduced over time due to physical and chemical processes such as dispersion.

volatilization and biodegradation. A five-year review would be conducted to determine the continued effectiveness of this alternative.

Contaminated materials generated as a result of this alternative include spent ion exchange resins from the Building 891 water treatment system. These resins are currently regenerated on site. There are no significant risks associated with handling these resins.

Reduction of Toxicity, Mobility, or Volume Through Treatment

The continued operation of the french drain would reduce the mobility and volume of COCs in groundwater at OU 1; however, residual concentrations of COCs would remain in subsurface soils and groundwater in IHSS 119.1.

Extracted groundwater would be treated in the Building 891 water treatment system. This is a destructive treatment process and thus would result in decreased toxicity. Contaminant removal through groundwater extraction would be irreversible; however, DNAPLs in IHSS 119.1 may continue to act as a source.

Short Term Effectiveness

This alternative includes the implementation of institutional controls, the continued operation of the french drain system, and groundwater monitoring. Because no additional remedial action would be initiated, no additional short term risks to the local community or environment would be created by implementing this alternative.

There would be no additional potential impacts to workers as a result of this alternative. Existing safety measures used for permanent workers and visitors would offer effective and reliable protection from the COCs associated with OU 1. Adherence to appropriate health and safety measures would be required for as long as monitoring activities are continued at OU 1.

Implementability

This alternative is readily implementable because it includes only institutional controls the continued operation of the french drain and Building 891 water treatment system and the continuation of groundwater monitoring activities. The implementability of this alternative would not be limited by the availability of services and materials nor would there be any significant technical difficulties associated with this alternative. There are no technical problems associated with continued operation of the existing french drain system. Groundwater monitoring would effectively track any additional migration of COCs.

Institutional controls proposed under this alternative such as deed restrictions could be implemented with no significant administrative problems.

Cost

Capital costs associated with Alternative 2 include decommissioning of the french drain. O&M costs for Alternative 2 include operation of the Building 891 water treatment system and groundwater monitoring for 30 years. Capital cost for this alternative is \$149,600. The annual O&M cost for this alternative is \$15,603,300, the post-closure cost is \$1,740,400. The total cost for this alternative is \$17,493,300. A detailed cost estimate for this alternative is included in Appendix E.

4.2.4 Alternative 3. Modified French Drain With Additional Extraction Wells

Similar to Alternative 2, this alternative would continue to operate the existing french drain system. However, approximately four to six additional six inch diameter groundwater extraction wells would be added to the existing system to enhance its effectiveness in capturing contaminated groundwater migrating from the IHSS 119.1 sources and other areas. The additional wells could be installed in the southeastern corner of the operable unit to capture any contaminated groundwater potentially flowing around the french drain in the IHSS 119.1 source.

area to assist the existing recovery well in front of the french drain in a suspected low permeability sandstone lens which may potentially form a conduit for groundwater transport beneath the drain or in the area south of Building 881 to enhance the recovery of contaminated groundwater in that area

Based on the performance of the french drain and existing extraction well the new extraction wells are expected to remove less than one gpm of groundwater combined Each well would be equipped with a sump pump with high and low level switches that would pump intermittently This would ensure that the pumps would not be burned out by operating in a dry well Groundwater recovered from the new extraction wells would be routed to the french drain system to be transferred to the Building 891 water treatment system and treated with the UV/peroxide/ion exchange process

As with the institutional controls alternatives, this alternative would also continue groundwater monitoring activities to determine if any changes occur in contaminant concentrations or in contaminant migration patterns Groundwater monitoring would continue for as long as institutional controls are active at the site or until it is determined that monitoring is no longer required For the purposes of this detailed analysis a 30-year institutional control period is assumed for groundwater monitoring The french drain would be decommissioned after monitoring confirms that COC concentrations have been reduced to acceptable levels

The evaluation of the two threshold and five balancing criteria for Alternative 3 Modified French Drain With Additional Extraction Wells are summarized as follows

Overall Protection of Human Health and the Environment

This alternative would be protective of human health and the environment assuming that the site would not be abandoned during the institutional control In this alternative, the french drain and several additional extraction wells would be used to capture contaminated groundwater and prevent downgradient migration of COCs Low level groundwater contamination east of the

french drain would also be addressed in this alternative by installing one or more of the new extraction wells in this area

Key ARARs would be met under this alternative. Specifically, MCLs would continue to be achieved for groundwater COCs at Woman Creek. This alternative would provide long term effectiveness in achieving the MCLs.

Risk levels achieved by this alternative would be the same as those achieved under Alternative 2, however these levels would be achieved in a shorter time because the additional groundwater extraction wells would increase the rate at which contaminated groundwater is removed. Therefore, the magnitude of residual risk that would result from the implementation of Alternative 3 falls well below the acceptable risk range of 10^{-4} to 10^{-6} . Additionally, risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6} .

There would be no increase in potential risks to the public or to on site workers under this alternative. Existing health and safety procedures would effectively protect on-site workers.

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling results projected 300 years from 1969 also indicate that there will be no exceedance of MCLs within the 300-year period. Groundwater modeling results demonstrate that the highest concentration of PCE during the 300-year period is 4.76×10^{-9} mg/l for this alternative. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

Alternative 3 compliance with action-specific RCRA program requirements is similar to that described under Alternative 2. RCRA requirements are relevant and appropriate and compliance of the alternative with the substantive requirements of Parts 262, Part 264, and Part 265 depends in part on implementation of Subpart S requirements. OU 1 designated as a CAMU would allow some contaminants to be left in place after closure at the 119.1 and 119.2 IHSS locations. Compliance with releases from solid waste management units Subpart F, closure of a disposal unit Subpart G, and post-closure of a disposal unit Subpart H can be achieved with the alternative.

Solid waste can be designated as part of the CAMU and therefore compliance with the State solid waste regulations would not be appropriate. The MCLs are met at Woman Creek according to groundwater modeling results.

The additional extraction wells would be constructed in accordance with the substantive parts of the Colorado Water Well and Pump Installation Regulations (2 CCR 402.2). Therefore, compliance with this ARAR is anticipated for Alternative 3.

Location-Specific ARARs

Alternative 3 is anticipated to comply with the location specific ARARs. Laws and regulations specific to wetlands and species which use wetlands will be complied with if Alternative 3 is implemented at OU-1. When the french drain is decommissioned, it is possible that the wetland/riparian habitat areas would increase in size in the long term, after short term disturbance of the wetland areas. Prior to disturbance, the State Division of Wildlife would be consulted on mitigation measures, to lessen impacts to species of concern such as Preble's meadow jumping mouse.

Long Term Effectiveness and Permanence

Under this alternative the existing french drain would continue to remove contaminated groundwater migrating from IHSS 119 1 Therefore the residual risk would gradually be reduced over time

This alternative would also minimize the human health risk associated with contaminated groundwater by restricting access to any wells impacted by OU 1 contaminants and by eliminating the possibility of building construction above areas known to be contaminated with VOCs The alternative would provide long term protection for potential human receptors assuming that the institutional controls initiated are guaranteed to remain in place

This alternative would also be effective in protecting the environment over the long term by addressing groundwater with low levels of contaminants that is potentially flowing around the east end of the french drain system and could otherwise potentially impact on groundwater and soils outside of OU-1 However, modeling indicates that under this alternative groundwater would continue to meet MCL for PCE at Woman Creek Concentrations of contaminants in groundwater would gradually be reduced over time due to physical and chemical processes such as dispersion volatilization and biodegradation A five year review would be conducted to determine the effectiveness of this alternative

Contaminated materials generated as a result of this alternative include spent ion exchange resins from the Building 891 water treatment system These resins are currently regenerated on site There are no significant risks associated with handling these resins

Reduction of Toxicity, Mobility, or Volume Through Treatment

This alternative would reduce the mobility and volume of contaminants in OU 1 through continued operation of the french drain and the addition of several groundwater extraction wells The french drain effectively reduces the migration of contaminated groundwater from OU 1

Groundwater extraction would reduce the volume of COCs in groundwater however residual concentrations of COCs would remain in subsurface soils and groundwater in IHSS 119 1 The french drain and extraction wells would also continue to prevent migration of COCs beyond OU 1 thus reducing the mobility of these contaminants

Extracted groundwater would be treated in the Building 891 water treatment system This is a destructive treatment process and thus would result in decreased toxicity Contaminant removal through groundwater extraction would be irreversible, however, DNAPLs may continue to act as a source

Short Term Effectiveness

Because the only additional remedial action initiated by this alternative would be the extraction well installation, a procedure that can be accomplished in a relatively short period of time the additional short term risks to the local community or environment would be minimized

The only potential impacts to workers implementing this alternative would be those associated with the installation of the new extraction wells These potential impacts would be minimized through existing worker safety procedures governing construction activities at RFETS Existing safety measures for the french drain operation and monitoring activities would offer effective protection for workers and visitors at OU 1 Adherence to appropriate health and safety measures would be required for as long as monitoring activities are continued at OU 1

Implementing this alternative would have limited impacts on the environment or the public at OU 1 Installing additional extraction wells would have minor impacts on site soils and flora

Implementability

This alternative is also readily implementable It includes the continuation of groundwater monitoring activities the continued operation of the french drain, and installation of several

additional groundwater extraction wells. The implementability of this alternative would not be limited by the availability of services and materials nor would there be any significant technical or administrative difficulties associated with this alternative. Continuation of groundwater monitoring activities would effectively track any additional COC migration. Also, implementing this alternative would not limit the ability to perform future remedial actions if they are determined to be necessary.

Extraction wells can be installed with little difficulty using standard drilling techniques and standard construction materials that are commonly used for well construction and are readily available. Operating the additional extraction wells would not require any additional specialized personnel or training. The Building 891 water treatment system has sufficient capacity to treat the quantities of water extracted by the additional wells.

No administrative difficulties are anticipated under this alternative. Coordinating activities with agencies and obtaining the appropriate permits is not expected to present any problems.

Cost

Capital costs for this alternative include the costs for four new groundwater extraction wells (six inch diameter, 20-foot depth) and associated piping, pumps, and instrumentation. The capital cost for Alternative 3 is \$305,000. O&M and post-closure activities for Alternative 3 include the operation of the additional extraction wells and the Building 891 water treatment system and groundwater monitoring. The present worth value for O&M for Alternative 3 is \$15,603,300; the post-closure cost is \$1,740,400. The total cost of this alternative is therefore \$17,648,700. A detailed cost estimate for capital and O&M costs for this alternative is included in Appendix E.

4.2.5 Alternative 4. Groundwater Pumping and Soil Vapor Extraction

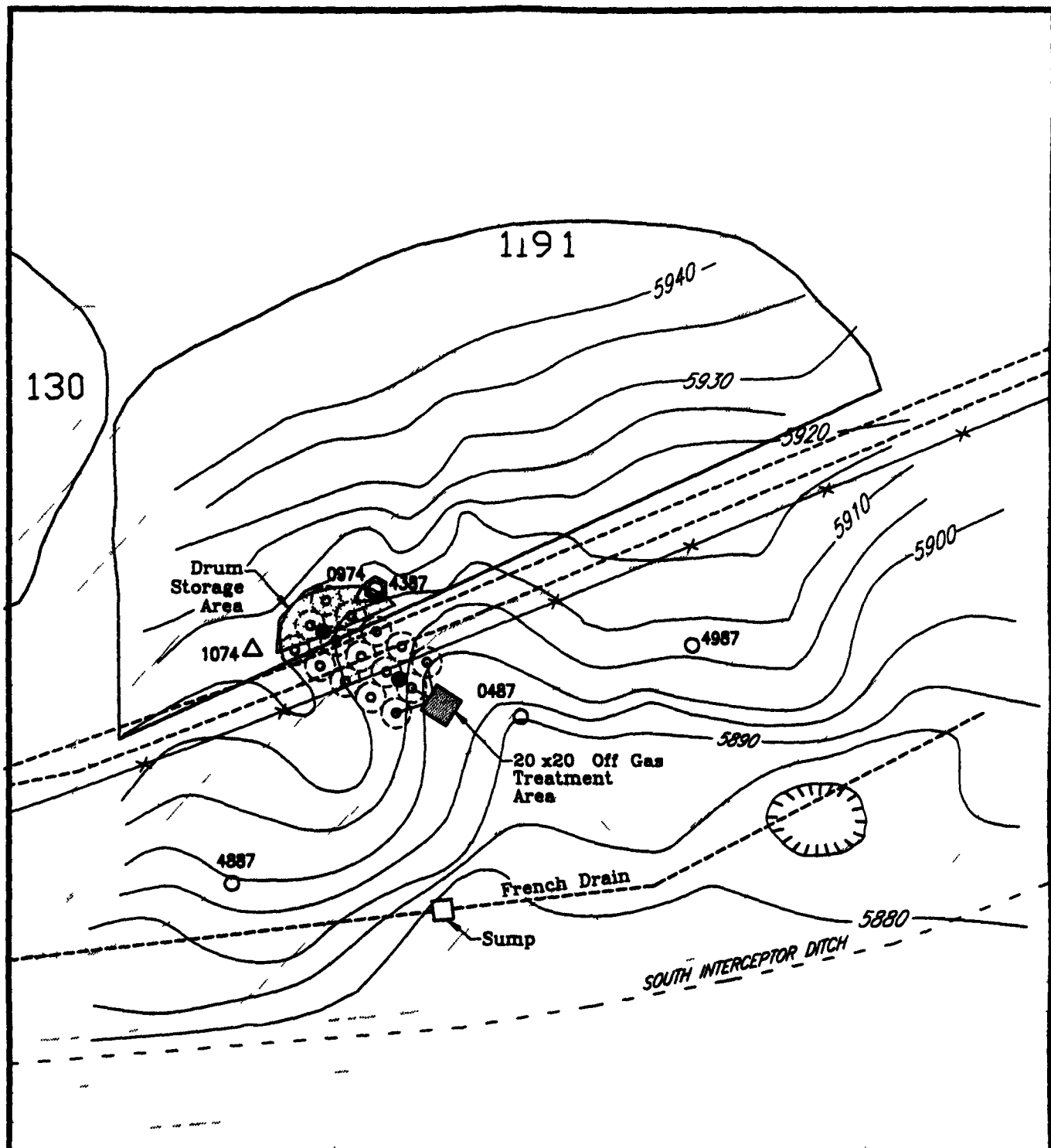
This alternative would remove contaminated groundwater by pumping it to the surface and

treating it in the Building 891 water treatment system. It would also remediate residual sources of contamination in the subsurface soils by SVE. The alternative targets all contaminants dissolved in the groundwater as well as sources of residual organic contamination such as DNAPLs. A detailed soil gas survey consisting of approximately 100 soil gas probes would be conducted to determine more precise locations of residual contaminants. The survey would take approximately six months. The results of this survey would be used to determine areas requiring treatment. For the purposes of detailed analysis, however, only the previously identified source area within IHSS 119.1 will be considered.

The treatment zone would first be dewatered by pumping from the existing groundwater extraction well and two additional extraction wells. These pumps would be operated intermittently to keep the source area dewatered. All groundwater collected would be piped to the french drain sump for transfer to the Building 891 water treatment system. Initial dewatering is expected to take 60 to 80 days, with intermittent operation continuing afterwards to keep the treatment zone dewatered throughout the entire remedial action. Approximately 80,000 gallons of groundwater is expected to be recovered, assuming that this alternative is implemented during a low water table elevation period.

Once dewatered, SVE would be applied at the source area to volatilize and remove any residual sources of organic contamination. Approximately 15 vapor extraction wells would be drilled in the source area and connected to a vacuum pump, thus inducing a vapor flow in the subsurface toward the extraction wells. The increased vapor flow would volatilize aqueous adsorbed and free phase contaminants and remove vapor phase contaminants with the evacuated air. A plan view of the SVE system layout is illustrated in Figure 4-1.

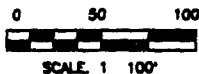
The size and number of vapor extraction wells, pump sizes, flow rates, and vacuum requirements were determined based on the results of the ongoing SVE treatability study at OU 2 in conjunction with the model HyperVentilate. Preliminary results from the OU 2 study indicate a 45 foot radius of influence was achieved in the silty sand matrix of OU 2 soils. This radius of influence was estimated by extrapolating observations of a six inch H₂O vacuum 20 feet from



OU1-SVE1.DWG

EXPLANATION

- | | | |
|--|--|---|
| | INDIVIDUAL HAZARDOUS SUBSTANCE SITE | |
| | ALLUVIAL WELL | |
| | PRE-1986 WELL | BEDROCK TOPOGRAPHY
C.I. = 5 |
| | BOREHOLE | TOPOGRAPHIC CONTOUR
INTERVAL = 10 FEET |
| | SVE WELL
WITH 10' RADIUS OF INFLUENCE | |
| | GROUND WATER EXTRACTION WELL | |



U S DEPARTMENT OF ENERGY
Rocky Flats Environmental Technology Site
Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

Plan View for
Alternative #4

Figure 4-1

the extraction well. The extraction well recovers two standard cubic feet per minute for each foot of well screening using a vacuum pump which produces 140 inches H₂O of vacuum at its inlet. The permeabilities observed at OU 2 have ranged from 0.1 darcies to 1 darcy which are an order of magnitude above those observed at OU 1. Initial contaminants concentrations at OU 2 were similar to those in the OU 1 source area however physical features of the OU 2 site including an excavated ditch near the treatability study zone have an undetermined influence on system parameters. For this reason estimates for parameters at OU 1 were made using technical literature to supplement the results from OU 2. Based on these two sources it is assumed that four inch diameter extraction wells operated at 10 scfm and 120 inches H₂O of vacuum would have a radius of influence of approximately 10 feet at OU 1.

Off gas from the SVE system would be treated with either GAC or catalytic oxidation prior to discharge. For costing purposes, GAC usage was estimated based on expected COC concentrations in extracted soil gas. Based on soil sample data from the Phase III RFI/RI Henry's Law was used to estimate the partitioning of COCs in the soil vapor. Calculations suggest that the equilibrium concentrations of each VOC would be on the order of 1 mg/l. However, because the quantities of COCs at OU 1 are assumed to be small equilibrium concentrations are not likely to be reached in extracted soil vapor. Therefore concentrations one order of magnitude below the Henry's Law calculated equilibrium values were used to determine the usage rate for GAC in the off gas treatment system. These concentrations are considered to be conservative estimates of soil vapor COC concentrations. Based on these assumptions the SVE off gas treatment system would require approximately 3,000 pounds of fresh GAC every three months.

Because of the low adsorption efficiency of 1,1-DCE on GAC the proposed SVE system would require two skid-mounted GAC vessels in series each containing 1,500 pounds of activated carbon. The activated carbon in the vessels would be replaced approximately every three months the spent carbon could be treated at an off site regeneration facility.

Off gas from the vacuum pump would be monitored to determine the effectiveness of the SVE.

system Intermittent operation could be employed to increase the recovery of residual sources and decrease operating costs

The remediation time frame for this alternative would be two years including six months for soil gas surveying four years for soil vapor extraction and dewatering operations and six months for mobilization/demobilization The french drain would be decommissioned upon completion of remedial activities

The evaluation of the two threshold and five balancing criteria for Alternative 4 Groundwater Pumping and Soil Vapor Extraction is summarized as follows

Overall Protection of Human Health and the Environment

This alternative would be protective of human health and the environment by removing the contaminants from both groundwater and subsurface soils Exposure potential would be reduced by reducing the contaminant concentrations and removing the source The french drain would continue to capture contaminated groundwater and prevent downgradient migration of COCs until remediation activities are completed

This alternative would be protective of the environment both downgradient of and within OU 1 because in addition to utilizing the existing french drain to intercept contaminated groundwater migrating away from OU-1 the source at IHSS 119 1 would be treated using SVE

Key ARARs would be met under this alternative In particular MCLs would continue to be achieved for groundwater COCs at Woman Creek providing long term effectiveness In addition this alternative would provide a large degree of permanence because the source area at IHSS 119 1 would be treated However, there is also some degree of uncertainty as to the level of cleanup that could be achieved for DNAPLs with SVE

Because this alternative would remediate the source at IHSS 119 1 a risk level of 3.22×10^{-7}

would be achieved at Woman Creek. Therefore, the magnitude of residual risk that would result from the implementation of this alternative falls well below the acceptable risk range of 10^{-4} to 10^{-6} . Additionally, risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6} .

This alternative would be completed in five years. During implementation, there would be no unacceptable short-term risks to the public. There may be potential risks to on-site workers from exposure to COCs in groundwater or soil vapor, and safety hazards associated with drilling and construction activities. However, risks would be minimized through standard health and safety practices.

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling projected that there will be no exceedance of MCLs at Woman Creek through implementation of this alternative. Groundwater modeling results demonstrate that the highest concentration of PCE is 5.94×10^{-4} for this alternative. The peak concentration occurs over a short duration and is below the MCL. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

Compliance with the RCRA program, involving releases of hazardous constituents from SWMUs under Subpart F, is a relevant and appropriate requirement. Since remediation is focused mostly on the source of contamination, the CAMU is a relevant and appropriate requirement for OU 1. Residual contamination may be left in place in areas of the unit; however, the overall quantity of constituents will be low. Compliance with the closure requirements as a disposal unit can be

achieved with Alternative 4 under the CAMU approach

Using a soil vapor extraction treatment system will create a temporary RCRA unit as the system treats hazardous waste constituents. Therefore the temporary unit requirements of Subpart S (6 CCR 1007.3 Subsection 264.553) are applicable to this treatment unit. In addition, any pre filters, HEPA filters, and activated carbon used to remove volatile organics in the off gas treatment require compliance with the following RCRA provisions:

- identification of hazardous waste (Part 261)
- air emission standards for process vents (Subsections 264.1032 and 264.1033)
- air emission standards for equipment leaks (Subsections 264.1056 and 1057)
- land disposal restrictions (Part 268)

It is anticipated that the applicable requirements of RCRA can be complied with in operating and decommissioning the SVE treatment unit and residuals.

The Colorado solid waste regulations are not appropriate to the CAMU unit created but are an ARAR for disposal of any residual materials that are not hazardous waste. If solid waste disposal is necessary with the alternative, it will be in accordance with 6 CCR 1007.2 and the solid waste disposal regulations.

Installation of additional extraction wells would be in accordance with the Colorado Water Well and Pump Installation Regulations (2 CCR 402.2). Compliance with this action-specific ARAR would be achieved.

The State's air pollution control Regulation 7, for the control of VOC emissions is an ARAR and will be achieved with this alternative. It is anticipated the level of emissions will be below the two ton/yr (two lbs/hr) threshold for use of reasonably available control technology (RACT).

Location Specific ARARs

Laws and regulations specific to wetlands and species which inhabit wetlands will be complied with if this alternative is implemented. There would be a short term impact to wetlands from decommissioning the french drain but the long term impact may be an increase in wetland areas. The State Division of Wildlife would be consulted prior to disturbance of wetland habitat and to implement adequate mitigation measures to protect species of special concern.

Long Term Effectiveness and Permanence

Under this alternative the source area at IHSS 119 1 would be remediated and the existing french drain extraction well, and Building 891 water treatment system would continue to extract and treat contaminated groundwater migrating from IHSS 119 1. Therefore the residual risk would be reduced as compared to the no action and institutional controls alternatives.

There is some uncertainty that SVE will effectively remediate the residual COCs at IHSS 119 1 due to the low permeability of the soils and the general lack of documented experience in effective DNAPL treatment at any site. If residual COCs are not effectively removed during remedial activities, they may continue to act as a source of groundwater contamination.

Following treatment of the source, contaminated groundwater within OU 1 would continue to migrate away from IHSS 119 1. Modeling indicates that under this alternative groundwater would continue to meet MCLs for the COCs at Woman Creek. A five year review would be conducted, however to determine the effectiveness of this alternative.

This alternative would provide long-term protection for potential human receptors and minimize the human health risk associated with contaminated groundwater by continuing to achieve MCLs. However although the MCLs will continue to be achieved with this alternative by remediating the soils and groundwater, this alternative may not be completely effective at removing DNAPLs if they are present at OU 1.

Reduction of Toxicity, Mobility, or Volume Through Treatment

This alternative satisfies the NCP preference for treatment as a principal element of the alternative. This alternative would effectively reduce the mobility and volume of contaminants in OU 1 by removing any secondary source of contaminants from the subsurface. Groundwater extraction would reduce the volume of COCs in groundwater, and soil vapor extraction would remove COCs from the unsaturated zone. Removing the secondary sources of contaminants in conjunction with the continued operation of the french drain and extraction well will also reduce their mobility by preventing potential additional migration.

Extracted groundwater would be treated in the Building 891 water treatment system. This is a destructive treatment process and thus would result in decreased toxicity. GAC from the SVE off gas treatment system would be regenerated off site resulting in additional reduction in TMV.

Contaminated materials generated as a result of this alternative include GAC from the off gas treatment system and spent ion exchange resins from the Building 891 water treatment system. GAC could be shipped off site to be regenerated, and ion exchange resins would be regenerated on site. There are no significant risks associated with handling and shipping either the spent activated carbon or ion exchange resins.

Short Term Effectiveness

Potential short term impacts on the environment associated with this alternative include a minor amount of disturbance to the soil and displacement or loss of vegetation during construction activities such as building and drilling. Additional short term risks to the public are minimal for this alternative.

Potential risks to workers during remediation activities include potential exposures to COCs in extracted groundwater or soil vapor. There are also safety hazards associated with drilling and other construction activities. Risks to workers would be minimized through standard

construction and process equipment operation safety practices

Implementability

This alternative would be readily implementable. Soil vapor extraction is a commonly used technology that does not require any unique or unusual equipment. The implementability of this alternative would not be limited by the availability of services and materials, nor would there be any significant technical or administrative difficulties associated with this alternative.

Implementing this alternative would not limit the ability to perform future remedial actions, if any are determined to be necessary. Groundwater monitoring programs continued under this alternative would continue to track any movement of COCs.

Vapor extraction wells can be installed using standard drilling techniques and standard construction materials that are readily available. Operation of the SVE system would not require highly specialized personnel or training. Spent GAC from the off-gas treatment system could be sent off site for regeneration. Spent ion exchange resin from the Building 891 water treatment system would be sent off site for appropriate disposal. Administrative requirements for this alternative would include obtaining an air emissions permit for the SVE off gas.

Cost

Costs for this alternative include the costs for the following items:

- soil gas survey (approximately 100 probes)
- two groundwater extraction wells (six inch diameter 20-foot depth)
- 15 vapor extraction wells (four inch diameter 20-foot depth)
- two vapor extraction systems with blowers, filters, and other appurtenances
- activated carbon adsorption system (two vessels containing 1,500 pounds each)
- associated piping, pumps, and instrumentation
- UV/peroxide/ion exchange water treatment system operation
- groundwater monitoring for 30 years

The capital cost for Alternative 4 is \$929,300 O&M and post closure activities for Alternative 4 include the operation of the SVE system for one year the operation of the existing french drain and Building 891 water treatment system until completion of remediation and groundwater monitoring for 30 years The present worth of the O&M costs for this alternative is \$5 358 700 the post closure cost is \$1 853 800 The total cost of this alternative is \$8 141 800 A detailed cost estimate for capital and O&M costs for this alternative is included in Appendix E

4 2 6 Alternative 5. Groundwater Pumping and Soil Vapor Extraction With Thermal Enhancement

This alternative would remove contaminated groundwater by pumping it to the surface and treating it in the Building 891 water treatment system It would also remediate residual sources of contamination in the subsurface soils by SVE with thermal enhancement The alternative targets all contaminants dissolved in the groundwater as well as sources of residual organic contamination such as DNAPLs Initially a detailed soil gas survey, consisting of approximately 100 soil gas probes would be conducted to determine more precise locations of residual contaminants The survey would take approximately six months The results of this survey would be used to determine areas requiring treatment For the purposes of this detailed analysis only the previously identified source area within IHSS 119 1 will be considered for treatment

As with Alternative 4 the treatment zone would first be dewatered by pumping from the existing groundwater extraction well and two additional extraction wells All groundwater collected would be piped to the french drain sump for transfer to the Building 891 water treatment system Initial dewatering is expected to take 60 to 80 days with intermittent operation continuing afterwards to keep the treatment zone dewatered throughout the entire remedial action Approximately 80,000 gallons of groundwater are expected to be recovered

Once dewatered, thermally enhanced SVE would be applied to the treatment zone to volatilize and remove any residual sources of organic contamination Just as with Alternative 4 approximately 15 vapor extraction wells would be drilled in the source area The SVE system

utilized in this alternative would be the same as was described for Alternative 4 based on the results of the OU 2 treatability study. To reduce the remediation time frame, subsurface temperatures would be raised to approximately 260 °C using radio frequency antennae. All of the vapor extraction wells would be constructed to accommodate the RF antennae. The increased vaporization caused by the elevated temperatures would result in a reduction in remediation time as compared to unenhanced SVE.

Off gas from the SVE system would be treated with GAC or catalytic oxidation prior to discharge as described for Alternative 4. The proposed SVE system would require two skid mounted GAC vessels in series, each containing 1,500 pounds of activated carbon. The activated carbon in the vessel would be replaced approximately every three months; the spent carbon could be treated at an off site regeneration facility. Carbon replacement may be required more frequently than every three months, depending on the efficiency of the RF heating process.

Off gas from the vacuum pump would be monitored to determine the effectiveness of the enhanced SVE system. Intermittent operation could be employed to increase the recovery of residual sources and decrease operating costs. The remediation time frame for this alternative is three years, including six months for soil gas surveying, two years for soil vapor extraction and dewatering operations, and six months for mobilization/demobilization. The french drain would be decommissioned upon completion of remedial activities.

The evaluation of the two threshold and five balancing criteria for Alternative 5, Groundwater Pumping and Soil Vapor Extraction With Thermal Enhancement, are summarized as follows:

Overall Protection of Human Health and the Environment

This alternative would protect human health and the environment by removing the contaminants from both groundwater and subsurface soils. Exposure potential would be reduced by reducing the contaminant concentrations and removing the source. The existing french drain and

extraction well would continue to capture contaminated groundwater and prevent downgradient migration of COCs

This alternative would be protective of the environment both downgradient of and within OU 1 because in addition to utilizing the existing french drain to intercept contaminated groundwater migrating away from OU 1 the source at IHSS 119 1 would be remediated

RF heating may have some negative impacts on the soils at OU 1 due to the high temperatures that are reached during operation While the elevated temperatures may increase the effectiveness of treatment for the COCs, they may be harmful to subsurface biota

Key ARARs would be met for this alternative In particular modeling shows that MCLs would continue to be achieved for groundwater COCs at Woman Creek providing long term effectiveness In addition this alternative would provide a large degree of permanence because the source area at IHSS 119 1 would be treated However there is also some degree of uncertainty as to the level of cleanup that could be achieved for source area DNAPLs with the enhanced SVE technology

Because this alternative would remediate the source at IHSS 119 1 modeling shows that a risk level of 3.22×10^{-7} would be achieved at Woman Creek Therefore the magnitude of residual risk that would result from the implementation of this alternative falls well below the acceptable risk range of 10^{-4} to 10^{-6} Additionally, risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6}

Alternative 5 would be completed within three years During implementation there would be no unacceptable short term risks to the public There may be potential risks to on-site workers from exposure to COCs in groundwater or soil vapor and safety hazards associated with drilling construction activities, and operating the RF heating elements However, risks would be minimized through standard health and safety practices

Compliance With Applicable or Relevant and Appropriate Requirements

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling results indicate that there will be no exceedance of MCLs at Woman Creek through implementation of this alternative. Groundwater modeling results demonstrate that the highest concentration of PCE is 5.94×10^{-4} mg/l for this alternative. The peak concentration occurs over a short duration and is below the MCL. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization. Alternative 5 and Alternative 4 are the same in terms of compliance with ARARs with few exceptions. The exceptions are noted in the following discussions of action specific and location specific ARARs.

Action Specific ARARs

The action specific ARARs associated with Alternative 5 are the same as presented in Alternative 4. Compliance with RCRA requirements for a temporary unit, CAMU, identification of hazardous waste, storage of hazardous waste, disposal of hazardous waste, and organic air emissions and leaks from the treatment unit will be achieved. The possible difference in the two alternatives may be the amount of organic constituents left in the OU 1 general area. It is possible that the thermal enhancement will leave slightly less contaminant pockets and soil vapor around extraction well locations, assuming air drying techniques are effective and the vapor generation rate is controlled. Closure of the CAMU would involve less contaminants left in place.

Other action specific ARARs are anticipated to be complied with in the same manner as was discussed under Alternative 4.

Location Specific ARARs

The areas of heaviest organic contamination are at IHSS 119 1 and 119 2. Assuming additional extraction wells are placed around these areas away from the french drain and Pond C 1, destruction of riparian vegetation and fauna during thermal enhancement will be minimal. Therefore, the habitat and fauna such as Preble's meadow-jumping mouse will remain intact. This species is of special concern, not threatened or endangered under State or Federal law. It is anticipated that compliance with DOE wetland protection regulations and the State's law concerning non game species will be achieved with implementation of this alternative. Should it be necessary, riparian habitat would be replaced if it is destroyed by the thermal technology.

Impacts from decommissioning the french drain would be the same as has been discussed for all previous alternatives. The wetland protection regulations and non game species laws would be complied with if this alternative is selected.

Long Term Effectiveness and Permanence

Similar to Alternative 4, this alternative would remediate the source area at IHSS 119 1 and the existing french drain extraction well, and Building 891 water treatment system would continue to extract and treat contaminated groundwater migrating from IHSS 119 1. Therefore, the residual risk would be reduced as compared to the no action and institutional control alternatives.

This alternative may provide more effective and permanent treatment as compared to Alternative 4, because enhancing SVE with RF heating may more effectively remove residual COCs trapped within the less permeable soils at OU 1. However, it is uncertain whether enhanced SVE or other in situ technologies can effectively remediate DNAPLs at 119 1 due to the general lack of documented success in treating DNAPLs at any site.

This alternative would provide long-term protection for potential human receptors and minimize the human health risk associated with contaminated groundwater by continuing to achieve

groundwater MCLs at Woman Creek. However, although this alternative may continue to effectively achieve the MCLs by remediating the soils and groundwater, it may not be completely effective at removing DNAPLs if they are present at OU 1. Following treatment of the source, contaminated groundwater within OU 1 would continue to migrate away from IHSS 119. Modeling indicates that under this alternative, groundwater would continue to meet MCLs for COCs at Woman Creek. A five year review would be conducted to determine the continued effectiveness of this alternative.

Contaminated materials generated as a result of this alternative include spent carbon from the SVE off gas treatment system and spent ion exchange resins from the Building 891 water treatment system. There are no significant risks associated with handling and shipping either the spent carbon or the ion exchange resins.

Reduction of Toxicity, Mobility, or Volume Through Treatment

This alternative satisfies the NCP preference for treatment as a principal element of the alternative. This alternative would effectively reduce the mobility and volume of contaminants in OU 1 by removing any secondary source of contaminants from the subsurface. Groundwater extraction would reduce the volume of COCs in groundwater, and soil vapor extraction would remove COCs from the unsaturated zone. Removing the secondary sources of contaminants would also reduce their mobility by preventing potential additional migration.

Extracted groundwater would be treated in the Building 891 water treatment system. This is a destructive treatment process and thus would result in decreased toxicity. GAC from the SVE off gas treatment system could be regenerated off site, resulting in additional reduction in toxicity.

Contaminated materials generated as a result of this alternative include activated carbon from the off gas treatment system and spent ion exchange resins from the Building 891 water treatment system. Both wastes would be shipped off site for treatment and eventual disposal.

There are no significant risks associated with handling and shipping either the spent activated carbon or ion exchange resins

Short Term Effectiveness

Potential short term impacts on the environment associated with this alternative include a minor amount of disturbance to the soil and displacement or loss of vegetation during construction activities. In situ soil heating may have additional impacts on subsurface biota due to the high temperatures that are reached during operation. Potential short term impacts to the public are minimal under this alternative.

Potential risks to workers during remediation activities include potential exposures to COCs in extracted groundwater or soil vapor. There are also safety hazards associated with drilling and other construction activities as well as with the operation of the RF heating devices. Risks to workers would be minimized through standard construction and process equipment operation safety practices.

Implementability

This alternative would be readily implementable. Soil vapor extraction is a proven and commonly used technology that does not require any unique or unusual equipment. Although RF heating is a less common technology, it is readily available through specialized vendors. The implementability of this alternative would not be limited by the availability of services and materials nor would there be any significant technical or administrative difficulties associated with this alternative.

Implementing this alternative would not limit the ability to perform future remedial actions if any are determined to be necessary. Groundwater monitoring continued under this alternative would track any movement of COCs.

Vapor extraction wells can be installed using standard drilling techniques and standard construction materials that are readily available. Operation of the SVE system would not require highly specialized personnel or training; however, using the RF heating antennae would require a certain amount of special training or assistance from the vendor. The RF antennae can be easily installed in one or several of the vapor extraction wells and can be easily moved from one well to another, as required. RF heating does not produce any residuals.

Administrative requirements for this alternative would include obtaining an air emissions permit for the SVE off gas. Spent GAC from the off gas treatment system could be sent off site for regeneration. Spent ion exchange resin from the Building 891 water treatment system would continue to be regenerated on site.

Cost

Costs for this alternative include the costs for the following items:

- soil gas survey (approximately 100 probes)
- two groundwater extraction wells (six inch diameter, 20-foot depth)
- 15 vapor extraction wells (four-inch diameter, 20-foot depth)
- two vapor extraction systems, with blowers, filters, and other appurtenances
- activated carbon adsorption system (two vessels containing 1,500 pounds each)
- radio frequency heating unit
- associated piping, pumps, and instrumentation
- UV/peroxide/ion exchange water treatment system operation
- groundwater monitoring for 30 years

The capital cost for Alternative 5 is \$1,845,700. O&M and post-closure activities for Alternative 5 include operation of the enhanced SVE, operation of the existing french drain, and operation of the Building 891 water treatment system until the completion of remediation. Groundwater monitoring for 30 years is also included. The present worth of the O&M costs for this alternative is \$3,845,700; the post-closure cost is \$1,811,700. The total cost of this alternative is \$7,503,100. A detailed cost estimate for capital and O&M costs for this alternative is included in Appendix E.

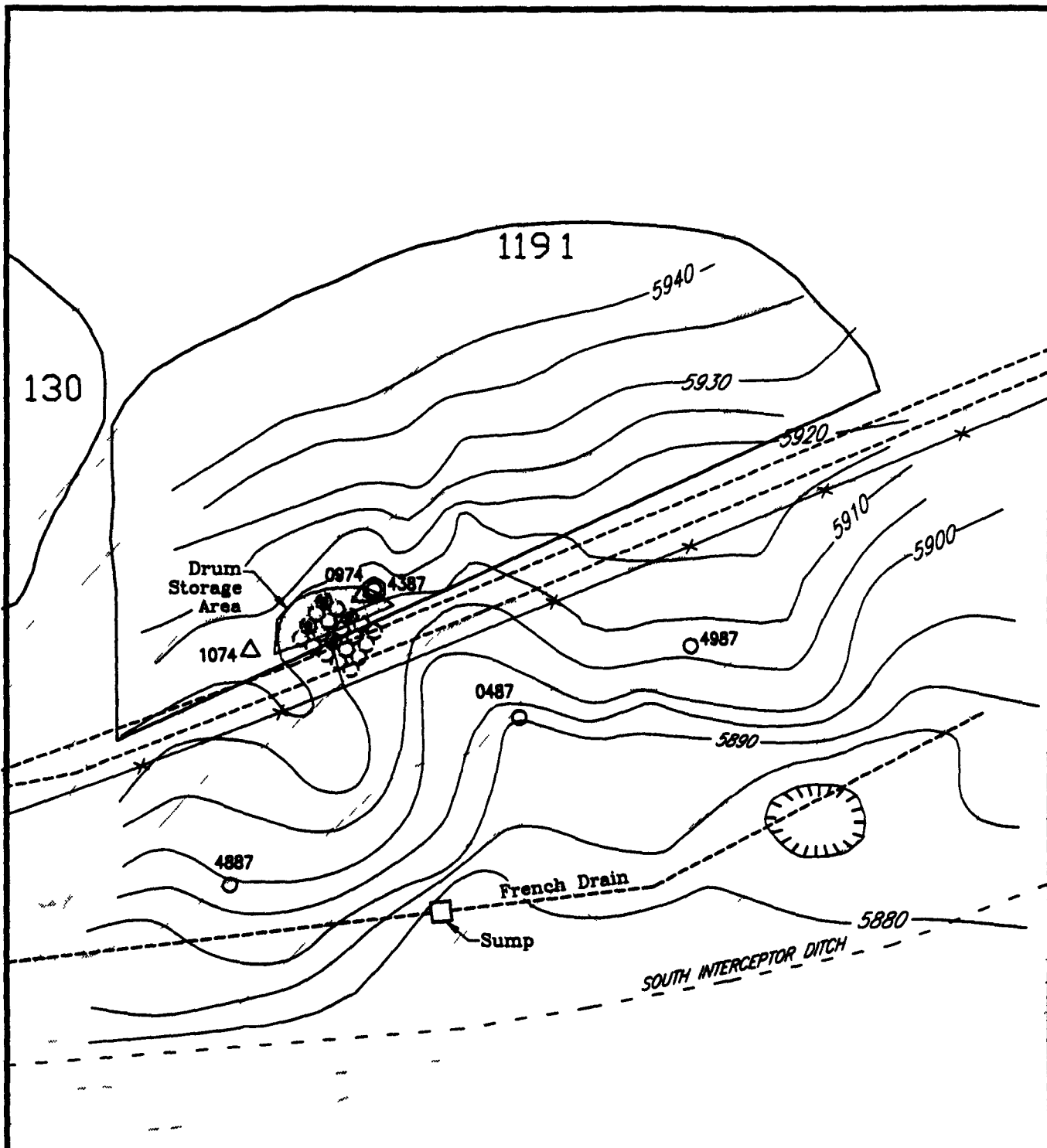
4 2 7 Alternative 6. Hot Air Injection With Mechanical Mixing

This alternative would remediate groundwater by pumping it to the surface and treating it in the Building 891 water treatment system. It would also remove residual sources of contamination from the soil with a combination of in situ mechanical mixing and soil vapor extraction with hot air injection. The alternative targets dissolved, adsorbed, vapor, and free phase DNAPL contamination. A detailed soil gas survey consisting of approximately 100 soil gas probes would be conducted initially to determine locations within the source area that require remediation.

The treatment zone would initially be dewatered by pumping the existing groundwater extraction well. This is expected to take approximately five days since the well is currently in operation and the water table in the vicinity is already depressed. The area near the existing well would then be treated using a portable system that combines mechanical mixing and thermal enhancement with conventional SVE techniques. This is an innovative, proprietary technology which increases the rate of soil vapor recovery through vigorous mixing and thermally enhanced volatilization. The rig-mounted equipment consists of a 10-foot diameter auger capable of mixing soil to a depth of 20 feet while injecting hot air. Soil vapors are recovered at the surface through a 12 foot diameter shroud. The physical mixing ensures exposure of all treated soils to high volume air flow and eliminates channeling and dead zones. The combination of mixing and enhanced volatilization reduces treatment time and increases contaminant removal effectiveness in heterogeneous or tightly consolidated media such as is present at OU 1. A plan view of Alternative 6 is shown in Figure 4-2.

As the system moves from one soil column to the next, groundwater extraction wells would be installed and removed from various locations in the treated unconsolidated material to ensure that the treatment zone is kept dewatered during treatment. Several wells would be installed permanently as groundwater monitoring locations.

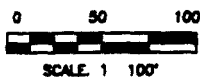
The remedial time frame for this treatment method is highly dependent on site conditions and



OU1-BRS DWG

EXPLANATION

- | | |
|--|-------------------------------------|
| | INDIVIDUAL HAZARDOUS SUBSTANCE SITE |
| | ALLUVIAL WELL |
| | PRE-1986 WELL |
| | BOREHOLE |
| | 10 DIAMETER BORE |
| | GROUND WATER EXTRACTION WELL |
- BEDROCK TOPOGRAPHY
C.I. = 5'
- TOPOGRAPHIC CONTOUR
INTERVAL = 10 FEET



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Plan View for
Alternative #6

Figure 4-2

the volume of soil requiring remediation The system would operate for approximately three months and treat approximately 4 500 cubic yards of soil The french drain would be decommissioned upon completion of remedial activities

The evaluation of the two threshold and five balancing criteria for Alternative 6 Hot Air Injection With Mechanical Mixing are summarized as follows

Overall Protection of Human Health and the Environment

This alternative would be protective of human health and the environment by removing the contaminants from both groundwater and subsurface soils at IHSS 119 1 Exposure potential would be reduced by decreasing the contaminant concentrations and removing the source The existing french drain and extraction well would continue to capture contaminated groundwater and prevent downgradient migration of COCs

This alternative would be protective of the environment both downgradient of and within OU 1 because in addition to utilizing the existing french drain to intercept contaminated groundwater migrating away from OU 1, the source at IHSS 119 1 would be remediated

Hot air injection may have some negative impacts on the soils at OU 1 due to the high temperatures that are reached during operation While the elevated temperatures may increase the effectiveness of treatment for the COCs they may be harmful to subsurface biota This technology may also have a higher potential to spread the contaminants away from the treatment zone

Alternative 6 would meet key ARARs In particular MCLs would continue to be achieved for groundwater COCs at Woman Creek, providing long term effectiveness In addition this alternative would provide permanence to the degree that it could effectively treat the source area at IHSS 119 1 However, there is some uncertainty as to the level of cleanup that could be achieved for DNAPLs with this technology

Because this alternative would remediate the source at IHSS 119 1 modeling shows that a risk level of 3.22×10^{-7} would be achieved at Woman Creek respectively Therefore the magnitude of residual risk that would result from the implementation of this alternative falls well below the acceptable risk range of 10^{-4} to 10^{-6} Risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6}

This alternative would be completed in approximately one year During implementation there would be no unacceptable short term risks to the public There may be potential risks to on site workers from exposure to COCs in groundwater or soil vapor and safety hazards associated with construction activities the hot air injection and operation of the mechanical mixing tool However risks would be minimized through standard health and safety practices

Compliance With Applicable or Relevant and Appropriate Requirements

The designation of ARARs for this alternative is the same as has been presented in Alternative 4 and 5 Similarly, the compliance with these ARARs is the same that is Alternative 6 will comply with chemical-specific, action specific and location specific ARARs The following discussions mention the variations of technology and implementation where it is important to compliance with ARARs

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs Modeling results also indicate that there will be no exceedance of MCLs at Woman Creek through implementation of this alternative Groundwater modeling results demonstrate that the highest concentration of PCE period is 5.94×10^{-4} mg/l for this alternative The peak concentration occurs over a short duration and is below the MCL Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics

Action Specific ARARs

Compliance with the RCRA requirements for

- identification of hazardous waste treatment residuals (off gas treatment system)
- storage of hazardous waste treatment residuals
- disposal of hazardous waste treatment residuals
- organic air emission leaks from the treatment system
- a temporary unit
- a CAMU can be achieved with this alternative

The CDPHE would need to designate OU 1 a CAMU and establish the closure and any groundwater monitoring requirements in compliance with the RCRA regulations (6 CCR 1007 3) Alternative 6 similar to Alternative 5 may enhance the amount of organics that can be extracted from the soil and soil vapor, assuming mechanical mixing and hot air injection work effectively. It is possible that these combined technologies will leave slightly less contaminants in the CAMU area at the time of closure than Alternative 4.

Other action specific ARARs would be complied with in a manner similar to Alternative 4 and 5.

Location Specific ARARs

It is assumed that mechanical mixing, hot air injection, and additional extraction wells will be placed within or near IHSS 119 1, away from the french drain and Pond C 1. Therefore, the habitat and fauna such as Preble's meadow jumping mouse will remain intact. This species is of special concern, not threatened or endangered, under State or Federal law. It is anticipated that compliance with DOE wetland protection regulations and the State's law concerning non game species will be achieved with implementation of this alternative. Should it be necessary, riparian habitat would be replaced if it is destroyed by the thermal technology.

Long Term Effectiveness and Permanence

This alternative would remediate the source area at IHSS 119 1 and the existing french drain extraction well and treatment system would continue to extract and treat contaminated groundwater migrating from IHSS 119 1. Therefore the residual risk would be reduced as compared to the no action and institutional control alternatives.

This alternative may provide more effective and permanent treatment as compared to the SVE alternatives. Alternatives 4 and 5, because mechanical mixing and hot air injection would more thoroughly remove residual COC concentrations from the soil. There is some uncertainty however, that the COCs at IHSS 119 1 can be effectively remediated due to the low permeability of the soils and the general lack of documented experience in effective DNAPL treatment at any site. If residual COCs are not effectively removed during remedial activities they may continue to act as a source of groundwater contamination.

Following treatment of the source contaminated groundwater within OU-1 would continue to migrate away from IHSS 119 1. Modeling indicates that under this alternative groundwater would continue to meet MCLs for the COCs at Woman Creek. A five year review would be conducted to determine the continued effectiveness of this alternative.

This alternative would provide long-term protection for potential human receptors and minimize the human health risk associated with contaminated groundwater by continuing to achieve MCLs at Woman Creek. However, although the MCLs can continue to be achieved with this alternative by remediating the soils and groundwater, this alternative may not be completely effective at removing DNAPLs if they are present at OU 1.

Contaminated materials generated as a result of this alternative include spent carbon from the remediation equipment off-gas treatment system and spent ion exchange resins from the Building 891 water treatment system. The spent GAC would be shipped off site for regeneration and the spent ion exchange resin could continue to be regenerated on site. There are no significant risks.

associated with handling and shipping either the spent carbon or the ion exchange resins

Reduction of Toxicity, Mobility, or Volume Through Treatment

Alternative 6 would satisfy the NCP preference for treatment as a principal element of the alternative. This alternative would effectively reduce the mobility and volume of contaminants in OU 1 by removing any secondary source of contaminants from the subsurface. Mechanical mixing and hot air injection would reduce the volume of COCs in subsurface soil in both the saturated and unsaturated zones. Removing the residual sources of contaminants will also reduce their mobility by preventing potential additional migration.

Extracted groundwater would be treated in the Building 891 water treatment system. This is a destructive treatment process and thus would result in decreased toxicity. GAC from the off gas treatment system would be regenerated, resulting in additional reductions in toxicity.

The hot air injection and mechanical mixing technology utilizes soil vapor extraction to control COCs volatilized by the mixing and increased temperatures. However, the technology may actually increase the mobility of contaminants by spreading them beyond the boundaries of the treatment zone.

Contaminated materials generated as a result of this alternative include activated carbon from remediation equipment off gas treatment system and spent ion exchange resins from the Building 891 water treatment system. There are no significant risks associated with handling and shipping either the spent activated carbon or ion exchange resins.

Short Term Effectiveness

This alternative would have significant short term impacts on the environment within the treatment zone at IHSS 119.1. Impacts include disturbance to the soil and displacement or loss of vegetation during remedial activities. The hot air injection and thorough mechanical mixing

would have severe impacts on subsurface biota due to the high temperatures that are reached during operation. Potential short term impacts to the public are minimal under this alternative.

Potential risks to workers during remediation activities include potential exposures to COCs in extracted groundwater or in soil vapor. There are also safety hazards associated with the operation of the mechanical mixing equipment. Mixing the soil would increase the risks associated with operating heavy equipment in OU 1 which is characterized by highly unstable soils. The risks to workers would be minimized through standard construction and process equipment operation safety practices.

Implementability

This alternative would be readily implementable. Although the technology that would be utilized in this alternative is not as common as those for the other remediation alternatives, equipment for hot air injection and mechanical mixing is readily available. The implementability of this alternative would not be limited by the availability of services and materials, nor would there be any significant administrative difficulties associated with this alternative.

There are several potential technical problems that may be encountered if this alternative is implemented. First, the technology may be difficult to implement at OU 1 due to the claystone material that is found in the subsurface. The formation is highly unstable which may present safety problems during remedial operations. In addition, as the remediation progresses the remediation zone would become completely mixed, saturated, and soft. Installing the necessary dewatering and monitoring wells into the treatment zone may not be possible if a drill rig can not be driven onto this material.

Administrative requirements for this alternative would include obtaining an air emissions permit for the off gas system. Spent ion exchange resin from the Building 891 water treatment system could continue to be regenerated on site.

Cost

Costs for this alternative include the costs for the following items

- soil gas survey (approximately 100 probes)
- five groundwater extraction wells (six inch diameter 20 foot depth)
- mechanical mixing unit
- associated piping pumps and instrumentation
- UV/peroxide/ion exchange water treatment system operation
- groundwater monitoring for 30 years

The capital cost for Alternative 6 is \$1 354 400 O&M and post-closure activities for Alternative 6 include the operation of the existing french drain and Building 891 water treatment system until completion of remedial activities and groundwater monitoring for 30 years The present worth O&M cost for this alternative is \$1 887 300 the post-closure cost is \$1 789 100 The total cost of this alternative is \$5 030 800 A detailed cost estimate for capital and O&M costs for this alternative is included in Appendix E

4 2 8 Alternative 7. Soil Excavation and Groundwater Removal With Sump Pumps

This alternative consists of the excavation of approximately 20 000 cubic yards of unsaturated and potentially saturated soil from IHSS 119 1 The amount of groundwater collected from dewatering the excavation would be approximately 80,000 gallons depending on the seasonal level of the water table Standard submersible pumps would be utilized to pump the groundwater to the OU 1 french drain sump pump which would then direct the water to the Building 891 water treatment system for final treatment and discharge

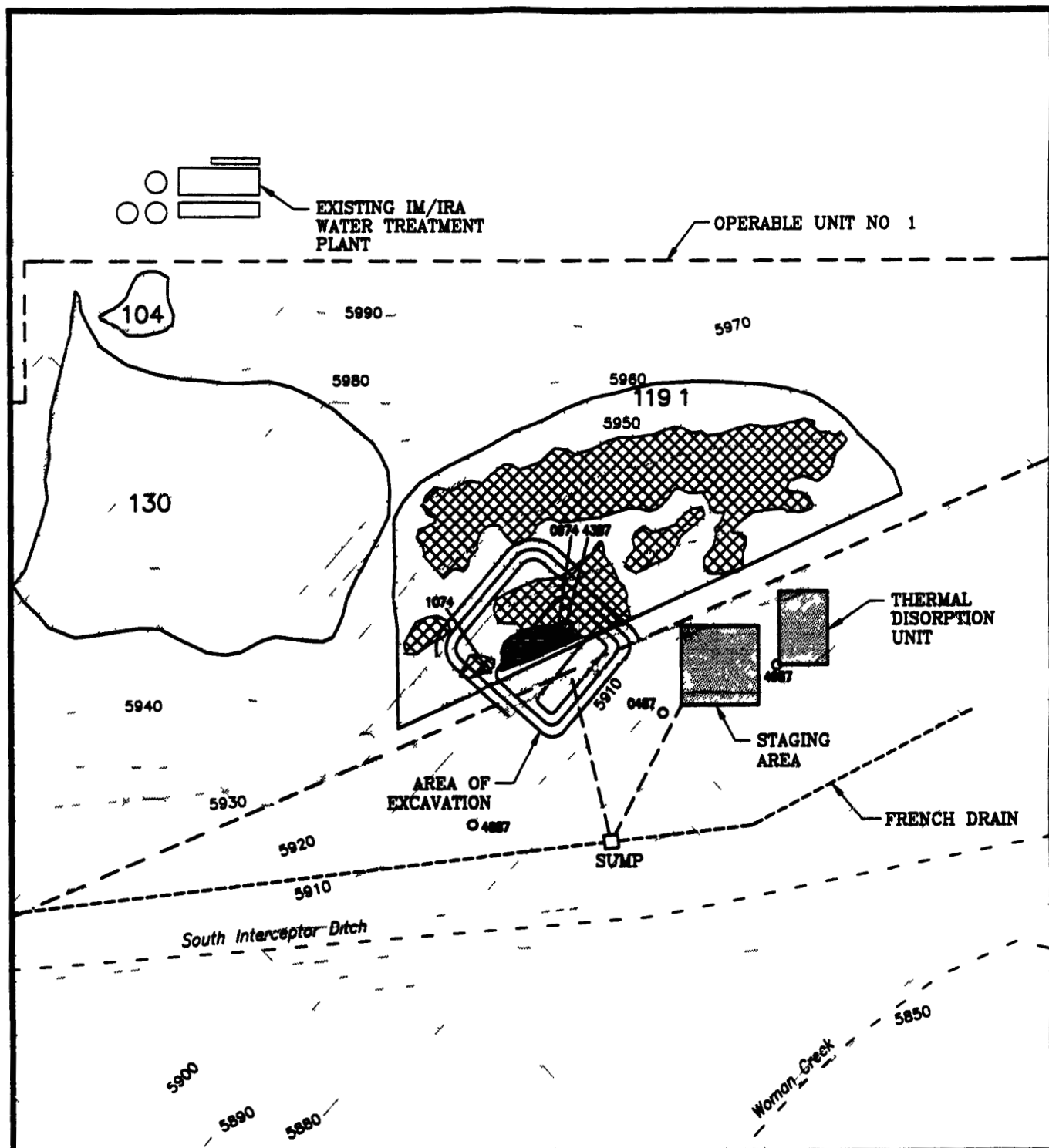
The top soil located within the excavation area would be scraped and stockpiled on site to be treated with top soils from OU 2 at a later time and the subsurface soils excavated and transported to a staging area constructed within 300 feet of the excavation The soil removed from the excavation could not be replaced because even treated soils could not be certain to be protective of groundwater MCLs The low organic carbon content of the soil favors partitioning

to the aqueous phase, therefore very small concentrations of VOCs in soil could result in aqueous phase concentrations greater than the MCLs. Based on a fraction of organic carbon (f_{oc}) of 0.002 for the soils at OU 1 (DOE 1994a) the concentrations of COCs in soil that would be required to be protective of groundwater MCLs are below detection limits and therefore too low to verify. Therefore all excavated soils would need to be disposed off site. Prior to being disposed in an off site hazardous waste landfill however the soil would require treatment to meet the Land Disposal Restriction treatment standards for each constituent. The treatment standards for 1,1-DCE, CCL_4 , PCE and 1,1,1-TCA are 33 mg/kg (proposed), 5.6 mg/kg, 5.6 mg/kg and 5.6 mg/kg respectively.

The excavated soil stockpiled in the staging area would be dewatered and then treated by a skid mounted thermal desorption unit. The duration of this alternative has been estimated to be nine months although it is highly dependent upon the capacity of the thermal desorption unit. Radiological monitoring would be conducted for the duration of the alternative. A plan view of Alternative 7 is illustrated in Figure 4-3.

Thermal desorption is a commercially available proven technology for the removal of volatile organics from soil. A conveyor feeds the soil to the thermal desorption unit which raises the temperature of the soil to 343 °C to volatilize the VOCs. The thermal desorption unit would be equipped with a baghouse to remove any particulates from the exhaust stream and an off gas treatment unit such as a catalytic oxidizer, for the destruction of any remaining VOCs and carbon monoxide. The treated air could then be emitted to the atmosphere. A thermal desorption unit could be obtained locally and mobilized on site in one day.

Following treatment, the soil would be packaged and shipped to a licensed facility located within 100 miles of the site. It should be noted that there is currently a proposed rule which states that if the soil were to be treated to meet the proposed Universal Treatment Standard for each contaminant of concern it may be possible to put the soil back into the excavation following treatment. However given the geology of the site there is a possibility that the soils treated to either of these standards could contaminate the groundwater above MCLs if it were placed



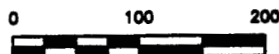
EXPLANATION

104 INDIVIDUAL HAZARDOUS SUBSTANCE SITE (IHSS) AND IHSS DESIGNATION DASHED WHERE DISTURBED DURING CONSTRUCTION OF FRENCH DRAIN

ACTUAL SCRAP METAL AND DRUM STORAGE AREAS IN IHSS 119 BASED ON AERIAL PHOTOGRAPHS

ACTUAL DRUM STORAGE AREA IN IHSS 119.1 BASED ON AERIAL PHOTOGRAPHS

- 8301888 ALLUVIAL WELL
- △ 0271 PRE-1986 WELL
- BH1587 BOREHOLE



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Plan View for
Alternative #7

Figure 4-3

OUT-EXC1 DWG

back into the excavation. It has therefore been assumed that the soil would require shipment to an off site disposal facility.

Extensive groundwater monitoring would not be required for this alternative because the source of contamination would be removed. Groundwater monitoring would be continued to evaluate the effectiveness of the removal action. Short term air monitoring would be required during excavation activities. The french drain would be decommissioned upon completion of remedial activities.

The evaluation of the two threshold and five balancing criteria for Alternative 7 Soil Excavation and Groundwater Removal With Sump Pumps are summarized as follows:

Overall Protection of Human Health and the Environment

This alternative would be protective of human health and the environment by excavating and treating subsurface soils, and pumping and treating groundwater at IHSS 119.1. Exposure potential would be reduced by decreasing the contaminant concentrations and removing the source. The existing french drain and extraction well would continue to capture contaminated groundwater and prevent downgradient migration of COCs.

This alternative would be protective of the environment both downgradient of and within OU 1 because in addition to utilizing the existing french drain to intercept contaminated groundwater migrating away from OU 1, the source at IHSS 119.1 would be remediated.

This alternative would meet key ARARs. In particular, MCLs would continue to be achieved for groundwater COCs at Woman Creek, providing long-term effectiveness. This alternative would also provide the greatest degree of permanence of all the treatment alternatives because excavating the source area would be the most effective way to remove DNAPLs.

Alternative 7 would have a significant impact on the environment due to the large excavation.

and material transportation requirements. Excavating the entire source area would negatively impact the flora and subsurface biota. However, there are no significant long term effects anticipated from this alternative.

Because this alternative would remediate the source at IHSS 119.1, a risk level of 3.22×10^{-7} would be achieved at Woman Creek, respectively. Therefore, the magnitude of residual risk that would result from the implementation of this alternative falls well below the acceptable risk range of 10^{-4} to 10^{-6} . Additionally, risk from surface soil contaminants would remain within the acceptable risk range of 10^{-4} to 10^{-6} .

This alternative would be completed within one year. During implementation, there would be the potential for risk to the public due to potentially contaminated air-borne dust generated during excavating activities and the transportation of large quantities of excavated soils off site. There may also be potential risks to workers from exposure to COCs in groundwater or air-borne dust. Workers would also have potential safety hazards associated with operating the earth-moving equipment and the thermal desorption unit. However, risks to workers would be minimized through standard health and safety practices.

Compliance With Applicable or Relevant and Appropriate Requirements

The ARARs associated with this alternative are very similar to those presented and discussed for Alternatives 4, 5, and 6. Alternative 7 can comply with chemical-specific, location-specific, and action-specific ARARs. The major difference of this alternative from 4, 5, and 6 is the excavation of soil to groundwater, the subsequent aboveground treatment of soils, and disposal of soils in a RCRA facility. The management of soils will require compliance with the hazardous waste land disposal restrictions. The following paragraphs summarize the compliance with ARARs status for this alternative.

Chemical Specific ARARs

The results of groundwater modeling and groundwater monitoring indicate that groundwater at Woman Creek currently meets MCLs. Modeling results also indicate that there will be no exceedance of MCLs at Woman Creek through implementation of this alternative. Groundwater modeling results demonstrate that the highest concentration of PCE is 5.94×10^4 mg/l for this alternative. The peak concentration occurs over a short duration and is below the MCL. Assumptions of the model are discussed in Appendix B and have included factors for natural degradation but not volatilization of organics.

Action Specific ARARs

Compliance with the following RCRA requirements will be accomplished

- identification of hazardous waste treatment residuals
- treatment of contaminated soil in a temporary unit
- disposal of hazardous waste treatment residuals and soil
- monitoring of organic air emission equipment leaks and process vents
- a CAMU for OU 1

Although the source of the majority of contamination would be removed with this alternative because some unknown amount contamination would be left in place, CDPHE would need to designate OU 1 as a CAMU. It is anticipated that once groundwater is removed and treated in the Building 891 water treatment system and soil is treated and disposed of off site, the requirements for groundwater monitoring and post-closure of the CAMU could be minimal.

It is possible that EPA's proposed definition of soil contaminated with hazardous waste could be promulgated prior to the final CAD/ROD. It is anticipated that this alternative would meet the proposed changes to the definition of hazardous waste.

Other action specific ARARs such as Regulation 7 Control of VOC Emissions of the Colorado air pollution regulations, will be complied with in operating the desorption unit. Levels of

emissions are anticipated to be below two tons/year of VOCs

Location Specific ARARs

This alternative would involve placement of a PVC pipe from the excavated area of contamination to the existing french drain sump. Although the area involved in construction activity would be small, there would be a short term impact to riparian/wetland areas around the french drain. Any riparian vegetation which is destroyed would be replaced according to DOE regulations on wetland protection. This assumes it is determined that there are no other alternatives which could achieve the same or similar result.

This alternative could result in a negative short term impact to a State species of special concern. Mitigation measures would need to be discussed within the State Division of Wildlife to enable the least disruption to habitat and compliance with the State law on protection of species of special concern, such as Preble's meadow jumping mouse.

Long Term Effectiveness and Permanence

This alternative would remove contaminated groundwater and soil including the residual concentrations and any non aqueous phase that may act as a source thereby significantly reducing potential risks to human health and the environment. The existing french drain extraction well and Building 891 water treatment system would continue to extract and treat contaminated groundwater migrating from IHSS 119.1 until the concentrations of COCs in groundwater at OU 1 are reduced below MCLs. Therefore, the residual risk would be reduced as compared to the no action and institutional controls alternatives.

Following treatment of the source, contaminated groundwater within OU 1 would continue to migrate away from IHSS 119.1. Modeling indicates that under this alternative groundwater would continue to meet MCLs for the COCs at Woman Creek. A five-year review would be conducted to determine the continued effectiveness of this alternative.

This alternative would provide long term protection for potential human receptors and minimize the human health risk associated with contaminated groundwater by continuing to achieve MCLs at Woman Creek. However, although the MCLs can be achieved with this alternative by remediating the soils and groundwater, this alternative may not be effective at removing DNAPLs if they are present at OU 1 outside the area to be excavated.

Excavated soils would be treated on site by thermal desorption and shipped for off site disposal at a properly permitted facility. There are no significant risks associated with handling and shipping the treated soils.

Reduction of Toxicity, Mobility, or Volume Through Treatment

Alternative 7 satisfies the NCP preference for treatment as a principal element of the alternative. This alternative would effectively and irreversibly reduce the mobility and volume of contaminants in OU-1 by removing any secondary source of contaminants from the subsurface. Excavating the soils within the treatment zone would reduce the volume of COCs in subsurface soils in both the saturated and unsaturated zones. Dewatering the excavation during remediation would likewise reduce COC volumes in groundwater. Removing the residual sources of contaminants will also reduce their mobility by preventing potential additional migration.

COC-contaminated soils removed from the excavation would be treated using thermal desorption. This would reduce the volume and toxicity of COCs in soils prior to off site disposal at a properly permitted facility. In addition, extracted groundwater would be treated in the Building 891 water treatment system, which is a destructive treatment process and thus would result in a further decrease in toxicity.

Contaminated materials generated as a result of this alternative include spent ion exchange resins from the Building 891 water treatment system and treated soils excavated from the treatment zone. There are no significant risks associated with handling and shipping either the ion exchange resins or the treated soils.

Short Term Effectiveness

This alternative would have significant short term impacts on the environment within the treatment zone at IHSS 119 1. Short term impacts to human health and the environment associated with this alternative include potential worker and public health exposure to airborne dust created during excavation and the displacement or loss of vegetation.

Alternative 7 would have a significant short term impact on the environment due to the large excavation and material transportation requirements. Excavating the entire source area would negatively impact the site flora as well as subsurface biota.

During implementation there would be the potential for risk to the public due to potentially contaminated air borne dust generated during excavating activities and the transportation of large quantities of excavated soils off site. There may also be potential risks to workers from exposure to COCs in groundwater or air borne dust. Workers would also have potential safety hazards associated with operating the earth-moving equipment and the thermal desorption unit. However, risks to workers would be minimized through standard health and safety practices.

Implementability

Excavation would be implemented using standard earth-moving equipment. However, the potential for radionuclide contamination in the excavated soils may limit the ability to transfer the soils off site. A large area would be required for stockpiling and treating the excavated soils, however, there is sufficient space available adjacent to the area to be excavated.

Excavated soils would be treated by thermal desorption, a proven and commonly applied soil remediation technology. The implementability of this alternative would not be limited by the availability of services and materials, nor would there be any significant technical or

administrative difficulties associated with this alternative Groundwater monitoring would be used to determine the effectiveness of this alternative

Cost

Costs for this alternative include the costs for the following items

- Conventional excavation/backfill earth moving equipment
- UV/peroxide/ion exchange water treatment system operation
- Thermal desorption unit
- Disposal of treated soil at licensed facility
- Associated piping, pumps and materials
- Groundwater monitoring for 30 years

The capital cost for Alternative 7 is \$11 326,100 O&M and post-closure activities for Alternative 7 include groundwater monitoring for 30 years The present worth O&M cost for this alternative is \$0 the post-closure cost is \$1,767,600 The total cost of this alternative is \$13 093 700 This cost estimate does not include the cost for disposal of radionuclide contaminated soil, if any is encountered A detailed cost estimate for capital and O&M costs for this alternative is included in Appendix E

4 3 Comparative Analysis of Alternatives

This section presents the comparative analysis which assesses the relative performance of each alternative in relation to each specific evaluation criteria excluding the two modifying criteria (state and EPA acceptance and community acceptance) The comparative analysis is summarized in Table 4-1

4 3 1 Overall Protection of Human Health and the Environment

All of the seven remedial action alternatives proposed meet key ARARs at OU-1 Specifically under each alternative groundwater would continue to meet the MCL for PCE the indicator

chemical at Woman Creek. Other contaminants appear in orders of magnitude lower concentrations (see Appendix B). All of the remedial action alternatives also satisfy the criterion of overall protection of human health and the environment by continuing to achieve groundwater MCLs at Woman Creek for the COCs identified at OU 1. For all alternatives, risk levels for a receptor at Woman Creek would be less than 2×10^{-6} . Alternatives 0, 1, 2, and 3 would each satisfy the criterion, although they do not address the source area at IHSS 119.1. Treatment alternatives 4, 5, 6, and 7 would likewise satisfy the criterion, although to a greater extent.

Alternative 0, No Action, would result in the highest risk of all the alternatives, although this risk (2×10^{-6}) is still within the established protective risk range of 10^{-4} to 10^{-6} . Alternative 1 would not present any human health risk because the entire RFETS site would be restricted under this scenario, thus eliminating the potential for exposure. Alternatives 2 and 3 would have the next lowest risk for a receptor at the edge of the operable unit; these two alternatives involve the continued operation of the existing french drain and extraction well and result in a risk of 4.76×10^{-9} at Woman Creek. The four alternatives that address the source at IHSS 119.1, Alternatives 4, 5, 6, and 7, have slightly higher risk levels because under these alternatives the french drain is decommissioned as soon as the source is remediated, allowing low concentrations of contaminants to continue migrating away from OU 1. However, these concentrations drop off rapidly after remediation is complete and result in lower long term risks to human receptors: 3.22×10^{-7} .

To the extent that COCs would be removed and treated, Alternatives 2, 3, 4, 5, 6, and 7 would all provide some degree of permanence. However, the source treatment alternatives (4, 5, 6, and 7) would be more effective in the long term. Alternative 7, which would remove the source by excavation, would provide the most permanence because of the uncertainty associated with the other alternatives' ability to effectively remove DNAPLs.

There would be no significant increase in potential risks to the public under Alternatives 0, 1, 2, 3, 4, 5, or 6. Alternative 7, however, would increase the risks to the public due to the potentially COC-contaminated dust that could become airborne during excavation activities and

off site transportation of excavated soils

There would be no significant risks to on site workers under any of the alternatives. These risks would be minimized through standard health and safety practices. Alternatives 5, 6, and 7 would have the most significant impacts on the environment due to in situ heating, mixing, and excavation, all of which are disruptive to the subsurface.

Under this criterion, the alternatives are ranked as follows:

<u>Rank</u>	<u>Alt. #</u>
1	2, 3
2	7
3	4, 5, 6
4	0, 1

4.3.2 Compliance with Applicable or Relevant and Appropriate Requirements

All alternatives would comply with the chemical specific, action specific, and location specific ARARs identified for OU 1. Compliance with the action specific ARARs, specifically state RCRA regulations, depends on CDPHE designating the OU 1 area a CAMU under Subpart S of the hazardous waste management regulations (6 CCR 1007.3.264.552). The closure, post closure, and groundwater monitoring requirements would be specified in the designation according to the selected alternative. Remediated waste, including groundwater and soil contaminated with hazardous waste constituents, is allowed to be left in place at the time of closure of the unit under the CAMU rule, providing certain provisions of the regulations are complied with and a determination is made that in taking such action there would not be risks to human health or the environment. Since all the alternatives meet the MCLs at Woman Creek, other differences in compliance with the ARARs are presented. There are possibly two differences in how some of the alternatives comply with the CAMU rule: the State Non game Endangered or Threatened Species Conservation Act and DOE regulations on compliance with wetlands protection.

Under this criterion the alternatives are ranked as follows

<u>Rank</u>	<u>Alt. #</u>
1	4 5,6 7
2	2 3
3	0 1

4 3 4 Reduction of Toxicity, Mobility, and Volume Through Treatment

Alternatives 0 and 1 which would not continue to operate the french drain or remediate the source would not reduce either toxicity mobility or volume of contaminants through treatment Alternative 2 and 3 would reduce the volume of contaminants and prevent migration of contaminants away from the operable unit by continuing to operate the existing french drain and extraction well These alternatives would also reduce toxicity by treating extracted groundwater with through the existing UV/peroxide/ion exchange process (Building 891 water treatment system)

Alternatives 4 5 6 and 7 would reduce the toxicity mobility and volume of contaminants further and more rapidly than the no action and institutional control alternatives because they would treat the source area at IHSS 119 1 Alternative 5 enhanced SVE would reduce volume and mobility more effectively than Alternative 4, SVE without subsurface heating Alternative 6 hot air injection and mechanical mixing, would reduce toxicity, mobility, and volume more effectively than Alternative 5 and Alternative 7 excavation would still more effectively reduce contaminant toxicity, mobility, and volume

Under this criterion the alternatives are ranked as follows

<u>Rank</u>	<u>Alt. #</u>
1	4 5 6 7
2	2 3
3	0 1

4 3 5 Short Term Effectiveness

Alternative 0 which would include only groundwater monitoring would not create any short term risks for the local community or the environment Likewise Alternative 1 which only includes institutional controls in addition to monitoring would not create any short term risks Alternatives 2 and 3 which would continue to operate the existing french drain and extraction well would not create any additional risks to the community or the environment as compared to the existing conditions at OU 1

Alternatives 4 5 6 and 7, which would remediate the source area at IHSS 119 1 would not create any significant potential for risk to the local community with the exception of Alternative 7 which may generate dust during excavation due to the large quantities of soil to be moved Under Alternative 7 there would also be risks associated with transporting the large quantities of excavated soils off site Each of the source remediation alternatives would present potential risks to workers during construction and operation however, these risks would be minimized through standard health and safety practices Alternative 4 would have minor impacts on the environment Alternatives 5 and 6 would have significant short-term impacts on the environment due to subsurface heating which may have adverse effects on biota In addition Alternative 6 would thoroughly mix the soil, increasing the impact on the subsurface environment Alternative 7 would also have significant impacts on the environment because a large volume of soil would completely removed

Under this criterion, the alternatives are ranked as follows

<u>Rank</u>	<u>Alt. #</u>
1	0 1,2
2	3
3	4 5 6
4	7

4 3 6 Implementability

Alternative 0 would be the most readily implementable alternative because it requires only groundwater monitoring. Alternative 1 would require institutional controls over the entire RFETS site such as designating the site a wildlife refuge and this could present some administrative problems although it effectively represents current conditions. Alternatives 2 and 3 the other institutional controls alternatives, would be readily implementable as compared to Alternative 1 and the other treatment alternatives.

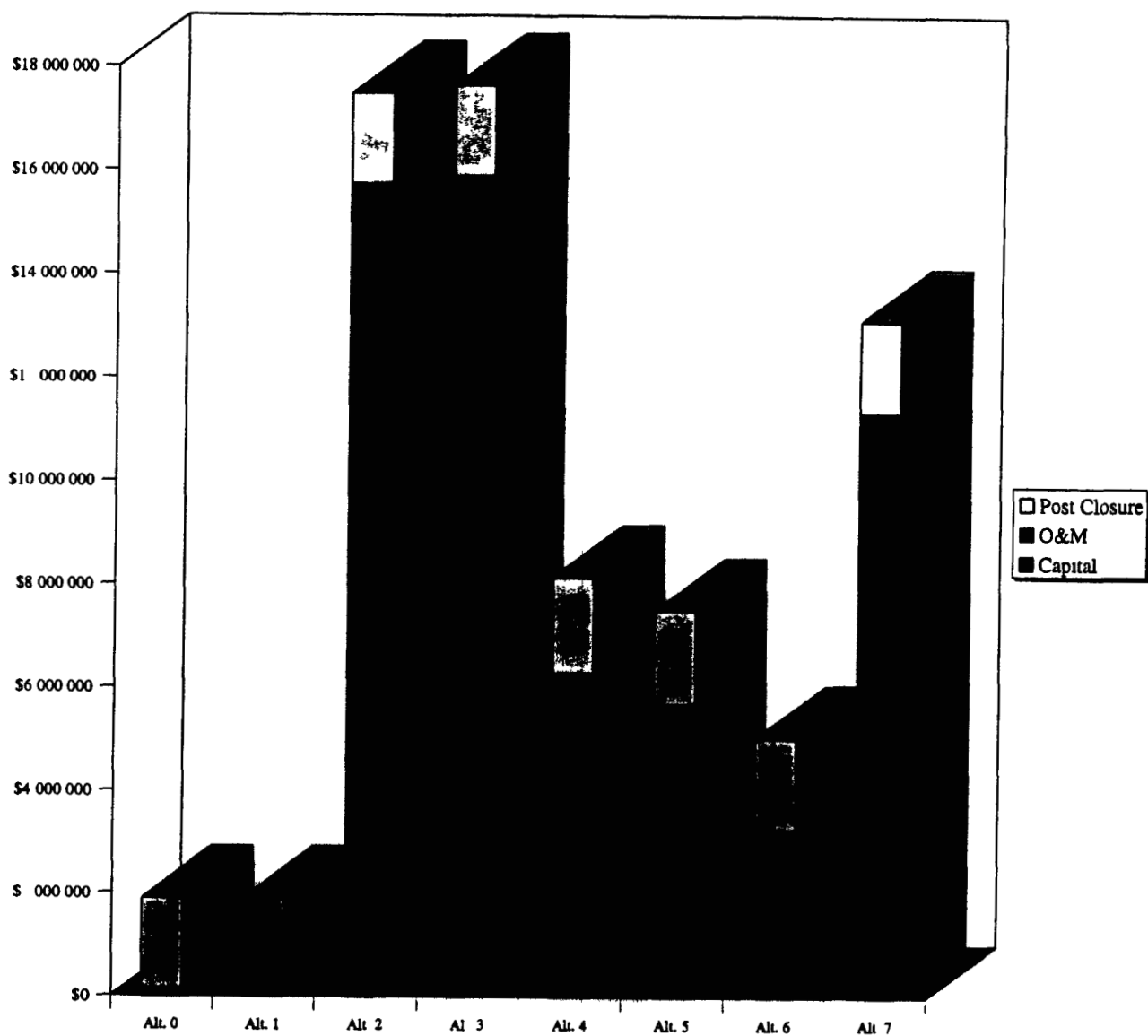
Alternatives 4 and 5 would also be readily implementable with no significant technical or administrative difficulties anticipated. However, Alternative 6 may be difficult to implement because of the instability of the subsurface soils within the treatment zone. Alternative 7 which would require large quantities of soil stockpiled and treated on site, and then transported off site for ultimate disposal would also be more difficult to implement.

Under this criterion, the alternatives are ranked as follows:

<u>Rank</u>	<u>Alt. #</u>
1	2
2	0 1
3	3
4	4,5,6
5	7

4 3 7 Cost

Figure 4-4 presents a comparison of the costs for the seven remedial alternatives. Alternatives 0 and 1 are the least costly alternatives because they involve only the continuation of groundwater monitoring. Alternatives 2 and 3 are significantly more costly than the other alternatives including the source treatment alternatives (4, 5, 6 and 7) due to the cost of operating the Building 891 water treatment system for 30 years. Of the four source treatment alternatives, Alternative 7 has the highest cost due to the need to treat the excavated soils to



Cost Element	Alt 0	Alt 1	Alt 2	Alt 3	Alt 4	Alt 5	Alt 6	Alt 7
Capital	\$154 700	\$154 700	\$149 600	\$305 000	\$929 300	\$1 845 700	\$1 354 400	\$11 326 100
O&M	\$0	\$0	\$15 603 300	\$15 603 300	\$5 358 700	\$3 845 700	\$1 887 300	\$0
Post Closure	\$1 740 400	\$1 740 400	\$1 740 400	\$1 740 400	\$1 853 800	\$1 811 700	\$1 789 100	\$1 767 600
Total Cost	\$1 895 100	\$1 895 100	\$17 493 300	\$17 648 700	\$8 141 800	\$7 503 100	\$5 030 800	\$13 093 700

Note: Costs represent 1994 dollars at 5% discount rate

Figure 4 4 Summary of Remedial Action Alternative Costs

meet Land Disposal Restriction treatment standards and dispose of them at an off site hazardous waste landfill

Under this criterion, the alternatives are ranked as follows

<u>Rank</u>	<u>Alt. #</u>
1	0 1
2	6
3	4,5
4	7
5	2,3

4 4 Preferred Remedial Action Alternative

Based on the results of the comparative analysis of alternatives, **Alternative 1 Institutional Controls without the French Drain** is selected as the preferred remedial action alternative for OU-1 This alternative is intended to minimize the risk from contaminated groundwater in OU 1 by restricting access to any wells impacted by OU-1 contaminants, and by eliminating the possibility of building construction above areas known to be contaminated with VOCs This alternative would be implemented by maintaining the current institutional controls present at the RFETS to at a minimum the Woman Creek drainage

4 4 1 Description

This alternative assumes that the existing french drain system would not be actively pumped but instead would be maintained and monitored as a contingency, in case groundwater contaminant concentrations begin exceeding predicted values If this occurs, water collected in the french drain sumps would be pumped to the Building 891 water treatment system The alternative cost estimate includes decommissioning the drain because it is assumed that at some point the drain will no longer be required

Groundwater monitoring would be required for this alternative to determine when institutional

controls could be discontinued. Once acceptable groundwater contaminant concentrations are achieved through natural degradation and dispersion of contaminants, the area would be released from institutional controls. Groundwater monitoring would take place for as long as required to meet this criterion. It is assumed that six monitoring points would be sufficient under this alternative. For cost estimating purposes it is assumed that four new wells would be installed: one deep and shallow well cluster downgradient of IHSS 119.1 and possibly two additional wells upgradient of Woman Creek. Samples would also be collected from the french drain sump and from the existing recovery well. Samples would be analyzed for organic and inorganic contaminants and would be collected semiannually. Analysis of individual species of inorganic contaminants is also suggested, to identify individual metal species which have the potential to bioaccumulate. This additional analysis requirement should only be applied occasionally in the sampling program. PQLs for analysis of these samples would be established to meet CDPHE criteria.

4.4.2 Summary of Detailed Analysis

This alternative would be protective of human health assuming that the institutional controls are properly implemented and that the site is not abandoned during the institutional control period. The risk to human receptors is 1.99×10^{-6} at Woman Creek. Risks from surface soils at OU 1 are within the acceptable risk range of 10^{-4} to 10^{-6} . Likewise, environmental receptors would be protected because there is no current or future risk identified from groundwater COCs for these receptors. There are no potential short-term risks to the public or to on-site workers identified through implementation of this alternative.

Under this alternative, the french drain would not be used to actively remediate contaminated groundwater, however, concentrations of contaminants in downgradient groundwater would gradually be reduced over time due to natural physical and chemical processes, such as dispersion, volatilization, and biodegradation. Because these are natural processes, they are essentially irreversible and would effectively reduce the toxicity, mobility, and volume of contaminants permanently.

This alternative is the most implementable of all alternatives identified because it utilizes the existing controls present at the RFETS. Costs associated with this alternative are similar to the no action alternative, and are lower than all other alternatives.

Key ARARs would be met under this alternative. The results of groundwater monitoring and modeling indicate that groundwater at Woman Creek currently does not exceed MCLs. Modeled contaminant concentrations projected 400 years from 1969 also indicate that there will be no exceedance of MCLs at Woman Creek within the 400-year period. Groundwater modeling results demonstrate that the highest concentration of PCE (the indicator COC) during the 400-year period is 3.60×10^{-3} mg/l under this alternative. Other COCs in OU-1 result in significantly lower concentrations at Woman Creek (see Appendix B). These results are considered extremely conservative due to the assumptions used to develop the model. The model assumes an infinite source of contamination and does not account for volatilization of contaminants, a potentially significant loss mechanism. Appendix B includes details concerning the groundwater model.

Alternative 1 will meet substantive requirements of the State RCRA program assuming CDPHE staff designates the OU-1 area as a CAMU under the recently adopted Subpart S provisions (6 CCR 1007.3 Part 264, Section 552). The CDPHE is required to determine that the unit will facilitate implementation of a reliable, effective, protective, and cost-effective remedy. In addition, CDPHE can require closure, post-closure, and any groundwater monitoring determined necessary to detect direction and movement of hazardous constituents. This report provides documentation to meet CAMU requirements and to suggest an appropriate monitoring program. The final CAD/ROD document will incorporate public comments and will provide all necessary documentation required by the CAMU rule regarding public participation.

In summary, Alternative 1 is protective of human health and the environment at minimal additional expense. The alternative will comply with identified ARARs and will be effective in reducing contaminant concentrations through natural processes. The alternative provides for natural attenuation with minimal additional impacts. Due to the limited availability of

groundwater in the saturated zone beneath OU 1 this alternative presents the most cost-effective and practical remediation alternative for OU 1 groundwater

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APPENDIX A

INITIAL SCREENING AND EVALUATION OF TECHNOLOGIES AND PROCESS OPTIONS

A 1 0 INTRODUCTION

The purpose of this appendix is to summarize the initial screening and evaluation of technologies and process options for the RFETS OU 1 CMS/FS. This screening and evaluation was presented in detail in *Technical Memorandum #11 Development and Screening of Remedial Action Alternatives 881 Hillside Area (OU 1) (April 1994)*. The screening and evaluation matrices for the groundwater medium are presented.

Additionally, although radionuclide contamination in surface soils has been included within the scope of the OU 2 CMS/FS, technology identification and screening of remedial technologies were performed prior to the determination to include OU 1 surface soil radionuclides in the larger OU 2 contamination plume. Work completed to date on identification, screening, and evaluation of technologies appropriate for contaminants identified in OU 1 surface soils is also presented through the attached figures.

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
No Action	None	Not applicable	Required for consideration by the National Oil and Hazardous Substances Contingency Plan	Potentially applicable as baseline against which other GPRAs/alternatives can be compared during detailed analysis
	Monitoring	Groundwater Monitoring	Monitoring of groundwater in operable unit during and after remediation, or as part of an institutional control period associated with the No Action alternative	Potentially applicable for monitoring site-specific groundwater conditions
Institutional Controls	Access Restrictions	Legal restrictions on access	Restrictions on present and future access to land prevent unauthorized access to groundwater source	Potentially applicable for controlling access to groundwater sources and/or exposure to COCs
		Fencing or other physical barriers	Fencing, security posts, limited roads and other various physical restrictions limit access to groundwater sources	Potentially applicable for controlling access to groundwater sources and/or exposure to COCs
		Legal restrictions on land use	Restrictions on present and future use and/or purchase of land; includes actions such as zoning and deed restrictions	Potentially applicable for controlling use of land affected by contaminated groundwater zones

Double lines surrounding a process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability or feasibility

Figure A 1 Initial Screening of Technologies and Process Options for Groundwater

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
Containment	Vertical Subsurface Flow Control	Subsurface Drains	Gravity driven collection system which is used to redirect groundwater flow and/or collect it for treatment	Potentially applicable. Includes possibility of modifying existing french drain system for use during remediation
		Grout Curtains	Grout "columns" are injected vertically into the soil in close proximity of each other to form an impermeable wall	Would not contribute additional containment because of existing low hydraulic conductivity
		Slurry Walls	A soil/bentonite or cement grout wall formed by backfilling a trenched area, has a lower permeability than native soils	Not implementable because of hillside stability concern; trenching may lead to slumping of native soils
		Sheet Pile Walls	Steel forms which are driven into the ground and joined to form a barrier which is impermeable to groundwater	Very difficult to implement due to proximity of bedrock; not widely used or accepted in cleanups
		Cryogenic Barrier	A section of ground is frozen to reduce its permeability thus limiting the mobility of contaminants through the area	Only applicable as a short term measure to control the migration of contaminants through an area
	Horizontal Subsurface Flow Control	Grout Injection	Grout is injected in a horizontal pattern beneath surface soils to limit vertical migration of VOCs from groundwater	Not applicable for remediation of VOCs in groundwater in fractured bedrock
		Block Displacement	Innovative use of grout forms perimeter barrier around waste while displacing waste upwards to block pathway	Not applicable for control of VOC that result from volatilization of groundwater contaminants nor for use in fractured bedrock
	Vapor Containment	Surface Cap	Compacted soil and bentonite cap used to reduce water infiltration to subsurface, and to contain VOC emissions	Potentially applicable for reducing vapor phase transport to surface structures
		Environmental Isolation Enclosure	One of several types of temporary structures used to contain/collect fugitive vapors and dust during remedial action activities, utilizes additional off-gas treatment	Potentially applicable for scenarios which would involve excavating soil to reach groundwater

Double lines surrounding process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability or feasibility

Figure A 1 Initial Screening of Technologies and Process Options for Groundwater (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
Removal	Passive Removal	Subsurface Drains	Gravity driven collection system which is used to redirect groundwater flow and/or collect it for treatment	Potentially applicable: includes possibility of modifying existing french drain system for use during remediation
	Active Removal	Horizontal and/or Vertical Extraction Wells or Bumps	Systems consisting of well installed either vertically or horizontally that are used to collect/recharge groundwater	Potentially applicable for removing contaminated water for treatment, for diverting groundwater flow or for lowering localized water table
	Excavation	Loader/Excavator/Dozer	Tractor/wheel mounted vehicles commonly used to excavate or move large amounts of soil can operate at various depths	Potentially applicable for removal of subsurface soil to locate groundwater hotspots
In-Situ Treatment of Chlorinated Solvents	Biological	Bioremediation	Destroy organics through microbial degradation; methanotropic process is specific to chlorinated solvents	Potentially applicable for in situ treatment of organic compounds in groundwater; however degradation products may be more harmful than original contaminants
	Chemical	Polymerization Chemical Oxidation	Catalyst injected into groundwater causes polymerization of organic monomers forming a gel-like, non-mobile mass Breakdown of organics using chemical which are typically introduced into the subsurface via injection wells or by drilling directly into the edge or within contaminant plume	Contact between reagent and groundwater is eventually overly hindered by the formation of the gel-like mass Difficult to apply because of concerns over injecting additional chemicals into the subsurface which may result in the formation of hazardous oxidation products
Physical		Hot Air/Steam Stripping with Mechanical Mixing	Hot air or steam is injected into the groundwater to promote the volatilization of VOCs which have low vapor pressures	Potentially applicable to remove VOC which are less likely to be volatilized through conventional means
		Air Sparging	Pressurized air is injected below or within a contaminated groundwater plume to cause in situ stripping of VOCs	Potentially applicable for in situ treatment of VOCs in groundwater
		Vapor Extraction	Induced negative pressure above saturated zone collects volatilized contaminants for treatment	Potentially applicable for removal of VOC from groundwater or for supporting other technologies (e.g. air sparging)
		Permeable Treatment Beds	A fixed bed containing treatment media is placed down-gradient of a groundwater plume to treat water in situ	Potentially applicable for in situ treatment of organic compounds in groundwater (including VOC in vadose zone) however limited by site hydrogeology
		In Situ Adsorption w/Walls (proprietary process)	Adsorption of organic contaminants in groundwater through the use of proprietary resin beads placed in existing well	Potentially applicable for in situ treatment of organic compounds in groundwater (including VOC in vadose zone) but not implementable due to low hydraulic conductivity
		RF/Ohmic Heating	Contaminants destroyed and/or destroyed by energy absorbed from radio frequency or ohmic sources	Potentially applicable for in situ treatment of organic compounds; effectiveness not dependent on conductivity or presence of groundwater

Double lines surrounding process option or technology denote option that were screened out from further consideration on the basis of technical implementability, applicability or feasibility

Figure A 1 Initial Screening of Technologies and Process Options for Groundwater (Cont)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
<div></div>	Biological	Bioremediation	Destroy organics through microbial degradation methanotropic process is specific to chlorinated solvents	Potentially applicable for ex situ treatment of organic compounds in groundwater however degradation products may be more harmful than original contaminants
		Chemical	Solvent Extraction	Not feasible for groundwater with low VOCs concentrations solvent would still require treatment/disposal
			Ultraviolet Photolysis with Chemical Oxidation	Potentially applicable for destroying organic compounds in extracted groundwater may include modification of existing UV/peroxide treatment system
	Physical	Gamma Irradiation	Innovative technology which decomposes organic compounds by destroying their chemical bonds using gamma irradiation	Not widely documented as to its use in the treatment of organic wastes also not feasible for low contaminant levels
		Activated Carbon or Carbonaceous Adsorbents	Extracted groundwater is passed through activated carbon which adsorbs most of the organic contaminants	Potentially applicable for removing organic compounds from extracted groundwater carbon could be disposed of or regenerated
		Air Stripping	Water is sprayed through a packed tower designed to increase its surface area to air ratio, thereby promoting volatilization	Potentially applicable for removing volatile organics compounds from extracted groundwater
		Membrane Processes	Application of an osmotic pressure forces contaminants to flow through semi-permeable membrane against diffusion	Not directly applicable for treatment of VOC in groundwater more commonly used to remove particulates
		Hot Air/Steam Stripping	Similar to air stripping but uses hot air or steam to remove VOCs which have relatively low vapor pressure	Potentially applicable for removing volatile organic compounds from extracted groundwater
		Evaporation	Concentration method used to drive off solvent from an aqueous waste stream using man-made and/or natural means	Not applicable as a stand-alone treatment technology more often used as pre-treatment step for a process
		Freeze Crystallization	Method of removing dissolved organic species by freezing the supporting matrix and crystallizing the solvent for separation	Only feasible for aqueous waste streams where organic contaminant concentrations are above 3-7% by weight
		Incineration	Destruction of organics through combustion with oxygen using a thermal and/or catalytic process option	Generally not applicable for liquids treatment at low contaminant concentration level
	Thermal	Plasma Arc Discharge	Pyrolysis of organics by high temperature plasma induced through electrical discharge to carrier gas	Potentially applicable for destruction of refractory organics in extracted groundwater
		Catalytic Oxidation	Catalyst allows low temperature thermal degradation of halogenated hydrocarbons to carbon dioxide water and hydrogen chloride. Also destroys non-halogenated VOCs forming water and carbon dioxide	Potentially applicable for treatment of groundwater or air Can produce acid off-gas
	Ex Situ Treatment of Chlorinated Solvents			

Double lines surrounding a process option or technology denote option that were screened out from further consideration on the basis of technical implementability applicability or feasibility

Figure A 1 Initial Screening of Technologies and Process Options for Groundwater (Cont)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
In-Situ Treatment of Inorganics	Physical	Electrodes	Electrodes are inserted in boreholes and a current passed through the media, causing migration of ions to the oppositely charged electrode where they are extracted by conventional pumping	Potentially applicable for the removal of metal from contaminated groundwater
		TRU Clear (Proprietary Process)	A proprietary potassium ferrate solution is mixed with groundwater precipitating transuranic and heavy metals	Potentially applicable for treatment of extracted groundwater currently undergoing treatability studies at RFP
		Oxidation/Reduction	Chemicals are added to groundwater which alter the oxidation state of the metals, causing precipitation	Potentially applicable for treatment of extracted groundwater however limited by low treatment efficiency
		Ferrite Process	Ferrite particles sorb metals and precipitate out of solution	Potentially applicable for treatment of extracted groundwater similar to TRU-Clear process mentioned above
Ex-Situ Treatment of Inorganic	Physical	Magnetic Separation	A high gradient magnetic field is applied to groundwater which forces polar metal ions out of solution onto collector plates	Conceivably applicable for treatment of extracted groundwater however not an established technology for groundwater treatment
		Freeze Crystallization	Contaminated solution is evaporated to saturation and then crystallization is induced by heat removal	Not feasible for waste stream with low contaminant concentrations due to prohibitive energy costs
		Ion Exchange	Metal species exchanged for resin ions and bound onto ion exchange resin for disposal	Potentially applicable for treatment of some metal however not applicable to many metal species
		Evaporation	Contaminated waste volume reduced by evaporation of solution in which contaminants are dissolved/suspended	Not feasible for extracted OU1 groundwater due to prohibitive energy costs
	Chemical	Membrane Processes	Metals concentrated by passing the contaminated solution through a semi-permeable membrane	Potentially applicable for treatment of extracted groundwater however may have prohibitive energy costs
		Electrocoagulation	Neutralization and precipitation of metallic ions is induced by creation of neutralizing ions using electrical current	Potentially applicable for treatment of extracted groundwater
		Precipitation	Removal of inorganics from aqueous phase by changing oxidation state through addition of chemicals or energy	Potentially applicable for removal of inorganics from extracted groundwater however difficult to control for complex waste streams



 Doubled lines surrounding process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability or feasibility

Figure A 1 Initial Screening of Technologies and Process Options for Groundwater (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
No Action	None	Not applicable	May not achieve remedial action objectives although required for consideration by NCP	Difficult to implement if public concern is high regarding site conditions	Very Low Capital Very Low O & M
		Groundwater monitoring	Effective in monitoring site conditions, or with No Action alternative as an institutional control	Readily implementable depending on remedial alternative selected	Low Capital Low O & M
Institutional Controls	Access Restrictions	Legal restrictions on access	Effective for relatively short-term control of present and future access to groundwater	Difficulty in obtaining necessary legal restrictions may reduce implementability	Low Capital Very Low O & M
		Fencing or other physical barriers	Moderately effective for relatively short term control of present and future access to area	Readily implementable if area under consideration is already site property	Moderate Capital Low O & M
		Legal restrictions on land use	Effective for control of present and future use of land which is affected by remedial actions	Difficulty in obtaining necessary legal restrictions may reduce implementability	Low Capital Very Low O & M
Containment	Vertical Subsurface Flow Control	Subsurface Drains	Effective in diverting flow of groundwater around targeted areas to limit the mobility of contaminants	May be difficult to implement upgradient of plume due to proximity of buildings	Moderate Capital Low O & M
		Surface Cap	Effective in dispersing vapor plume and reducing localized atmospheric emissions	Readily implementable using common construction equipment	High Capital Moderate O & M
		Environmental Isolation Enclosure	Effective in preventing the inadvertent release of VOCs and dusts during remediation	Readily implementable with many vendors available as suppliers	Moderate Capital Low O & M
Removal	Passive Removal	Subsurface Drains	Effective in collecting ground water if the system is designed appropriately for site conditions	Modification of existing french drain would be readily implementable if required	Moderate Capital Very Low O & M
		Horizontal and/or Vertical Extraction Well or Sumps	Effective in diverting collecting or recharging groundwater when gradient is relatively flat	Readily implementable based on existing site conditions if few wells are involved	Low Capital Low O & M
		Excavation	Tractor/wheel mounted shovels commonly used to excavate or move large amounts of soil can operate at various depths	Readily implementable although may be limited by bedrock formations	Low Capital Moderate O & M

Figure A 2 Evaluation of Process Options for Groundwater

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
In-Situ Treatment of Chlorinated Solvents	Biological	Bioremediation	Effective in treating organics but difficult to monitor progress during in situ treatment; may result in residuals which require further treatment	Requires extensive treatability work to determine viability of microbial growth for existing site-specific condition	Moderate Capital Moderate O & M
		Hot Air/Steam Stripping with Mechanical Mixing	Most effective in removing VOCs and SVOCs from groundwater but air/steam must be collected	Innovative technology which is considered moderately difficult to implement	High Capital High O & M
	Physical	Air Sparging	Effective in removing volatile organics and volatile inorganics from groundwater in situ	Requires horizontal drilling below water table so air will reach contaminant areas	Moderate Capital Moderate O & M
		Vapor Extraction	Moderately effective in removing VOCs from saturated soils although limited by nature of contamination	Would require the use of extraction wells to temporarily depress the water table	Low Capital Moderate O & M
		RF/Ohmic Heating	Effective in removing organics from groundwater but off-gas collection and treatment required	Treatability studies required to optimize frequency and phase settings for RFP	Moderate Capital Moderate O & M
Ex Situ Treatment of Chlorinated Solvents	Biological	Bioremediation	Effective in treating organics but may possibly result in residuals which require further treatment	Readily implementable if all contaminants can be degraded under similar condition	Moderate Capital Moderate O & M
		Ultraviolet Photolysis with Chemical Oxidation	Effective and proven method of destroying organic contaminants in extracted groundwater	UV treatment system already exists on site and may be used w/o significant modification	High Capital High O & M
	Physical	Activated Carbon or Carbonaceous Adsorbents	Effective if used as a final polishing step in a groundwater treatment system	Readily implementable as this is a common technology supported by many vendors	Moderate Capital Moderate O & M
		Air Stripping	Effective in removing VOCs and some SVOCs from extracted groundwater in large volumes	Readily implementable as this is common technology supported by many vendors	Low Capital Moderate O & M
		Hot Air/Steam Stripping	Effective in removing VOCs and some SVOCs from extracted groundwater	Readily implementable but more difficult than air stripping due to addition of steam	Moderate Capital Moderate O & M
	Thermal	Plasma Arc Discharge	Effective in destroying organics including refractory halogenated compounds	Treatability studies required to optimize energy level and treatment times for RFP	High Capital Moderate O & M
		Catalytic Oxidation	Effective in destroying organics including refractory halogenated compounds	Treatability studies required to determine catalyst, temperature and residence time	High Capital Moderate O & M

Figure A 2 Evaluation of Process Options for Groundwater (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
In-Situ Treatment of Inorganics	Physical	Electrodesis	Effective in removing ionic inorganic species from groundwater but unproven for most inorganics	Requires extensive treatability studies to determine applicability to site contaminants	Moderate Capital High O & M
		TRU Clear (Proprietary Process)	Effective in removal of inorganic species from extracted groundwater to extremely low levels	Use of proprietary chemical available from single vendor may limit implementability	Moderate Capital Moderate O & M
Ex-Situ Treatment of Inorganics	Physical	Oxidation/Reduction	Effective in precipitating many inorganics, however is difficult to control for multiple species	Treatability studies required to determine reagents required for site contaminants	Low Capital Moderate O & M
		Ferriite Process	Effective in removal of metal and radionuclides from extracted groundwater by precipitation	Readily implementable using commonly available equipment and chemicals	Low Capital Moderate O & M
		Ion Exchange	Effective in removing virtually all inorganics from water however may require extensive pretreatment	Treatment system already exists on site and may be used w/o significant modification	Moderate Capital Low O & M
	Chemical	Membrane Processes	Effective in removing many inorganics from water however may require extensive pretreatment	Implementability may be limited by influent water quality and low COC concentrations	Moderate Capital High O & M
		Electrocoagulation	Effective in removing most inorganic ions from water however it is nonselective process	Extensive treatability studies required due to innovative status of technology	Moderate Capital High O & M
		Precipitation	Effective in removing most inorganic ions from water however it is a nonselective process	Treatability studies required to determine chemicals which best address site conditions	Moderate Capital High O & M

Figure A 2 Evaluation of Process Options for Groundwater (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
No Action	Non	Not applicable	Required for consideration by NCP	Potentially applicable as a companion to against other GRAs
Institutional Control	Monitoring	Long-term surface soil and air monitoring	Monitoring of site conditions within operable unit after remediation or as part of an institutional control period associated with the no-action alternative	Potentially applicable for monitoring site-specific surface soil conditions (i.e. fugitive dust monitoring)
		Short-term surface soil and air monitoring	Monitoring of site conditions within operable unit during remediation activities	Potentially applicable for monitoring site-specific surface soil condition (i.e. fugitive dust monitoring)
	Access Restrictions	Legal restrictions on access	Restrictions on present and future access to land prevent unauthorized access to contaminated areas	Potentially applicable for controlling access to areas which are subject to dust emissions or surface contamination
		Fencing or other physical barriers	Fencing, security posts, limited roads, and other various physical restrictions limit access to contaminated areas	Potentially applicable for controlling access to areas which are subject to dust emissions or surface contamination
		Legal restrictions on land use	Restrictions on present and future use and/or purchase of land: Includes actions such as zoning and deed restrictions	Potentially applicable for controlling use of land which may be contaminated or subject to hazardous dust emissions

Double lines surrounding a process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability, or feasibility

Figure A 3 Initial Screening of Technologies and Process Options for Soils

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
Containment	Capping	Chemical Sealant	Chemical stabilizers are mixed with surface soil to form a stabilized cover which is resistant to water infiltration	Not applicable for remediation due to limitations on design if not widely used or accepted in cleanup
		Clay / Soil Cap	Compacted soil and bentonite (g bentonite liner) cap used to reduce water infiltration and to mitigate erosion and dust hazards	Potentially applicable to prevent contact with surface soil which may be contaminated and to contain them in situ
		Cement based Cap	Concrete slab over area of concern minimizes water infiltration and mitigates erosion and dust hazards	Would not be appropriate for soil required and would not provide additional protection beyond other cap types
		Asphalt based Cap	Asphalt cover over area of concern minimizes water infiltration and mitigates erosion and dust hazards	Would not be appropriate for soil required and would not provide additional protection beyond other cap types
		Synthetic Cap	Flexible liner used as sole cover source to reduce water infiltration to subsurface, and to prevent dust emissions	Not applicable for remediation due to limitation on design life, not widely used or accepted in cleanup
		Multimedia Cap	EPA recommended cap design which contains several layers (e.g. clay geomembrane, soil vegetative) minimizes water infiltration, erosion and dust emissions	Potentially applicable to prevent contact with surface soil which may be contaminated and to contain them in situ
	Erosion Control	Diversion / Runoff Control	Surface water routing measures that seek to redirect flow to minimize erosion of soil and spreading of contaminants	Not considered necessary because operable unit topography currently provides for sufficient runoff/runoff control
		Wind Breaks	Mesh-like barriers used to reduce wind speeds over small areas to minimize erosion of soils and reduce dust emissions	Potentially applicable to reduce erosion rate and for dust control however wind breaks would not be considered a permanent solution
		Surface Armoring	Surface soils are held in place by covering with riprap or other debris, minimizes wind and water effects	Potentially applicable to reduce erosion if riprap does not interfere with remedial alternative selected
		Vegetation	Natural vegetation is used to provide firm upper soil layer to limit dust emissions and surface water effects on soils	Potentially applicable as natural method for erosion control may be implemented during remediation

Double lines surrounding a process option or technology denote options that were screened out from further consideration on the basis of technical implementability applicability or feasibility

Figure A-3 Initial Screening of Technologies and Process Options for Soils (Cont)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
Removal / Disposal	Excavation	Loader / Excavator / Dozer	Tactor/wheel mounted vehicles commonly used to excavate or move large amounts of soil can operate at various depths	Potentially applicable for excavation of surface and subsurface soil
		Pneumatic	Vacuum extraction method for removal of loose dry surface soils or pumpable liquids into tank trucks or storage containers	Potentially applicable for excavation of surface soil not applicable to subsurface soil
	Dust Control	Dust Suppressants	Various synthetic and natural compounds which are sprayed on surface soils to reduce fugitive dust emissions (e.g. water)	Potentially applicable to reduce dust emission during remediation of operable unit
		Temporary Structures	Light, easily constructed structures used during remediation that provide protection from the effects of wind and rain	Not feasible because of areal extent of contamination and not considered necessary for low levels of contamination
On-Site Disposal	On-Site Disposal	Geotextile Cover	Flexible geotextile membranes that can be used during remediation to cover surface soils and reduce dust emissions	Applicability is limited because of possibility that cover would interfere with process equipment and personnel
		Engineered On-Site Disposal Facility	On-site disposal facility designed to contain site-specific waste for a single operable unit or for the entire site	Potentially applicable for storage of contaminated surface and subsurface soil or residuals which result from the treatment of soil
	Off-Site Disposal	Permitted Off Site Disposal Facility	Existing facility which is permitted to accept operable unit-specific waste or remedial action waste treatment residual	Potentially applicable for storage of contaminated surface and subsurface soils or residuals which result from the treatment of soil

Double lines surrounding a process option or technology denote options that were screened out from further consideration on the basis of technical implementability applicability or feasibility

Figure A 3 Initial Screening of Technologies and Process Options for Soils (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
In-Situ Treatment of PAHs and PCBs	Biological	Bioremediation	Destroys organics through microbial degradation several methods are available for treatment of PAHs	Feasibility limited for treating PAH in surfac soil du to very slow degradation rates and difficulty in process control
		Chemical	Breakdown of PAHs using chemical which are typically introduced into the subsurface via injection wells or by spraying directly over the surface soil requiring treatment	Difficult to apply because of concerns over injecting additional chemicals into the surface soil which may result in the formation of hazardous reaction products
	Physical	Soil Flushing	Aqueous flushing agents are forced through soils via injection wells which flush contaminants into a collection system	Not feasible for large areas which require shallow soil treatment
		Vitrification	Electrical soil melting process that destroys most organics while containing other contaminants in a solid, glassy matrix	Applicability is limited by stability of hillside also not appropriate for low level of surface soil contamination
		Radio Frequency Heating w/ Vacuum Extraction	Radio frequency energy is applied via electrodes to heat surrounding soils thereby promoting volatilization of PAHs	Not feasible for large areas which require shallow soils treatment
		Shallow Soil Mixing	Upper layer of surface soils are mixed with lower layers to reduce contaminant exposure to erosion problems	Potentially applicable for surfac soils to limit the mobility of contaminants although may not prevent volatilization
		Vacuum Extraction	A vacuum is applied to subsurface soils to volatilize organics and remove inorganics that are in the vapor phase	Not applicable for removal of PAHs from soil in situ PAHs do not volatilize significantly at normal temperatures
		Hot Air/Steam Stripping	Hot air or steam is injected below surface soil to force organic contaminants to the surface for capture and treatment of semi-volatile contaminants	Not feasible due to difficulty in sufficiently superheating steam to ensure adequate temperatures for volatilization of semi-volatile contaminants

Double lines surrounding a process option or technology denote options that were screened out from further consideration on the basis of technical implementability applicability or feasibility

Figure A 3 Initial Screening of Technologies and Process Options for Soils (Cont)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
Ex-Situ Treatment of PAHs and PCBs	Biological	Bioremediation	Destroys organics through microbial degradation several methods are available for treatment of PAHs	Potentially applicable for treating PAH in excavated soil although limited by slow degradation rate
		Land Application	Application of a thin layer of waste over an area to promote natural biodegradation may include addition of nutrients	Not implementable because of possible radionuclide contamination in soil which would be spread during treatment
	Chemical	Ultraviolet Photolysis w/ Chemical Oxidation	UV radiation is applied to assist in oxidizing organic compounds (using various oxidizing agents) thereby destroying them	Potentially applicable for destroying PAH in excavated soils many aromatic reactions are UV (free radical) catalyzed
		Solvent Extraction	Removal of organic compounds from soil by mass transfer to an organic solvent which is then collected and treated	Limited feasibility for removing PAH from soil with low contaminant concentrations, solvent would still require treatment/disposal
	Physical	Soil Washing	A variety of cleansing agents can be used to "wash" soil of organic contaminants, soils can be replaced after treatment	Potentially applicable for removing organic compounds from soil although agent would still require treatment/disposal
		Stabilization / Solidification	Binding agents are mixed with contaminated soils in either batch or continuous process to stabilize/immobilize contaminants	Not appropriate for very low level of contamination especially for contaminants with low initial mobility
		Incineration	Destruction of organics through oxidation and/or pyrolysis Various methods are available (i.e. rotary kiln fluidized bed)	Potentially applicable for treatment of PAH and PCBs in surface soil although may not reach PRG target level
		Thermal Desorption	Organics are volatilized from soil through the addition of heat; catalysts may be used to enhance process	Potentially applicable for treatment of PAHs and PCBs in surface soil although may not reach PRG target levels
		Vitrification	Electrical soil melting process that destroys most organics while containing other contaminants in solid glassy matrix	Potentially applicable for treatment of excavated soil but may not be appropriate for low level of contamination
		Molten Salt / Sodium Flushing	Molten salt, air and soil are combined in a reactor to destroy organics through oxidation and to trap halogens	Not applicable for treatment of PAHs more commonly used for remediating chlorinated solvent contamination

Double lines surrounding process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability or feasibility

Figure A 3 Initial Screening of Technologies and Process Options for Soils (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	DESCRIPTION	SCREENING COMMENT
In-Situ Treatment of Radionuclides	Physical	Soil Flushing	Complexing or chelating agents are used to solubilize radionuclides and then extracting the contaminated flushing solution	Difficult to apply because of concern over solubilizing radionuclides which could then migrate toward groundwater
		Stabilization / Solidification	Binding agents are injected into contaminated soils and then hardened to stabilize/solidify contaminants	Difficult to implement due to large surface area and minimal depth of contamination
		Shallow Soil Mixing	Upper layer of surface soils are mixed with lower layers to reduce contaminant exposure to erosion problems	Potentially applicable for surface soil to limit the mobility of contaminants and to reduce potential human exposure level
	Thermal	Vitrification	Electrical soil melting process that encapsulates radionuclides and other contaminants in solid, glassy matrix	Applicability is limited by stability of hillside also not appropriate for low levels of surface soil contamination or large, shallow areas
		Soil Washing	A variety of cleansing agents can be used to "wash" soil of radionuclides, soils can be replaced after treatment	Potentially applicable for excavated surface soil although "wash" solution would require additional treatment/disposal
Ex-Situ Treatment of Radionuclides	Physical	Stabilization / Solidification	Binding agents are mixed with contaminated soils in either a batch or continuous process to stabilize/solidify contaminants	Potentially applicable for excavated surface soils contaminated with radionuclides
		Manganese Dioxide Adsorption	Radionuclides are adsorbed from slurried soil onto manganese dioxide particles	Not feasible due to its experimental nature and the extent of research and feasibility studies it would require; thus it has been screened out from RFP Site-wide feasibility studies
		Magnetic Separation	A high gradient magnetic field is applied to slurried soil which forces polar radionuclides out of slurry onto collector plates	Potentially applicable for excavated surface soil currently undergoing feasibility studies at the Los Alamos National Laboratory
		TRU-Clean (proprietary process)	Radionuclides are "washed" from slurried soils by a proprietary precipitation process	Potentially applicable for excavated surface soils currently undergoing feasibility studies at the Nevada Test Site
		Segmented Gate System	Radioactive particles above threshold activities are removed from soil and concentrated by diversion gates attached to radiation monitors	Potentially applicable for excavated surface soil May exhibit low treatment efficiency on fine well distributed radionuclide sources such as those at OU1
		Vitrification	Electrical soil melting process that encapsulates radionuclides and other contaminants in solid glassy matrix	Potentially applicable for excavated surface soil contaminated with radionuclides
	Thermal	Vitrification	Electrical soil melting process that encapsulates radionuclides and other contaminants in solid glassy matrix	Potentially applicable for excavated surface soil contaminated with radionuclides

Double lines surrounding process option or technology denote options that were screened out from further consideration on the basis of technical implementability, applicability, or feasibility

Figure A 3 Initial Screening of Technologies and Process Options for Soils (Cont)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
No Action	None	Not applicable	May not achieve remedial action objectives although required for consideration by NCP	Difficult to implement if public concern is high regarding site conditions	Very Low Capital Very Low O & M
Institutional Controls	Monitoring	Long-term surface soil and air monitoring	Effective in monitoring long-term site conditions after remediation, or as part of no-action alt.	Readily implementable depending on remedial alternative selected	Low Capital Low O & M
		Short-term surface soil and air monitoring	Effective in monitoring short-term site conditions to protect worker and public health and safety	Readily implementable depending on remedial alternative selected	Low Capital Very Low O & M
		Access Restrictions	Legal restrictions on access	Effective for relatively short-term control of present and future access to designated area	Difficulty in obtaining necessary legal restrictions may reduce implementability
	Fencing or other physical barriers		Moderately effective for relatively short-term control of present and future access to area	Readily implementable if area under consideration is already site property	High Capital Low O & M
	Legal restrictions on land use		Effective for control of present and future use of land which is affected by remedial actions	Difficulty in obtaining necessary legal restrictions may reduce implementability	Low Capital Very Low O & M
	Containment	Capping	Clay / Soil Cap	Moderately effective in preventing precipitation from reaching and mobilizing contaminated areas	Easiest to implement relative to other types of cap/covers available
Multimedia Cap			Most effective form of cap which is resistant to weathering and cracking over its design life	Moderately difficult to implement based on variety of materials required	High Capital Very Low O & M
Erosion Control		Wind Breaks	Marginally effective in reducing dust emissions or erosion over large areas of soil	Readily implementable does not require subcontractor support	Low Capital Low O & M
		Surface Armoring	Effective in reducing soil erosion caused by wind and precipitation also minimizes dust	Readily implementable if local supply of riprap or debris is available	High Capital Very Low O & M
		Vegetation	Effective in reducing soil erosion due to wind and precipitation if vegetation is maintained	Readily implementable requires common landscaping equipment available locally	Moderate Capital Moderate O & M

Figure A-4 Evaluation of Process Options for Soils

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
Removal / Disposal	Excavation	Loader / Excavator / Dozer	Effective for excavating soils and sludges less than 30 feet deep	Readily implementable - see common road building and construction equipment	Low Capital Moderate O & M
		Pneumatic	Effective in removing loose dry soils or pumpable liquids from ground surfaces and surface waters	Readily implementable	Low Capital Low O & M
	Dust Control	Dust Suppressants	Moderately effective for reducing surface dust generation depending on type of suppressant	Readily implementable although certain suppressants may be considered hazardous	Low Capital Moderate O & M
	On-Site Disposal	Engineered On-Site Disposal Facility	Effective in containing treated or residual wastes assuming the facility is designed properly	Difficult to implement because of permit requirements and administrative concerns	Very High Capital High O & M
	On-Site Disposal	Permitted On-Site Disposal Facility	Effective in containing treated or residual wastes if proper facility is available	Readily implementable for wastes other than TNU or mixed (radioactive/hazardous)	Moderate Capital Very Low O & M

Figure A-4 Evaluation of Process Options for Soils (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
In-Situ Treatment of PAHs and PCBs	Biological	Bioremediation	Effective in treating organics but difficult to monitor progress during in situ treatment; may result in residue which requires further treatment	Requires extensive treatability work to determine viability of microbial growth	Moderate Capital Moderate O & M
	Physical	Shallow Soil Mixing	Effective in treating upper soil layers in situ to prevent migration of contaminants	Readily implementable uses commonly available agricultural engineering equipment	Low Capital Moderate O & M
Ex-Situ Treatment of PAHs and PCBs	Biological	Bioremediation	Effective in treating organics but may possibly result in residue which requires further treatment	Requires extensive treatability work to determine viability of microbial growth	Moderate Capital Moderate O & M
	Chemical	Ultraviolet Photolysis or Chemical Oxidation	Effective method for destroying some organic compounds depending on UV lamp used in system	Implementability will depend on oxidation method chosen to accompany UV process	High Capital High O & M
		Solvent Extraction	Effective in removing volatile and non-volatile organic compounds from soils although spent solvent will require treatment or disposal	Moderately difficult to implement relative to other ex situ treatment options	High Capital Moderate O & M
	Physical	Soil Washing	Effective for removal of organic and inorganic contaminants; several washing agents available	Implementable technology which is based on commonly used ore mining technologies	High Capital Moderate O & M
		Incineration	Effective in destroying or removing organic contaminants to levels in the low ppm range	Implementable technology which has been used extensively for treating organics	High Capital High O & M
		Thermal Desorption	Effective for removing volatile and semi-volatile compounds from soil requires off-gas treatment	Implementable technology which has been used extensively for treating organics	High Capital High O & M
		Volatilization	Very effective for destroying organic waste while containing other contaminants such as metals	Innovative technology which is difficult to implement based on the complexity of equipment required to verify waste	Very High Capital Very High O & M

Figure A-4 Evaluation of Process Options for Soils (Cont.)

GENERAL RESPONSE ACTION	REMEDIAL TECHNOLOGY	PROCESS OPTION	EFFECTIVENESS	IMPLEMENTABILITY	RELATIVE COST
In-Situ Treatment of Radionuclides	Physical	Soil Flushing	Innovative technology which is effective for removing certain radionuclides from soil particles	Moderately difficult to implement; requires system to collect and treat flushing agent	High Capital Moderate O & M
		Shallow Soil Mining	Effective in stabilizing upper layer of soils in situ to prevent migration/contact with radionuclides	Readily implementable uses commonly available construction equipment	High Capital High O & M
Ex-Situ Treatment of Radionuclides	Physical	Soil Washing	Effective for removal of radionuclides from soil if proper washing agents are used in the process	Implementable technology which is based on commonly used ore mining technologies	High Capital Moderate O & M
		Stabilization/Solidification	Effective in containing radionuclides by containing them in a stabilized or solidified matrix	Moderately difficult to implement because of problems with long-term leach resistance	Moderate Capital Moderate O & M
		Magcrete Separation	Effective for removing trace amounts of metals from liquid waste streams, including radionuclides	Moderately difficult to implement	Moderate Capital Moderate O & M
		TRU-Clay (proprietary process)	Innovative proprietary process which is form of soil washing used specifically with rad waste	Readily implementable but requires consent of proprietary vendor for implementation	High Capital Moderate O & M
		Segmented Gas System	Effective for removal of discrete radioactive particles. Effectiveness for contaminants distributed by weathering would be determined by feasibility studies	Readily implementable Uses common sand-and-gravel handling equipment and common radiation monitors. Control software and design are proprietary	Low Capital Moderate O & M
		Vitrification	Very effective for containing radionuclides in glassy solid matrix which is resistant to leaching	Innovative technology which is difficult to implement based on its complexity	Very High Capital Very High O & M
	Thermal				

Figure A-4 Evaluation of Process Options for Soils (Cont)

**APPENDIX B
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B 1 0 INTRODUCTION

Appendix B presents the results of a subsurface solute transport model of the OU 1 site. The purpose of the model is to provide a basis for residual risk calculations and design calculations for the feasibility study. In this section the following topics are discussed: the hydrogeological conceptual model of the site, the framework of the corresponding numerical model, the results and predictions of the model, and a qualitative discussion of model uncertainty. Tables and Figures are included in the back of this appendix, after references.

B.2 0 HYDROGEOLOGICAL CONCEPTUAL MODEL

The conceptual model of OU 1 is a description of the primary processes that control the movement of solutes in the subsurface. Such processes include groundwater flow rates and directions, solute release rates and timing, recharge and discharge rates, dispersion, degradation rates, and adsorption.

The groundwater flow system beneath the hillside at OU 1 is described in detail in the Phase III RFI/RI (DOE 1994). The description here is limited to features incorporated into the flow and transport model of the site, and is further limited to the area of IHSS 119.1. IHSS 119.1 is where most of the observed contamination at the site is located.

Groundwater flow beneath the hillside occurs in shallow colluvial, alluvial and bedrock units. Most of the flow is concentrated in the colluvium and alluvium (DOE 1994). Groundwater flow tends to be focussed in areas where colluvium is thickest, these areas generally correspond to surface-water drainage features. Such correspondence is likely due to deeper weathering of bedrock beneath surface channels. One such surface channel feature extends upslope into IHSS 119.1 with a corresponding thicker section of colluvium. It is along this subsurface feature that most of the groundwater flow in the vicinity of IHSS 119.1 occurs (Figures 3-23 and 3-24 of the Phase III RFI/RI). Therefore, groundwater flow is generally channelized along surface water features.

Recharge and discharge probably vary during the year at the hillside. However, over long periods of time an average rate of recharge or discharge is applicable. Recharge to the hillside flow system is assumed to occur as subsurface flow from the Rocky Flats alluvium and from bedrock beneath the Rock Flats alluvium. No site specific measurements of recharge or discharge on the hillside are available. Discharge at the surface along the hillside is assumed to occur most of the time due to low precipitation rates, runoff due to topography, partially saturated conditions (with corresponding smaller relative water permeabilities), the small permeability of the colluvium and alluvium (both of which are derived from the claystone and siltstone of the bedrock) and frozen ground during winter months. Discharge is also assumed to occur as flow into Woman Creek as observed by Fedors et al (1993a and 1993b) and as indicated by hydraulic gradients directed toward the creek.

The primary source of contaminants is located in the subsurface beneath IHSS 119.1. During the 1960s and 70s, drums of solvents were stored in IHSS 119.1 (DOE 1994). Releases from the drums have resulted in a residual DNAPL phase in the subsurface around Well 4387. The residual DNAPL phase has not been directly observed, but is indicated by high concentrations of chlorinated solvents like PCE in groundwater. The start of release to groundwater is not precisely known but is assumed to be 1970. The release mechanism to groundwater is dissolution of the residual (immobile) DNAPL phase.

The transport of contaminants in groundwater is controlled primarily by groundwater flow directions and rates. Other processes that affect contaminant movement and mobility are dispersion, degradation, adsorption, and volatilization from groundwater to soil gas. Of these, volatilization is not included in the model. Dispersion is simulated using dispersivity, groundwater flow velocity, and molecular diffusion. Degradation rates and sorption properties for solutes are discussed and reported in the Phase III RFI/RI.

B 3 0 MODEL FRAMEWORK

The computer simulation code TARGET_2DU (Dames & Moore 1985) was used to simulate contaminant transport in the subsurface. TARGET_2DU is a vertically oriented two-dimensional finite difference model that can simulate variably saturated conditions.

Because the model is two dimensional it cannot simulate dispersion (spreading) transverse (perpendicular) to the model section. Therefore dispersion in the plane of the model will be over predicted parallel and transverse to groundwater flow. Consequently the model is more conservative (over predicts concentration and travel time) because it does not account for spreading of contaminants in transverse to the model plane.

Another conservative aspect of the TARGET_2DU is that the mass adsorbed on soil is not decayed. As constituents desorb the concentration on soil decreases but remains undecayed. The result is a source in the model that decreases but at a rate slower than if decay were calculated for contaminants on soil. Consequently the concentration of contaminants that desorb into water is higher than for the case in which decay on soil were calculated. For contaminants with halflives that are short relative to the groundwater transport time the degree of over prediction is significant and is conservative.

The model grid is 296 (horizontal) by 170 (vertical) cells (Figure B-1) with approximately 25 000 active cells. The grid was designed to capture details of the bedrock/colluvium interface and topography, to accurately simulate the vadose zone and to minimize errors caused by numerical dispersion. The location of the section of the model is shown in Figures B-2 and B-3, and corresponds to the trend of thicker colluvium which passes through IHSS 119 1.

Two criteria are used to ensure minimal numerical dispersion: the Peclet number and the Courant number. The grid Peclet number is the ratio of grid spacing (length of a cell side) to dispersivity. To minimize numerical dispersion the Peclet number should be less than or equal to one. For this model dispersivity is much larger than cell lengths, so the Peclet number is

much smaller than one. The grid Courant number is the ratio of time step interval to groundwater travel time across a cell. Similar to the Peclet number, the Courant number should be less than or equal to one. Because gradients and hydraulic conductivities are small and because decay rates of the COCs are short, the Courant number for this model is much smaller than one.

The distribution of boundary conditions and material types are shown in Figure B-4. Properties associated with each of the material types and degradation rates and adsorption distribution coefficients for the contaminants of interest are listed in Table B-1. The list of contaminants is presented and discussed in Section 2.0.

Each material type is assumed to be homogeneous where specified. Therefore, heterogeneity in the model is limited to three zones: colluvium, alluvium, and bedrock. Fractures in the colluvium resulting from mass movement of the colluvium down the hill are assumed to be healed, so that fractures do not provide preferential flowpaths. This is justified because the mass movement is generally relict (probably occurred during the Pleistocene), thus having had considerable time to heal, and because the colluvial material, being residuum bedrock, is easily deformed so that voids cannot remain open over long periods of time.

For the French drain, a constant head/pressure cell was set at the bottom of the drain to simulate flow to the drain (elevation of the constant head is 5876.2 ft). The extraction well was simulated in the same manner but with an elevation of 5910.2 ft (see Figure B-5). These elevations are slightly above the interface between bedrock and colluvium material. This was done based on the assumption that the French Drain and extraction well could not draw the water table all the way down to the interface (otherwise, the saturated thickness approaches zero and flow decreases to zero). Simulations using the French Drain and extraction well are discussed in detail in following sections.

The bottom of the model was selected to be somewhat lower than the elevation of Woman Creek, which is considered to be the ultimate sink for groundwater flowing down the hillside. Because

flow rates in the bedrock are much lower than those in the colluvium the model is less sensitive to the location of this boundary

The contaminant source was simulated using a constant concentration boundary condition based on the assumption that slow dissolution of residual DNAPL is the source of contamination in groundwater. The source cell is located at the interface between bedrock and colluvium material in the model, in the area where high concentrations of contaminants in groundwater have been observed (Figure B 5)

B 4 0 CALIBRATION

The model was calibrated to steady state average conditions as observed prior to the installation of the French Drain. For calibration targets, observed groundwater levels for wells 4387, 0487, 4787 and 5587 were compiled and averaged (Table B 2). For the purposes of computing target water levels, dry (no measurable water) conditions were excluded from the average. The results of the calibrated flow model are shown in Figure B 6 with point comparisons to average observed water levels (results of the flow and transport simulations are discussed in more detail in section B 5 0). To achieve calibration, a net areal discharge of 2.96 in/yr from the water table was used, as discussed in Section B 2 0.

The flow mass balance provides a measure of how well the model is converged. Discrepancies in the mass balance should be smaller than about 5%, especially for groundwater flow; otherwise errors in the flow domain may adversely affect subsequent transport simulations. For the OU 1 model, the percent discrepancy between simulated inflows and outflows for various times is (approximate values): 1.87% for steady-state flow, 2.3% at the end of 23 years, and 1.38%, 0.18% and 5.31% at the end of the three predictive simulations, respectively (see Section B 7 0). Convergence of the model was good, exhibiting monotonic behavior.

After calibrating the steady state flow, transient transport simulations were done for each contaminant. Transport simulations started with the steady state flow field continued for 20

years then incorporated the French Drain and extraction well as shown in Figure B 7 Each transport simulation was calibrated in a manner similar to that for the flow calibration Figures B 8 through B-17 show breakthrough curves for each of the contaminants with average upper bound and lower bound observed concentrations Data used to compute average minimum and maximum concentrations are listed in Table B 3 Calibrated source concentrations are 6 41 mg/l for 1 1 DCE 9 63 mg/l for PCE 0 64 mg/l for CCl₄ 16 mg/l for 1 1 1 TCA and 160 mg/l for selenium

As with flow, the contaminant mass balance provides a measure of how well the model is converged Discrepancies in the mass balance should be smaller than about 10% For the OU 1 model and PCE, the percent discrepancy between the simulated mass in place, and mass influx and outflux for PCE at various times is (approximate values) 2 5% at the end of 23 years and -4 51% 5 76% and 10 09% at the end of the three predictive simulations respectively (see Section B 7 0) Convergence of the model is good exhibiting monotonic behavior

B 5 0 RESULTS

From the calibrated steady state flow simulation (Figure B-6) groundwater rates and directions can be obtained Figure B-18 shows the effects of the French Drain and extraction well on groundwater flow The French Drain and extraction well both draw down the water table with drawdown cones that extend upgradient into IHSS 119 1 As expected the drawdown cones are asymmetrical due to the slope of the hill The simulated water levels correspond well with observed low water table conditions

Results of transport simulations for PCE are discussed in detail Results of simulations for other contaminants are not shown because the chemicals tend to behave similarly The PCE plume after 22 years (pre-French Drain) and at 23 and 24 years is shown in Figures B-19 B-20 and B-21 The plume moves down gradient slowly and also penetrates into the bedrock a small distance The majority of movement is in the colluvium due to higher groundwater flow rates

Some migration in the vadose zone is also simulated corresponding to dispersion in soil moisture

After 24 years the French Drain and extraction well have a slight effect on the plume (Figures B 20 and B-21) The extraction well pulls the plume back toward IHSS 119 1 and the French Drain captures the plume trapped between it and the extraction well

B 6 0 UNCERTAINTY

This section is a qualitative discussion of uncertainties associated with the model In general uncertainties can be divided into two types The first type results from an incomplete knowledge of the system or processes A real system can often be too complex or lack the necessary information to be completely understood or modeled without making simplifying assumptions Parts of the system or processes may also be omitted because they are thought to be less important than others The second type of uncertainty relates to the values assigned to model input parameters used to describe the system or processes In reality input parameters are not single values but vary over a range of possible values

Table B-4 lists specific model assumptions or uncertainty factors (parameters) that could contribute to variations in model predictions The second column of the table gives the source of the uncertainty "Not simulated" means a particular transport or transformation process was not considered in the modeling Measurement Error indicates that there could be some unknown, unmeasured variability or heterogeneity in the corresponding property Not measured indicates that the parameter has not been measured under site-specific conditions either in the field or in the laboratory In the third column, "Incorrect Flows" indicates that a different flow could result by a corresponding change in the parameter or assumption The fourth column lists the relative degree of uncertainty

Table B-4
Model Assumptions and Uncertainty Factors

Model assumption or uncertainty factor	Cause of uncertainty or model error	Probable effect on model results	Relative degree of uncertainty
Two dimensional model	Three-dimensional transport not simulated	Incorrect spatial distribution of concentrations and incorrect flows	Low Model adequately matches general trends in the horizontal behavior of the observed plume Model is conservative due to over predication of lateral spreading
Porous media	Flow in fractures or other secondary porosity not simulated	Incorrect spatial distribution of concentrations and incorrect fluxes	Low Although slip subsurface failure planes have been mapped (DOE 1994) it is likely that such potential pathways have healed and are no longer permeable
Steady state flow	Transient flow is not simulated for calibration	Incorrect spatial distribution of concentrations and incorrect flows	Low Contaminant transport and fluctuations in flow become less important over long periods of time The model is conservative in simulating continually saturated conditions where seasonal wetting and drying is known to occur
Material properties are homogeneous within a model layer	Heterogeneity within model layers	Incorrect spatial distribution of contaminants incorrect flows	Low The primary hydrogeologic layers that effect transport are well characterized
Volatilization	Not simulated	Incorrect spatial distribution of contaminants	Low Model is conservative with regard to this process
Timing of release	Not well known	Incorrect spatial distribution of contaminants	Low Model is generally conservative
Nature of release	Processes other than dissolution are not modeled	Incorrect spatial distribution of contaminants	Low Model is generally conservative
Sorption	Linear sorption	Incorrect spatial distribution of contaminants	Low Organic carbon content of subsurface and surface materials is low
Natural recharge/discharge rates	Not measured	Incorrect spatial distribution of contaminants incorrect flows	Moderate Model is sensitive to this parameter

**Table B-4
(Continued)**

Model assumption or uncertainty factor	Cause of uncertainty or model error	Probable effect on model results	Relative degree of uncertainty
Decay and transformation	Multicomponent transport not simulated no site-specific data	Incorrect spatial distribution of contaminants	Low Model is conservative
Porosity	Measurement error	Incorrect spatial distribution of contaminants	Low Measurement error relatively small
Diffusion coefficient	Not measured	Incorrect spatial distribution of contaminants	Low Error is small and model is insensitive to this parameter
Dispersivity	Not measured	Incorrect spatial distribution of contaminants	Moderate Parameter is based on scale of site this is a standard assumption
Size of source	Not measured	Incorrect spatial distribution of contaminants	Low Model has been shown to be insensitive to source size (Fedors et al 1993)

The combination of parameters used in the model is not considered to be unique. Other combinations of the parameters may yield a similar result. However, the parameter values used generally lie within observed and accepted ranges and therefore the model is considered representative of site conditions. However, the model is conservative in that it does not account for volatilization of PCE, DCE, TCA, and CCl_4 , and it generally over predicts concentrations at Well 0487. For these reasons, the model is considered to be highly conservative.

B 7 0 PREDICTIONS

For predictions in which the source is not remediated, the source is assumed to be large enough to provide an infinite supply of groundwater contamination. Therefore, in such simulations, the source concentration is held constant throughout the simulations. For predicative simulations in which the source is remediated, the concentrations in a 200 foot long area of colluvium around IHSS 119 1 are set to the appropriate water quality standard.

B 7 1 No Action Alternatives

Under these alternatives (0 1) French Drain and extraction well are removed, but the source is not remediated. Transport simulations beginning from 1994 and continuing through 2024 were done for each of the contaminants of interest. Figure B 22 shows the predicted PCE plume in 1998. Under this scenario, the plume continues to grow with time because the source remains in place, providing a constant source of dissolved PCE. In addition, desorption begins to provide an undecayed source, which results in conservatively high predicted concentrations.

Figures B-23 through B-32 show the variation of concentration with time at the French Drain and Woman Creek. These curves are typically called breakthrough curves. At the French Drain, the installation of the drain and extraction well cause a dip in concentrations. After the drain and well are removed, concentrations begin to recover and increase due to a continuing source and to desorption. At Woman Creek, similar results are obtained, however, due to the longer travel distance and time, the features of the breakthrough curves are more subdued. Peak concentrations are simulated for PCE and DCE.

B 7 2 Institutional Control Alternatives With the French Drain

Under these alternatives (2, 3) the french drain and extraction well remain in operation in the future. No remediation of the source takes place under this scenario. Transport simulations beginning from 1994 and continuing through 2024 were done for each of the contaminants of interest. Figure B-33 shows the predicted PCE plume in 1998. Under this scenario the plume is drawn to and captured by the extraction well and french drain. In addition, desorption begins to provide an undecayed source which results in conservatively high predicted concentrations.

Figures B-34 through B-43 show the variation of concentration with time at the french drain and Woman Creek. At the french drain, the installation of the drain and extraction well cause a dip in concentrations. With the drain and well in place, concentrations peak for shorter half-life COCs. Desorption still provides a decreasing but undecayed source. At Woman Creek, similar results are obtained, however, due to the longer travel distance and time, the features of the breakthrough curves are more subdued. Peak concentrations are simulated for PCE and DCE.

B 7 3 Remediation Alternatives

Under these alternatives (4, 5, 6, 7) the french drain and extraction well are removed, and the source is remediated. Transport simulations beginning from 1994 and continuing through 2024 were done for each of the contaminants of interest. For these simulations, where the source is remediated, a 200 foot long strip of colluvium assumed to be cleaned up to the appropriate water quality standard (see Table 2.1). Figure B-44 shows the predicted PCE plume in 1998. Under this scenario, the plume that remains in place after the source is removed continues to move down gradient with time. In addition, desorption begins to provide an undecayed source, which results in conservatively high predicted concentrations.

Figures B-45 through B-54 show the variation of concentration with time at the source, french drain, and Woman Creek. At the french drain, the installation of the drain and extraction well

cause a dip in concentrations The breakthrough curves exhibit behavior that is a combination of the other sets of alternatives concentrations rise briefly after the drain and well are removed but rapidly decrease due to source remediation Desorption still provides a decreasing but undecayed source At Woman Creek similar results are obtained however due to the longer travel distance and time the features of the breakthrough curves are more subdued Peak concentrations are simulated for PCE and DCE

B 8 0 SUMMARY

A groundwater flow and contaminant transport model has been developed and calibrated for OU 1 The model was used to simulate and predict contaminant movement from IHSS 119 1 to the french drain and Woman Creek

The model is considered to be conservative for the following reasons

- The model is two dimensional therefore dispersion (spreading) in lateral to the plane of the model is not simulated This causes over prediction of concentrations
- The model does not account for decay of contaminants adsorbed to soil If desorption occurs then concentrations are conservatively over predicted
- The model does not account for volatilization of organic contaminants It is likely that volatilization is an important process because of high volatilization rates for these chemicals (high Henry's constants) and because of the short distance from groundwater to landsurface
- The model predicts increasing concentrations at locations like Well 0487 and 4387 where observed concentrations fluctuate around a generally constant average This most likely due to the way in which desorption is simulated and to ignoring the effects of volatilization

The model is calibrated to average site conditions for flow and transport with adequate agreement between the model and observed conditions The model has good mass balance and exhibits monotonic convergence indicative of accurate calculations

Three scenarios were simulated, each representing a set of alternatives. Predicted results for no action alternatives indicate that concentrations at the french drain and at Woman Creek will increase to peak concentrations hundreds of years in the future. Predicted results for institutional-controls and remedial alternatives indicate that concentrations at the french drain and at Woman Creek will increase slightly then decrease with time. Peak concentrations are also hundreds of years in the future.

The results of the model are used in characterizing risk associated with each of the alternatives.

B 9 0 REFERENCES

Dames & Moore 1985 TARGET *Dames & Moore Mathematical Model of Ground Water Flow and Solute Transport*

**DOE 1994 *Final Phase III RFI/RI Rocky Flats Plant 881 Hillside Area (Operable Unit 1)*
U S Department of Energy Rocky Flats Plant Golden Colorado November 1993**

Fedors et al 1993a *Numerical Modeling of Variably Saturated Flow and Transport 881 Hillside* Rocky Flats Plant Jefferson County, Colorado Groundwater Technical Report Number 20 Colorado State University

Fedors et al 1993b, *Characterization of Physical and Hydraulic Properties of Surficial Materials and Groundwater/Surface Water Interaction Study at Rocky Flats Plant* Golden Colorado Colorado Groundwater Technical Report Number 21 Colorado State University

Table B-1a
Media Specific Hydraulic Parameters Used in all Contaminant
Simulations

Hydraulic Parameter	Units	Bedrock	Colluvium	Alluvium
Horizontal hydraulic conductivity	ft/d	0 06	0 45	6
Vertical hydraulic conductivity	ft/d	0 06	0 2	3
Specific storativity	1/ft	0 0001	0 00015	0 00035
Porosity		0 35	0 36	0 45
Bulk density ratio		1 81	1 5	1 65
Distribution coefficient	ft ³ /lb	0 0578	0 0578	0 0578
Molecular dispersion	ft ² /d	0 0001	0 0001	0 0001
Longitudinal dispersivity	ft	20	30	40
Transverse dispersivity	ft	2	10	10
Coefficient for Sr (psi)	1/ft	0 24	0 0558	3
Coefficient for Sr (psi)		1 09	1 22	2 5
Coefficient for Sr (psi)		-0 0826	-0 18	-0 6
Residual moisture content		0 25	0 19	0 1
Saturated moisture content		0 35	0 36	0 45
Coefficient for Kr (psi)	1/ft	0 83	0 0148	3 48
Coefficient for Kr (psi)		0 41	0 44	1 93
Coefficient for Kr (psi)		3	10	3
Minimum Kr (psi)		0 1	0 1	0 1

Table B-1b
Contaminant-Specific Modeling Parameters

Contaminant	Distribution Coefficient (ft ³ /lb)
1 1-dichloroethene	0 0104
Tetrachloroethene	0 0578
Carbon tetrachloride	0 0704
1 1 1 trichloroethane	0 0243
Selenium	2 4

Table B-2a
Measured Water Levels at Well 4387

Date	Measured Water Level (ft)
4/27/89	5917.2
5/18/89	5917.2
6/10/89	5917.3
6/29/89	5917.0
7/14/89	5916.8
8/18/89	5916.9
8/25/89	5915.2
9/12/89	5916.4
10/26/89	5916.8
1/16/90	5916.6
2/1/90	5916.6
4/13/90	5920.4
6/7/90	5920.2
7/12/90	5919.3
8/9/90	5918.9
9/11/90	5918.3
9/12/90	5915.7
10/1/90	5917.9
11/7/90	5917.7
11/13/90	5917.7
12/6/90	5917.4
1/3/91	5917.3
3/18/91	5917.0
4/1/91	5916.6
5/7/91	5917.2
5/13/91	5917.2
6/11/91	5917.5
7/5/91	5917.2
8/6/91	5916.8
8/14/91	5916.9
9/5/91	5916.6
10/3/91	5916.5
11/5/91	5916.3

**Table B-2a
(Continued)**

Date	Measured Water Level (ft)
12/2/91	5916.4
12/10/91	5916.4
1/3/92	5916.3
2/13/92	5916.5
3/5/92	5916.2
4/1/92	5918.5
5/5/92	5917.4
6/1/92	5917.7
6/23/92	5917.1
7/2/92	5917.2
8/3/92	5917.2
8/6/92	5917.3
9/4/92	5917.4
10/1/92	5916.6
10/27/92	5917.1
11/2/92	5916.8
12/3/92	5917.1
1/20/93	5916.7
2/2/93	5916.8
3/26/93	5917.8
4/2/93	5917.1
5/13/93	5917.3
6/17/93	5917.0
6/28/93	5916.9
7/13/93	5916.2
Maximum	5920.4
Minimum	5915.2
Average	5917.2

Table B-2b
Measured Water Levels at Well 0487

Date	Measured Water Level (ft)
4/27/89	5900.5
5/19/89	5901.0
6/9/89	5902.0
6/29/89	5901.7
7/14/89	5900.6
7/26/89	5900.3
8/18/89	5900.0
9/13/89	5899.7
10/16/89	5899.2
1/16/90	5899.2
1/31/90	5899.2
4/12/90	5905.7
6/7/90	5904.1
7/11/90	5902.1
8/8/90	5901.6
8/29/90	5900.8
9/12/90	5900.3
10/1/90	5899.9
10/29/90	5899.3
11/7/90	5899.3
12/6/90	5899.0
1/2/91	5898.7
3/18/91	5898.5
4/1/91	5898.5
5/7/91	5899.8
5/9/91	5899.8
6/5/91	5901.1
7/2/91	5900.5
8/6/91	5899.0
8/20/91	5898.8
9/3/91	5898.4
10/2/91	5897.5

**Table B-2b
(Continued)**

Date	Measured Water Level (ft)
11/5/91	5896.6
1/3/92	5897.1
2/3/92	5897.2
2/11/92	5897.2
3/5/92	5897.6
4/6/92	5901.8
5/6/92	5901.8
5/11/92	5901.7
6/1/92	5901.6
7/1/92	5899.7
8/3/92	5900.5
8/12/92	5900.2
9/4/92	5900.7
10/1/92	5900.0
10/21/92	5899.5
11/3/92	5899.2
12/7/92	5898.6
1/20/93	5898.4
2/2/93	5898.3
3/10/93	5898.0
3/26/93	5898.2
4/8/93	5901.4
5/14/93	5901.6
5/20/93	5901.6
6/16/93	5901.1
7/13/93	5900.5
Maximum	5905.7
Minimum	5896.6
Average	5899.9

Table B-2c
Measured Water Levels at Well 4787

Date	Measured Water Level (ft)
4/27/89	
5/19/89	5875.8
6/10/89	5877.3
6/29/89	
7/14/89	
7/26/89	
8/25/89	
9/13/89	5876.4
10/20/89	
1/16/90	5878.4
2/15/90	5875.9
4/12/90	5876.4
5/3/90	5876.5
7/11/90	
8/8/90	5875.1
9/11/90	
9/12/90	
10/1/90	
10/25/90	
11/7/90	
12/10/90	
1/2/91	
4/1/91	
5/7/91	
6/5/91	5877.2
7/2/91	5875.7
8/6/91	
8/19/91	
9/3/91	
10/2/91	
11/5/91	5875.0
12/10/91	5876.4

**Table B-2c
(Continued)**

Date	Measured Water Level (ft)
1/10/92	5875.1
2/5/92	
2/11/92	
3/5/92	5875.0
4/6/92	
5/5/92	5879.2
6/10/92	5878.0
7/1/92	5877.2
8/5/92	5876.5
8/17/92	5876.2
9/4/92	5877.0
10/1/92	5876.2
10/21/92	5875.8
11/3/92	5875.2
12/7/92	
1/20/93	5875.2
2/2/93	5875.4
3/26/93	5875.4
4/2/93	5875.5
6/16/93	5876.2
7/2/93	5875.9
Maximum	5879 2
Minimum	5875 0
Average	5876 2

Table B-2d
Measured Water Levels at Well 5587

Date	Measured Water Level (ft)
4/27/89	5850.7
5/19/89	
6/1/89	5850.7
6/29/89	
7/10/89	5850.7
7/28/89	
8/25/89	
9/14/89	
10/16/89	
1/16/90	
4/12/90	5853.5
5/4/90	5853.2
7/10/90	5852.2
7/19/90	5854.1
8/7/90	5851.2
9/12/90	5851.3
10/1/90	5851.3
10/29/90	5851.2
11/7/90	5850.7
12/6/90	
1/2/91	5850.7
3/18/91	5850.7
4/1/91	
5/7/91	5850.7
6/5/91	5850.7
7/2/91	5851.0
8/6/91	5851.1
8/19/91	5851.1
9/3/91	
10/2/91	5850.7
11/5/91	5850.6
11/14/91	5850.6

**Table B-2d
(Continued)**

Date	Measured Water Level (ft)
12/2/91	5850.6
1/3/92	
2/3/92	
3/5/92	5850.6
4/1/92	5853.6
5/1/92	5852.9
5/7/92	5852.9
6/1/92	5851.7
7/1/92	5851.9
8/3/92	5851.8
8/17/92	5851.7
9/4/92	5851.0
10/1/92	5851.0
10/20/92	5851.1
11/3/92	5850.7
12/7/92	5850.7
1/19/93	5850.7
2/1/93	5850.7
3/4/93	5850.6
3/29/93	5850.7
4/7/93	5850.7
5/14/93	5850.9
5/18/93	5850.9
6/16/93	5850.9
7/6/93	5851.0
Maximum	5854 1
Minimum	5850 6
Average	5851 3

Table B-3a
Measured Concentrations of VOCs at Well 4387

Date	Material Type	Tetra chloroethene (ug/l)	1 1 Dichloroethene (ug/l)	1 1 1 Trichloroethane (ug/l)	Carbon Tetrachloride (ug/l)
Feb-01 1990	Colluvium	3600	4900	500 U	500 U
Jun-06-1990	Colluvium	61	38	110	5 U
Jun-07 1990	Colluvium	82	53	140	5 U
Sep-11 1990	Colluvium	1400 E	2400	3200 E	400 E
Nov 14-1990	Colluvium	1500 B V	1400 V	3000	120 U V
Mar-19 1991	Colluvium	2900	2900	5900 B	170 U
May 15 1991	Colluvium	6000 D JA	8200 D V	15000 D	5 U
Aug 15 1991	Colluvium	5700 V	7200 V	10000	500 U V
Dec-12 1991	Colluvium	3400 V	6000 D V	14000 D	100 U V
Feb-18-1992	Colluvium	3200 JA	4300 D V	7400 D	5 U V
Jun-24-1992	Colluvium	1400 V	1400 V	2600	40 U V
	Maximum	6000	8200	15000	500
	Minimum	61	38	110	5
	Average	3000	4750	5750	153

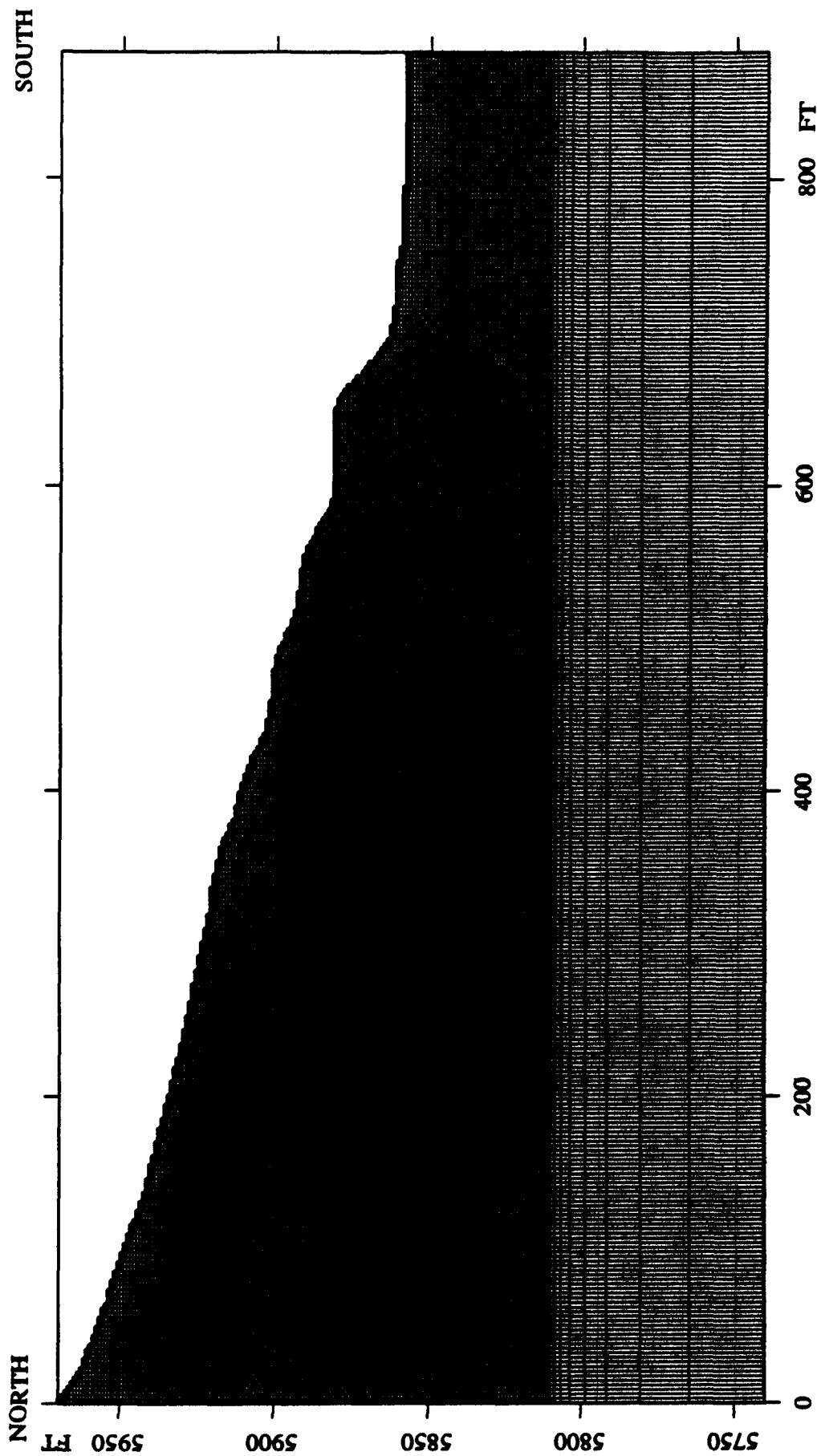
V - Valid
 R - Rejected
 B - Found in blank
 E - Value > calibration range
 JA - Estimated acceptable
 U - Not detected at/above method detection limit
 D - Dilution
 J - Estimated value

Table B-3b
Measured Concentrations of VOCs at Well 0487

Date	Material	Tetra chloroethene (ug/l)	1 1 Dichloroethene (ug/l)	1 1 1 Trichloroethane (ug/l)	Carbon Tetrachloride (ug/l)
Jan-31 1990	Colluvium	27	4 J	2 J	160
Jun-07 1990	Colluvium	14	4 J	7	55
Aug 29 1990	Colluvium	21	5	8	95
Oct 30-1990	Colluvium	5	5 U	5 U V	5 U V
Mar-19-1991	Colluvium	28	25 U	25 U	130
May 10-1991	Colluvium	46	7	5 U V	330 D V
Aug 21 1991	Colluvium	73	14	7 V	280 D V
Feb-11 1992	Colluvium	55	9 2	5	240
May 11 1992	Colluvium	10	20 U	10 U	120
	Maximum	73	25	25	330
	Minimum	5	4	2	5
	Average	41 3	8 7	5 7	205

V = Valid
 R = Rejected
 B = Found in blank
 E = Value > calibration range
 JA = Estimated, acceptable
 U = Not detected at/above method detection limit
 D = Dilution
 J = Estimated value

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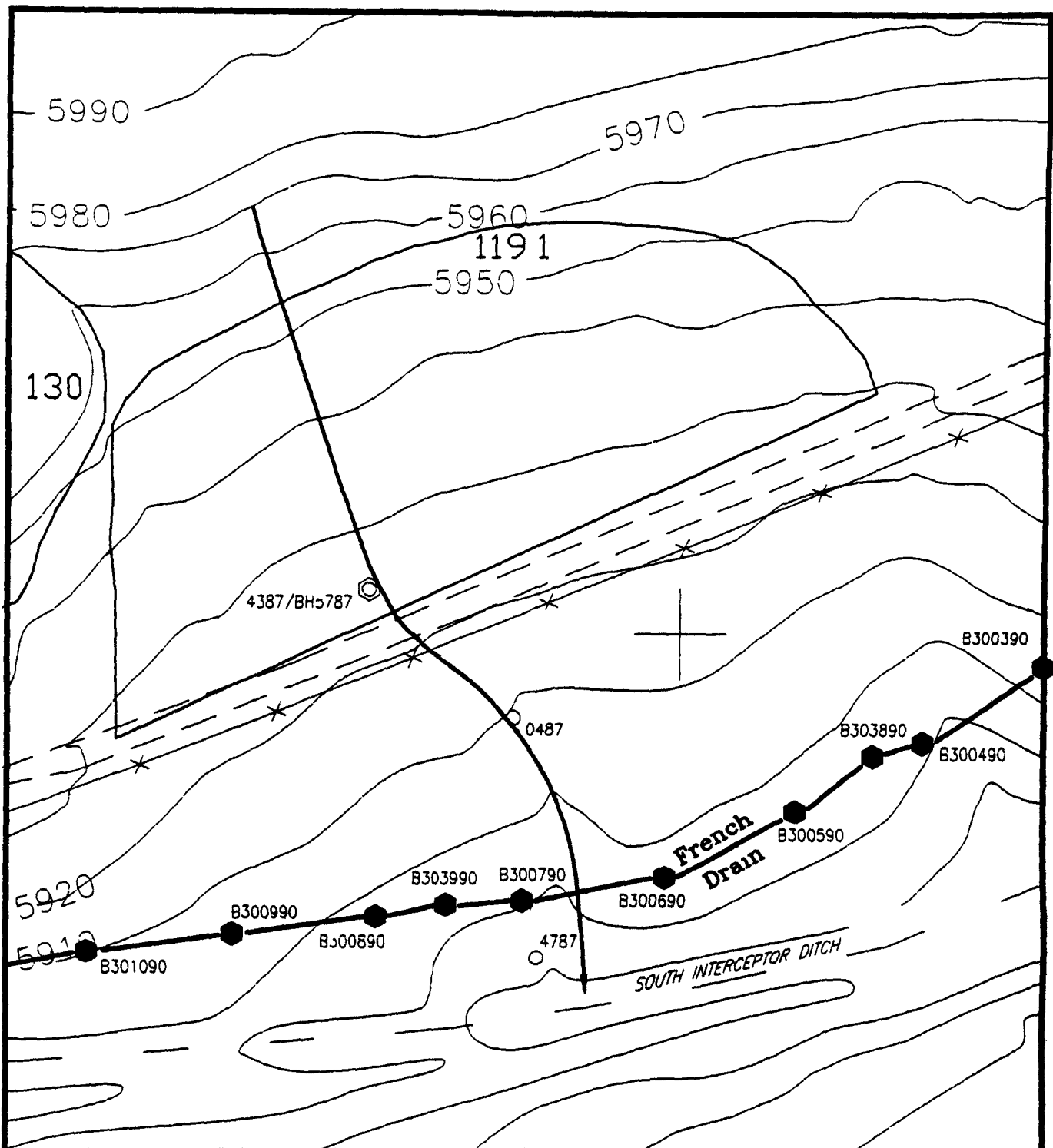
U S DEPARTMENT OF ENERGY
Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

MODEL DISCRETIZATION

Figure B 1

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EXPLANATION



INDIVIDUAL HAZARDOUS SUBSTANCE

- B301889
- BH1587
- B300390

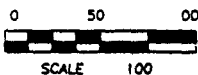
ALLUVIAL WELL

BOREHOLE

FRENCH DRAIN BOREHOLES



CONTOUR INTERVAL = 10 FEET

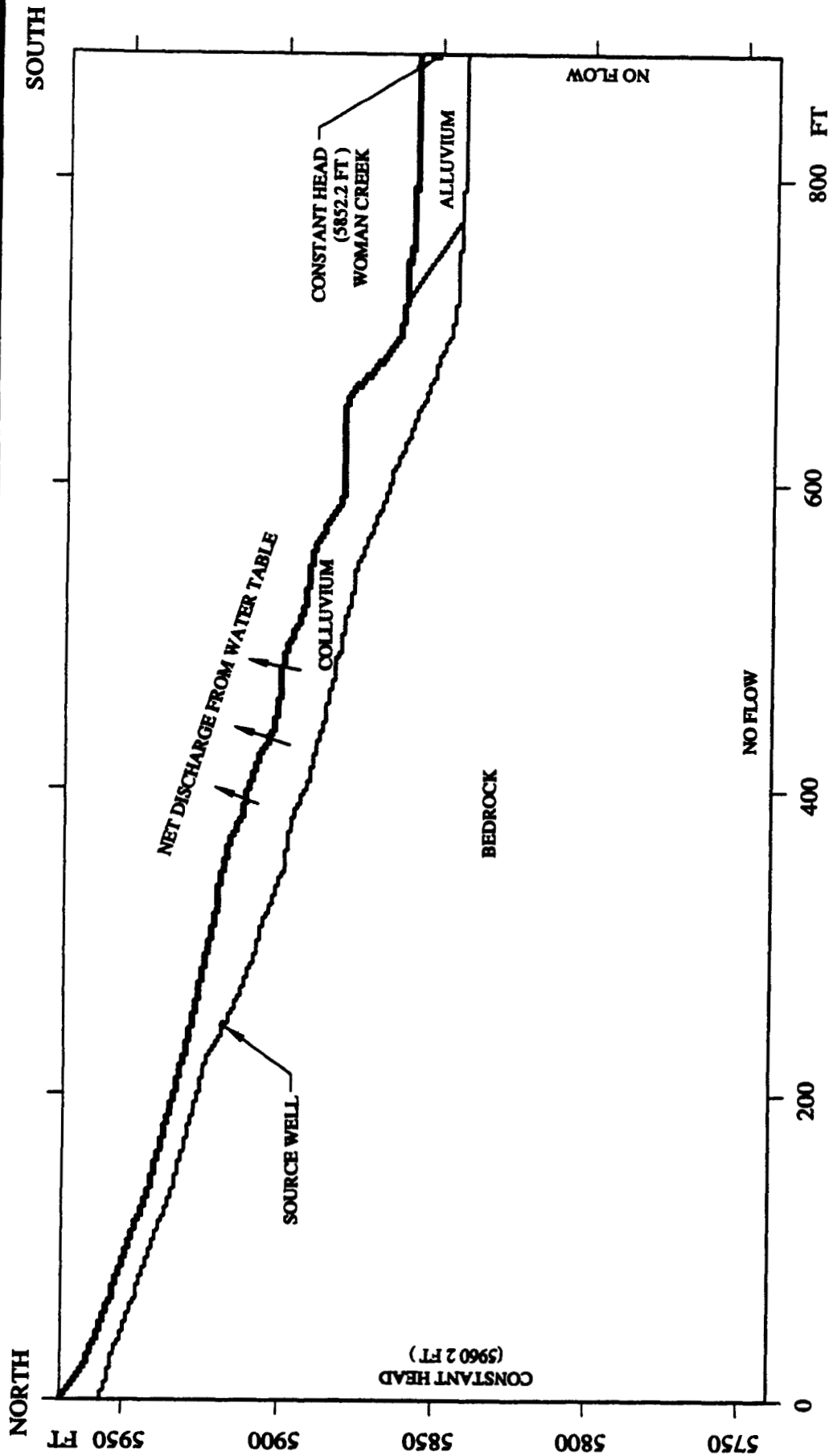


U S DEPARTMENT OF ENERGY Rocky Flats Plant, Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1
PHASE III RFI/RI REPORT

Location of Model Section
in IHSS 119 1

Figure B-3



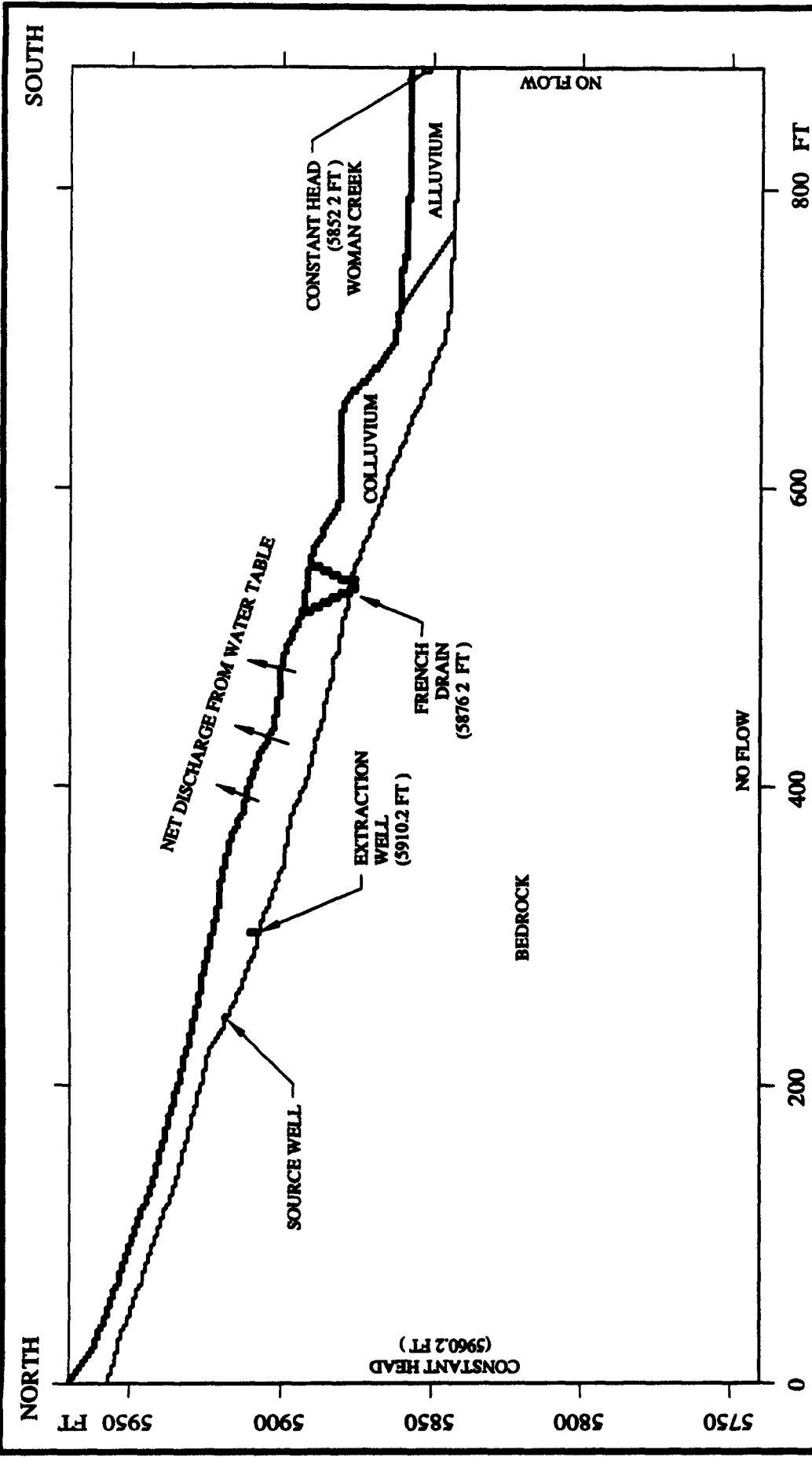
U S DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

MODEL BOUNDARY
CONDITIONS
1992

Figure B-4



U S DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

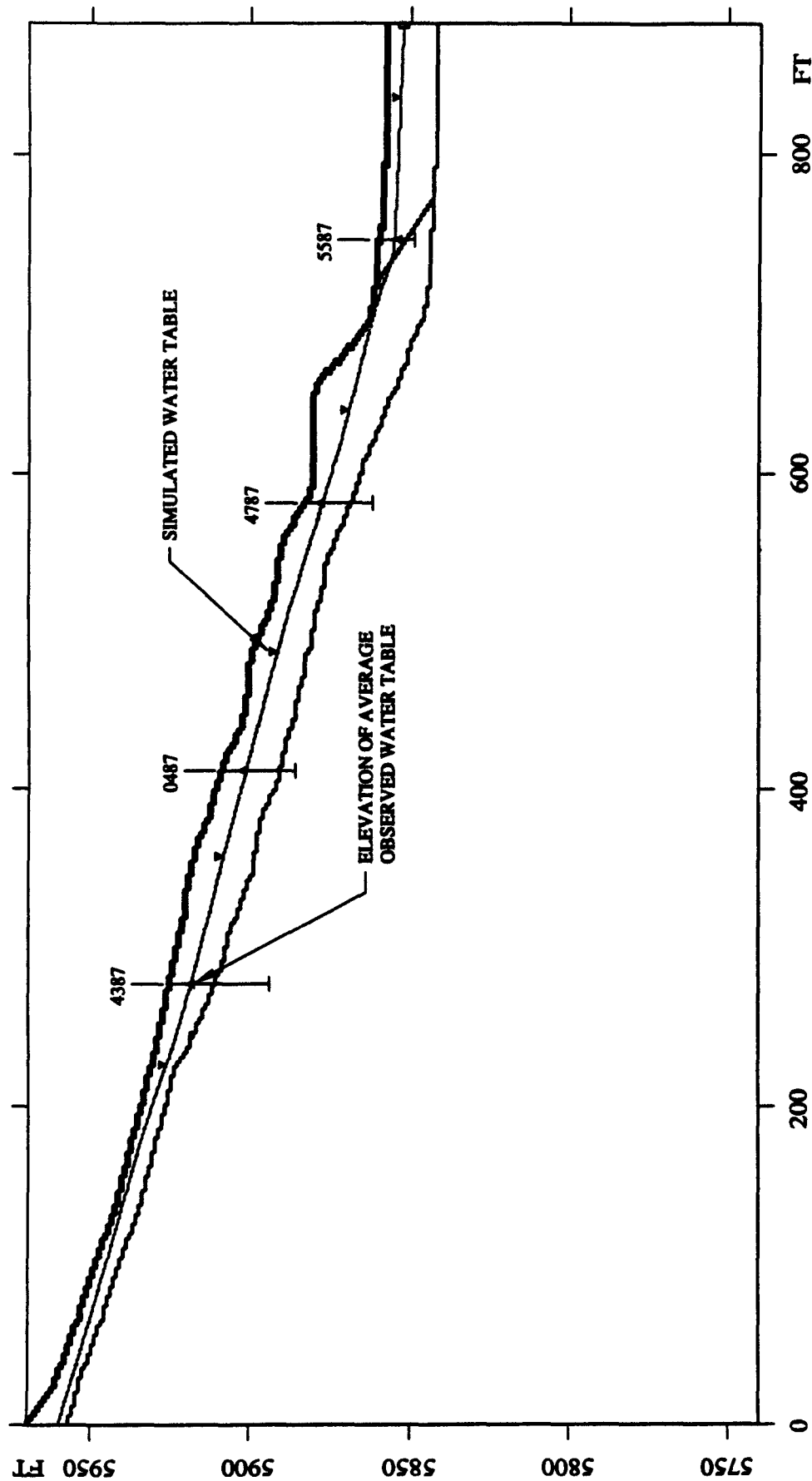
MODEL BOUNDARY
CONDITIONS
1994

Figure B 5

K:\DESIGN\ FLTSHILL\881\OU-1\GRAPH\SWATERBL.DWG

NORTH

SOUTH



U S DEPARTMENT OF ENERGY

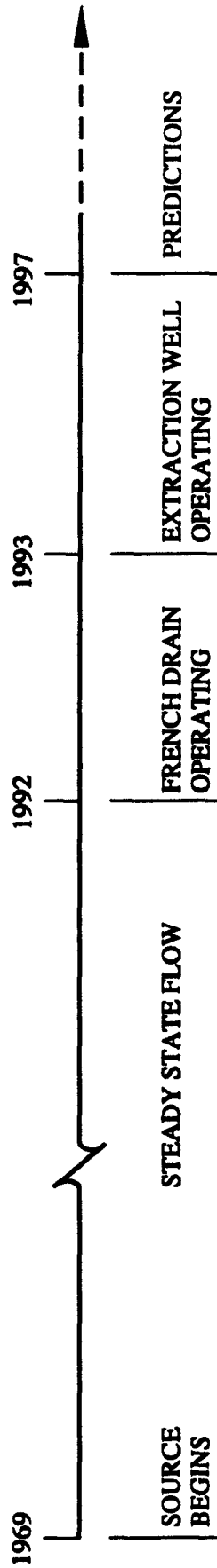
Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO. 1

WATER TABLE
SIMULATED VS OBSERVED

Figure B-6

TIME →



U S DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1
MAJOR TIME EVENTS
IN OU1 MODEL

Figure B 7

1,1 DCE Calibration of well 4387 Simulated Vs Observed

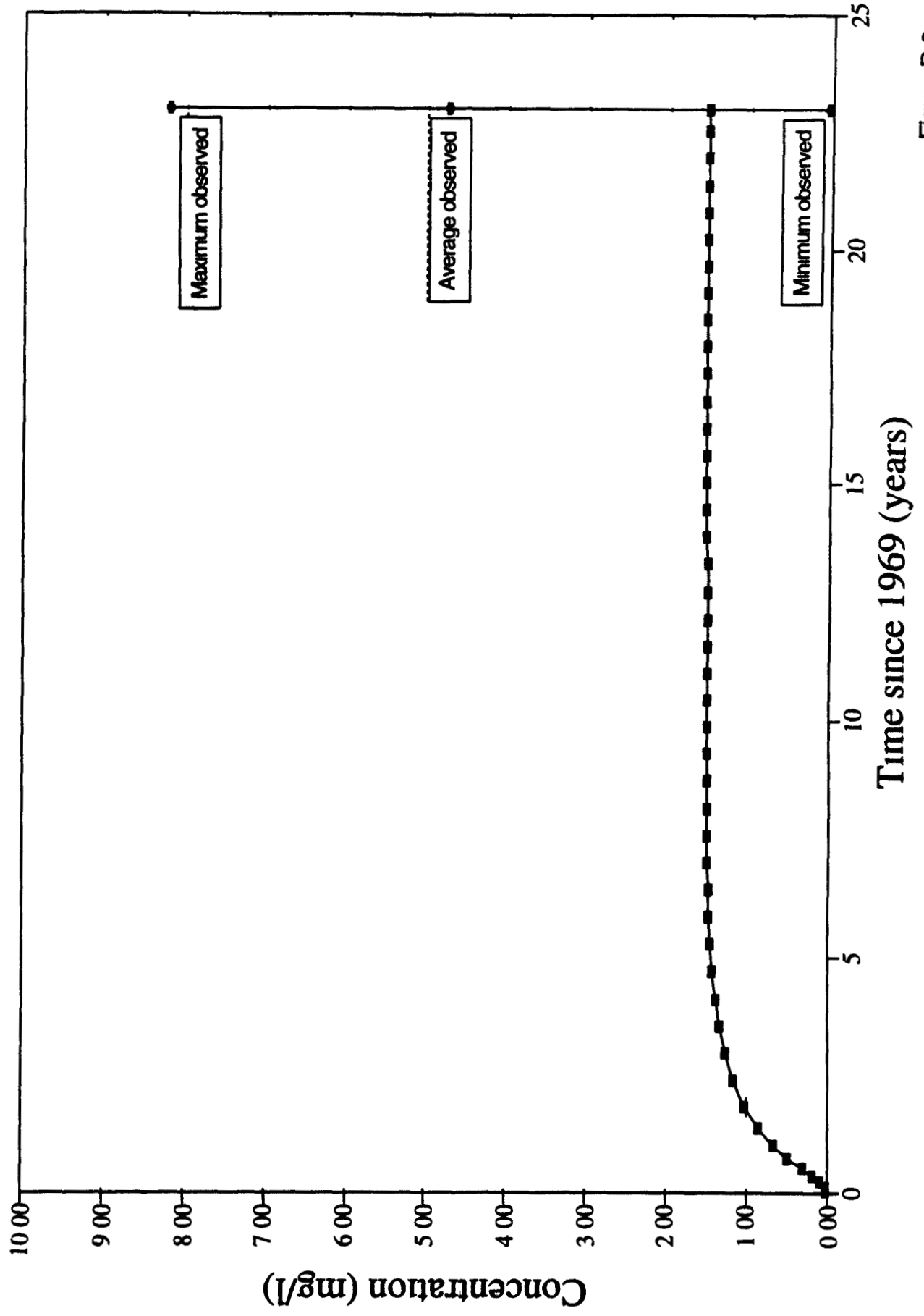


Figure B 8

1,1 DCE Calibration of well 0487 Simulated Vs Observed

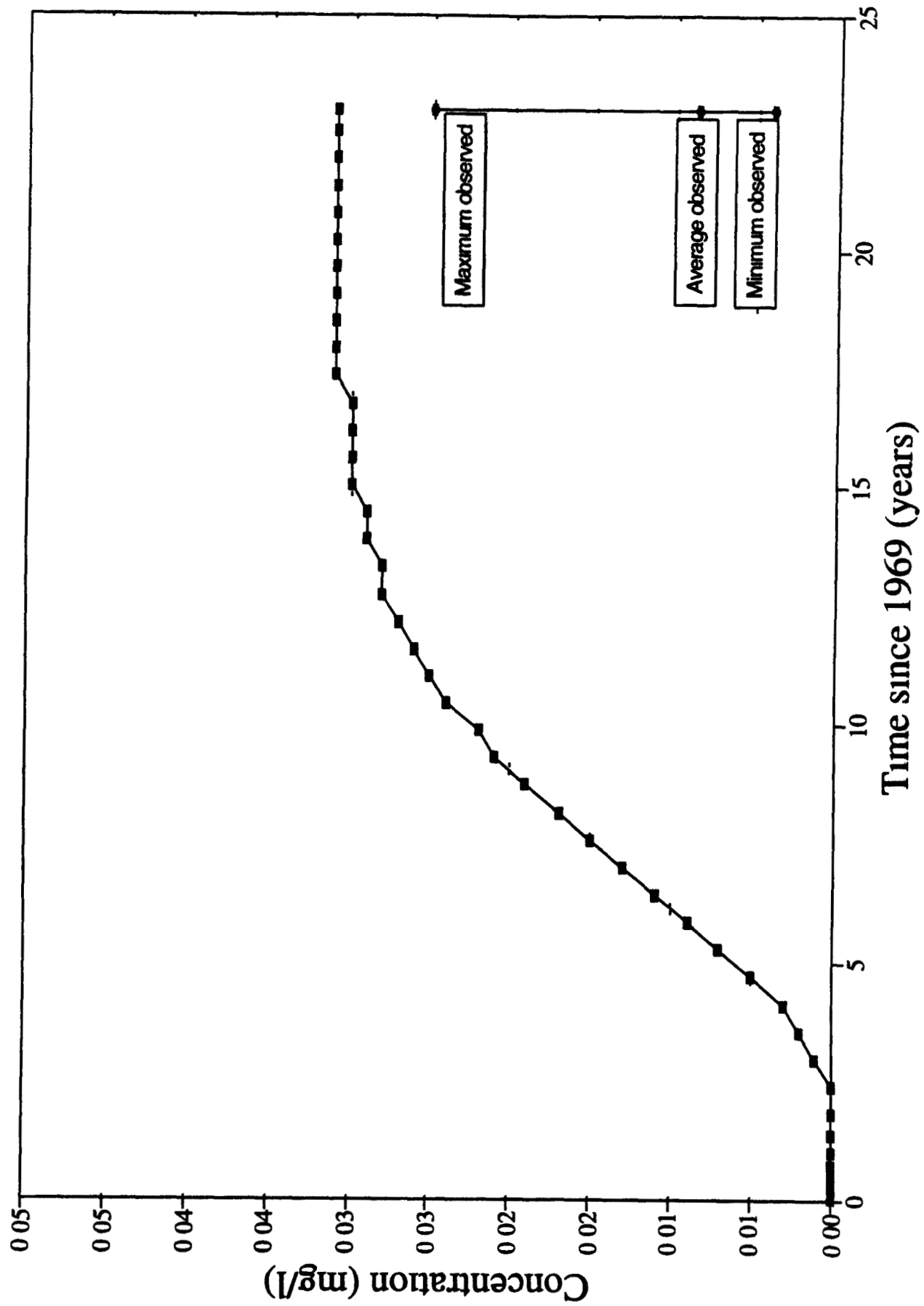


Figure B-9

PCE Calibration of well 4387 Simulated Vs Observed

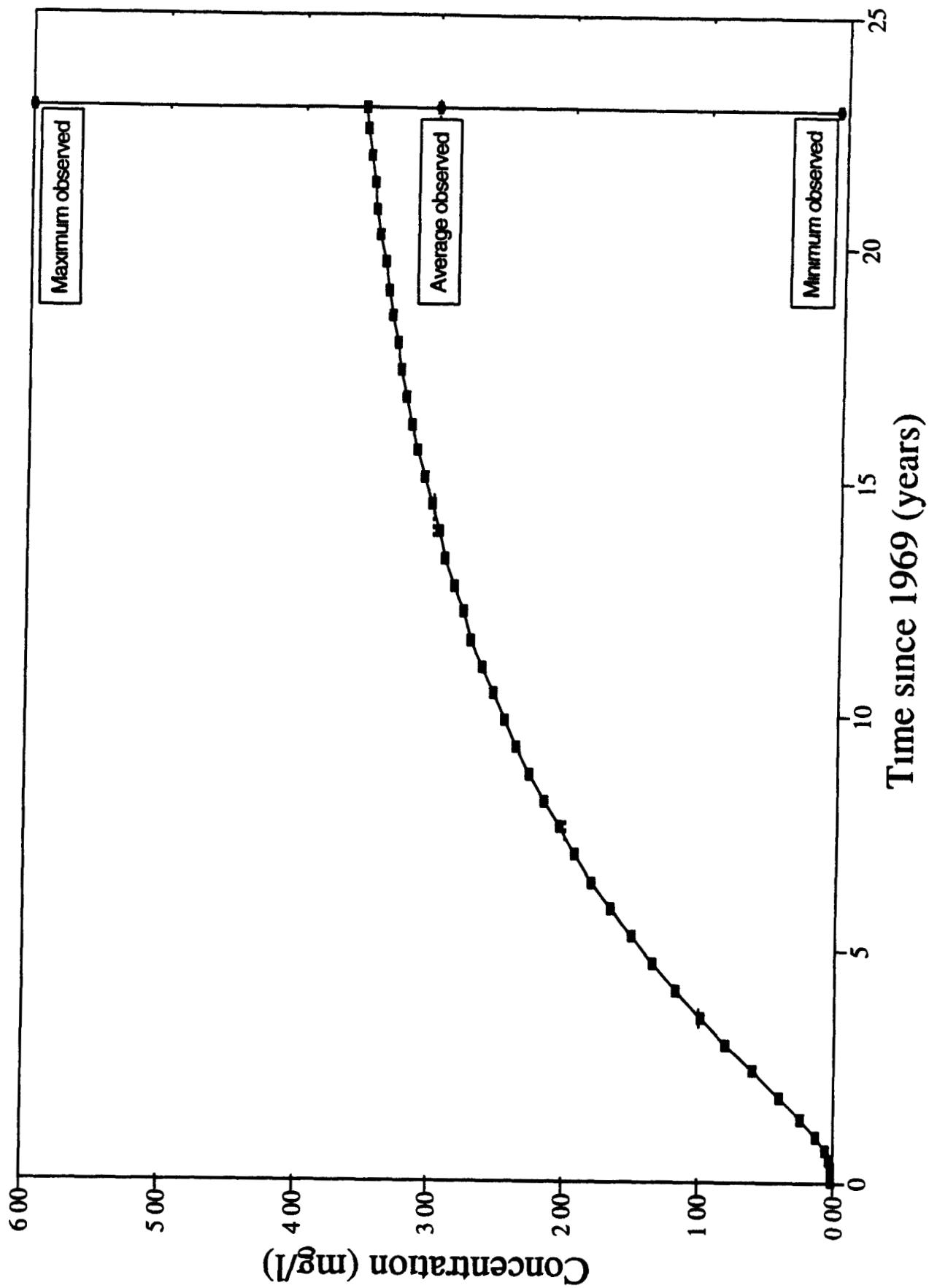


Figure B 10

PCE Calibration of well 0487 Simulated Vs Observed

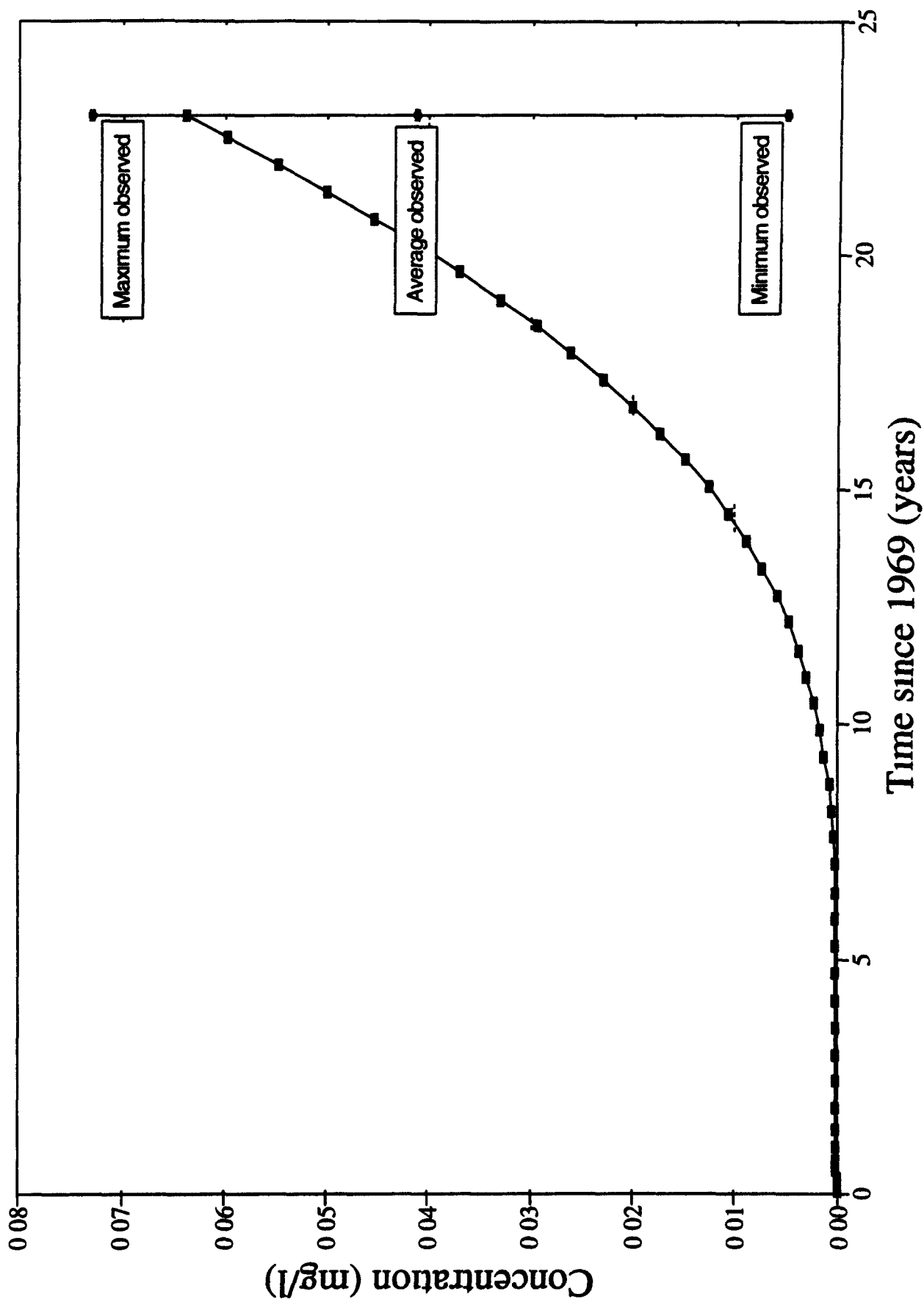


Figure B-11

CCL Calibration of well 4387 Simulated Vs Observed

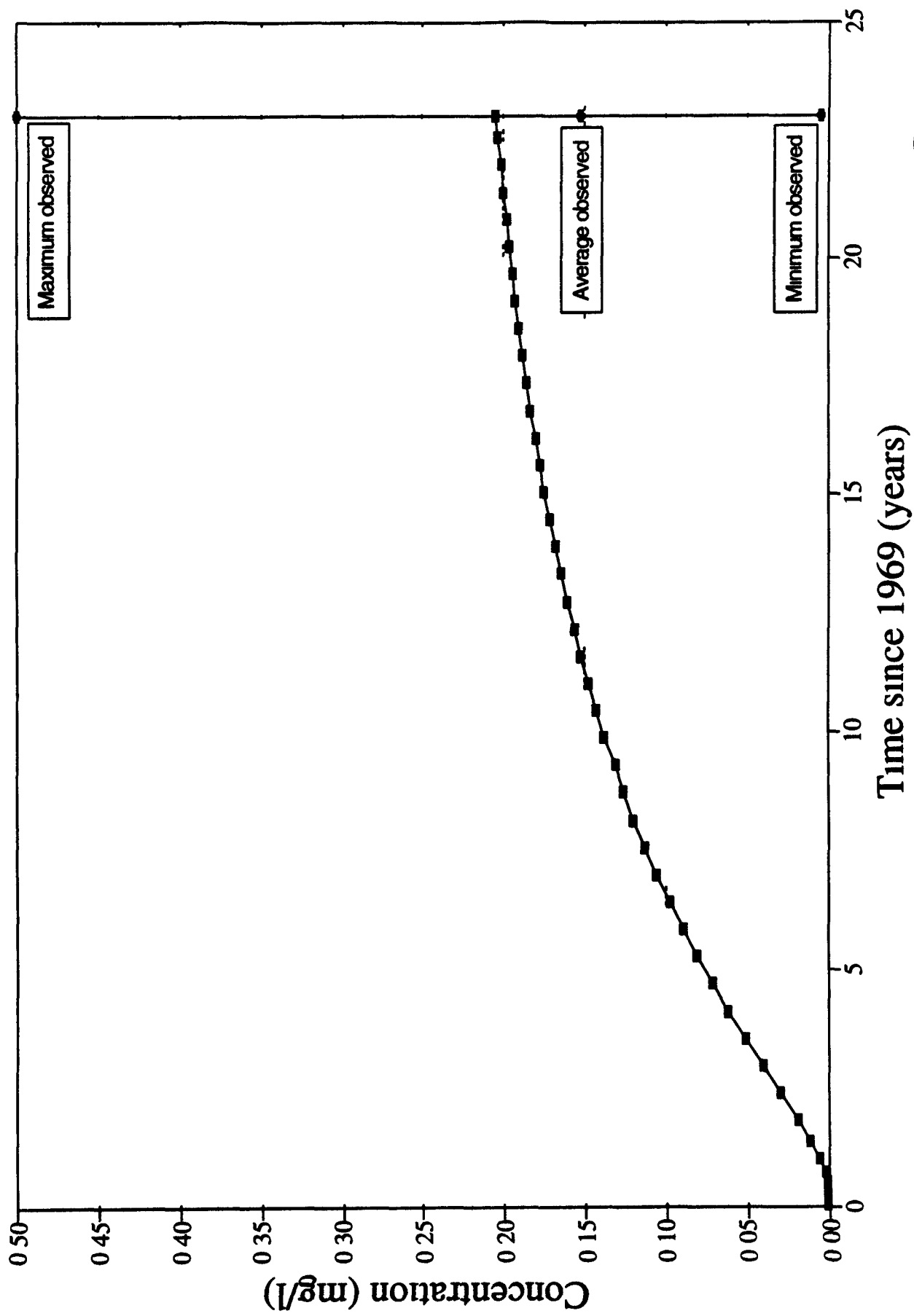


Figure B 12

CCL Calibration of well 0487 Simulated Vs Observed

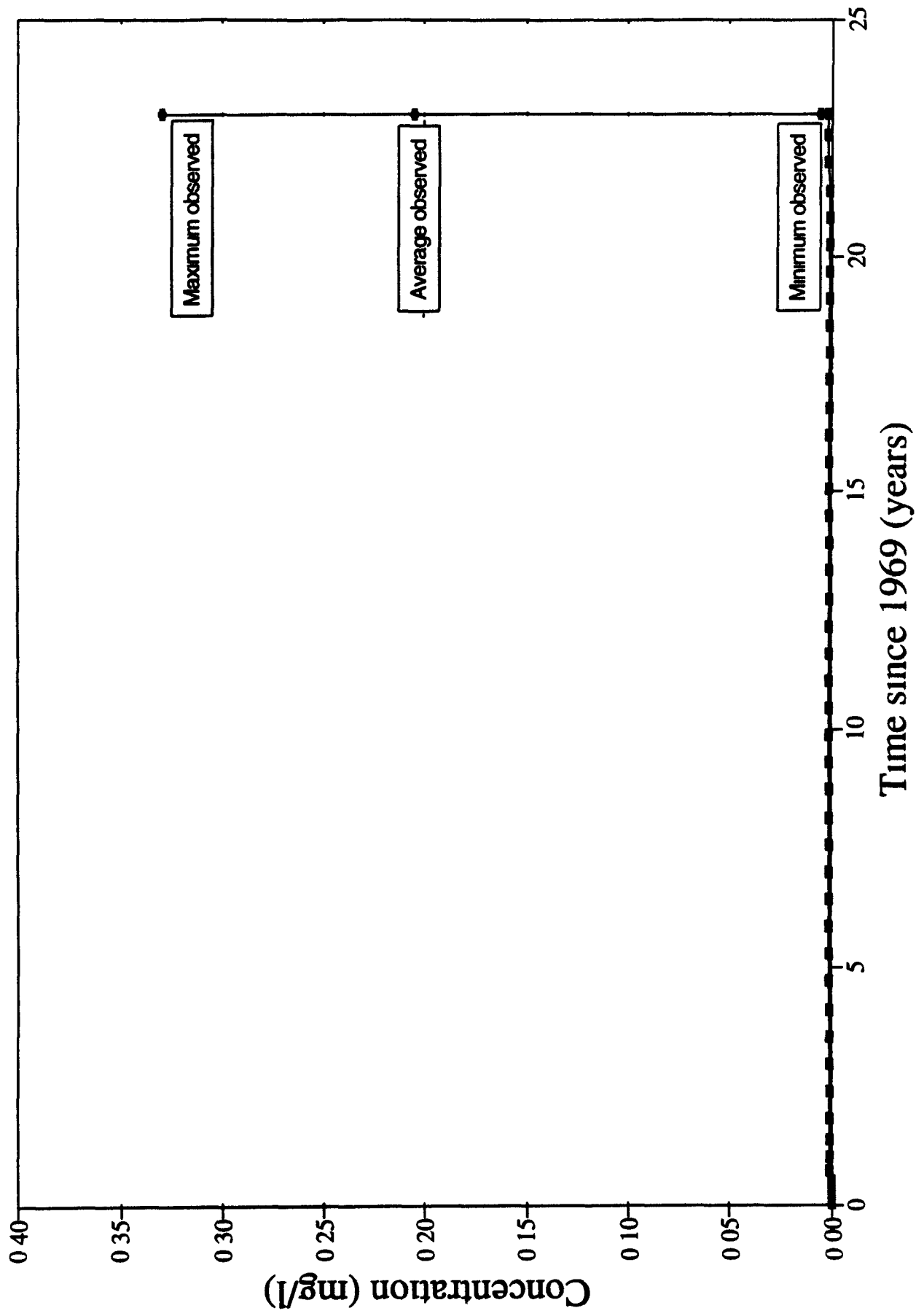


Figure B 13

1,1,1 TCA Calibration of well 4387 Simulated Vs Observed

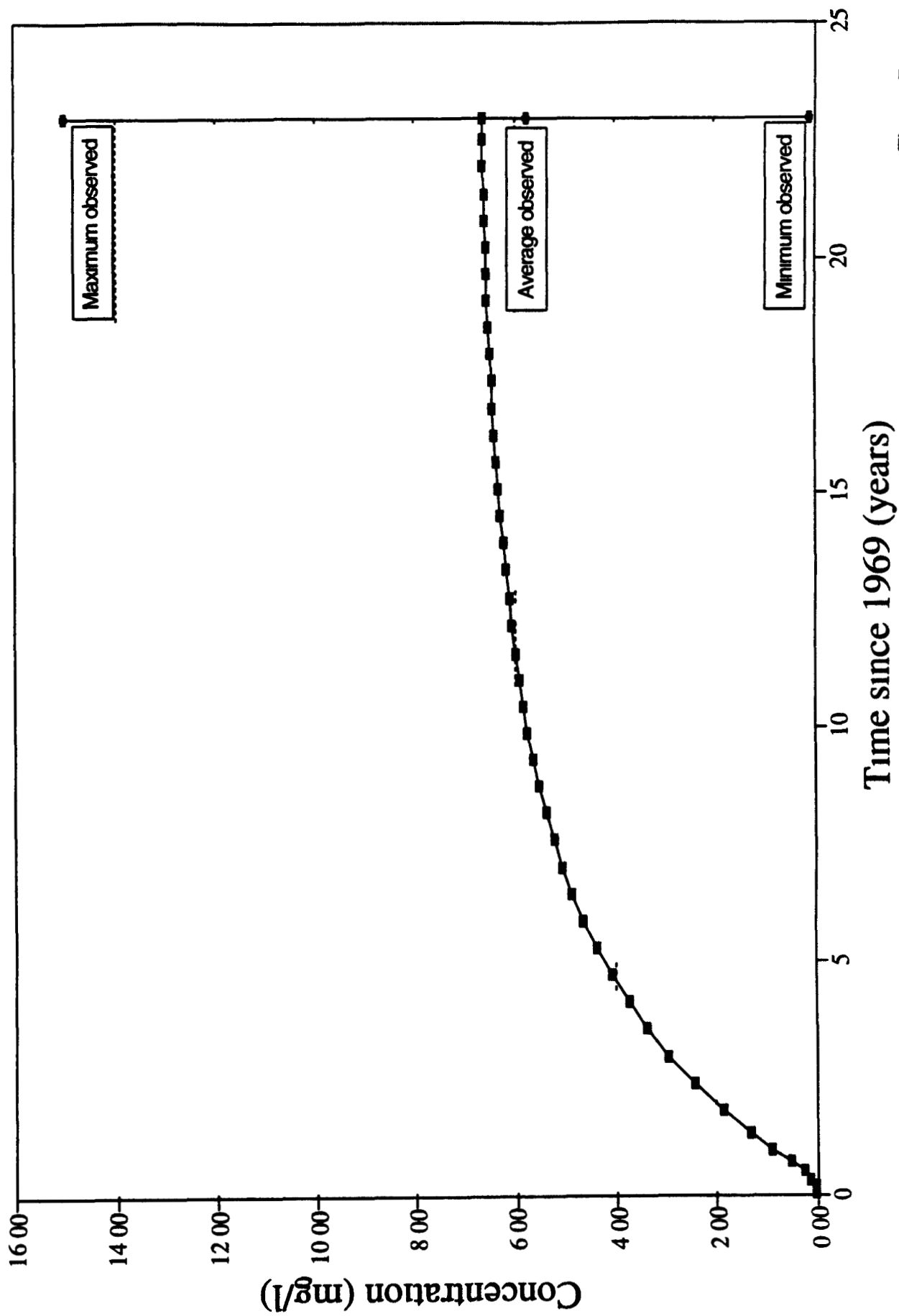


Figure B 14

1,1,1 TCA Calibration of well 0487 Simulated Vs Observed

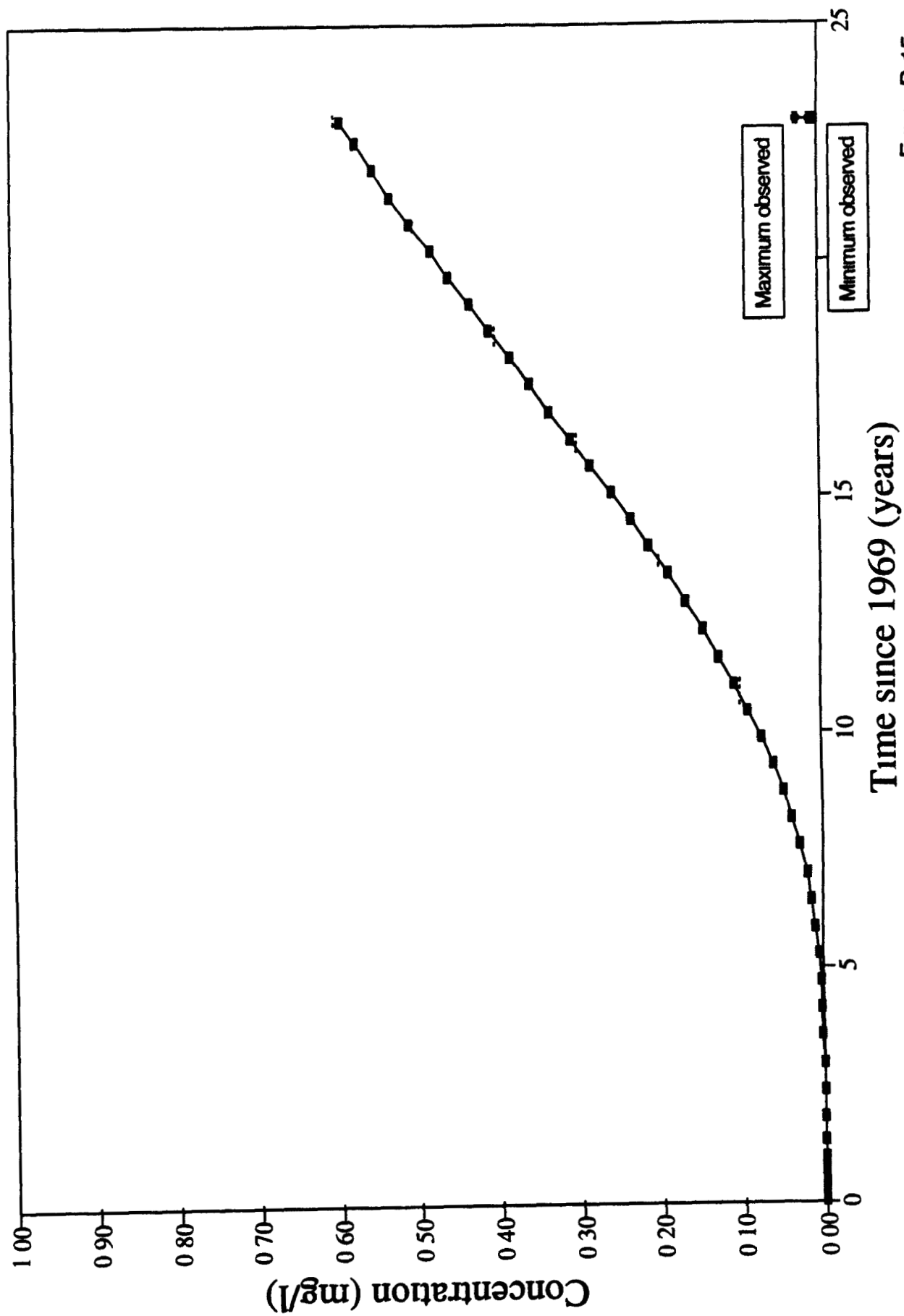


Figure B 15

Selenium Calibration of well 4387 Simulated Vs Observed

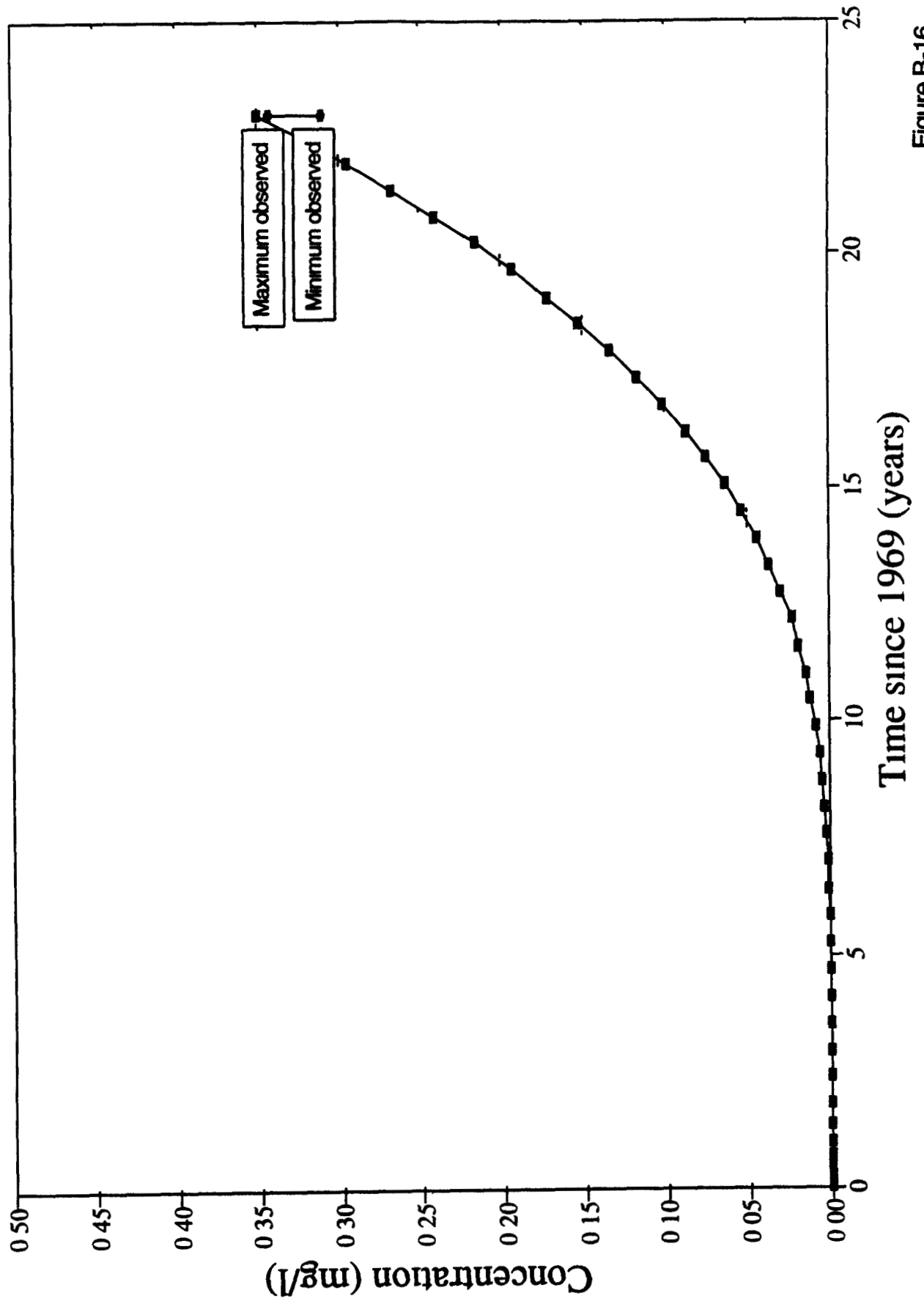


Figure B-16

Selenium Calibration of well 0487 Simulated Vs Observed

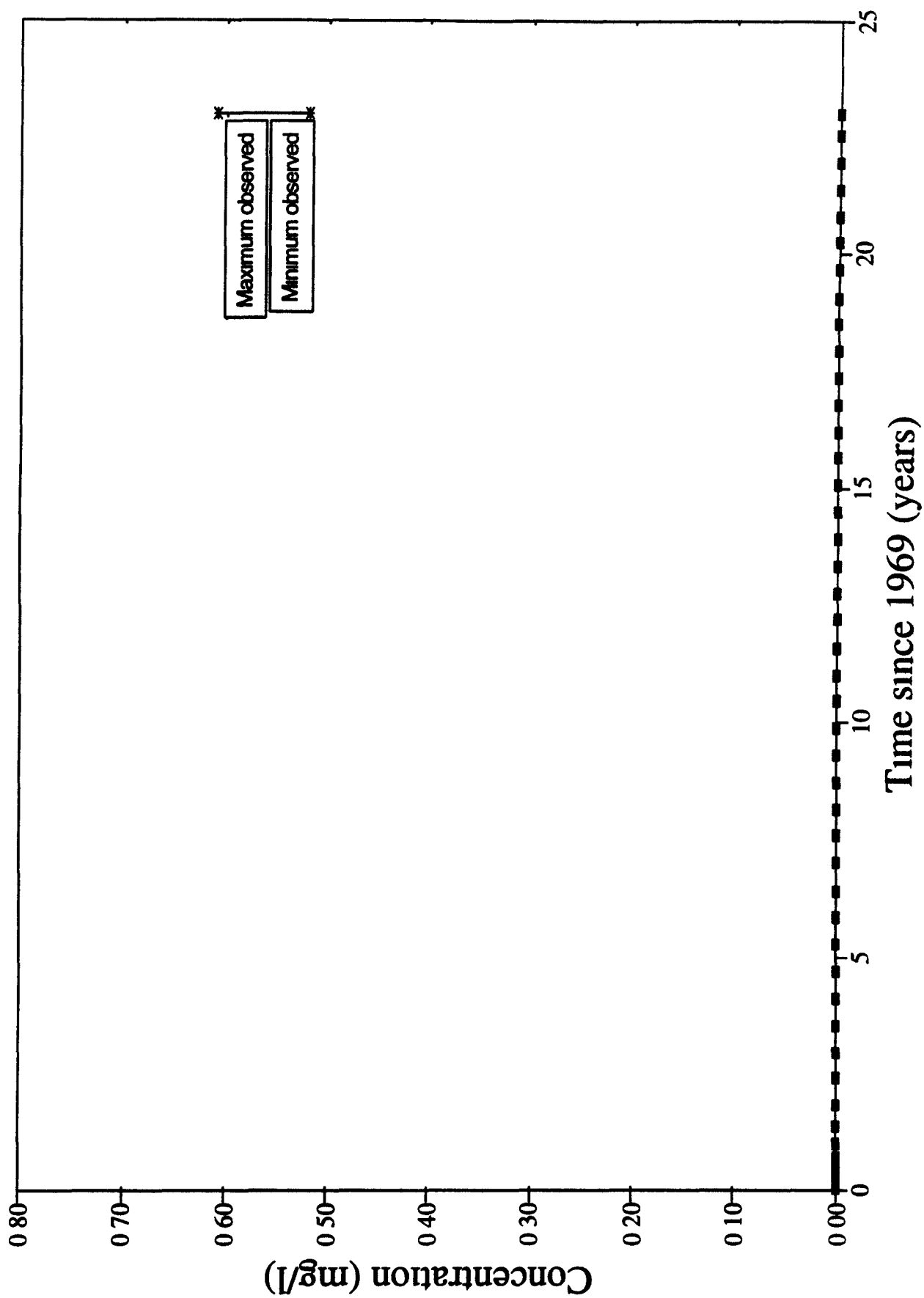
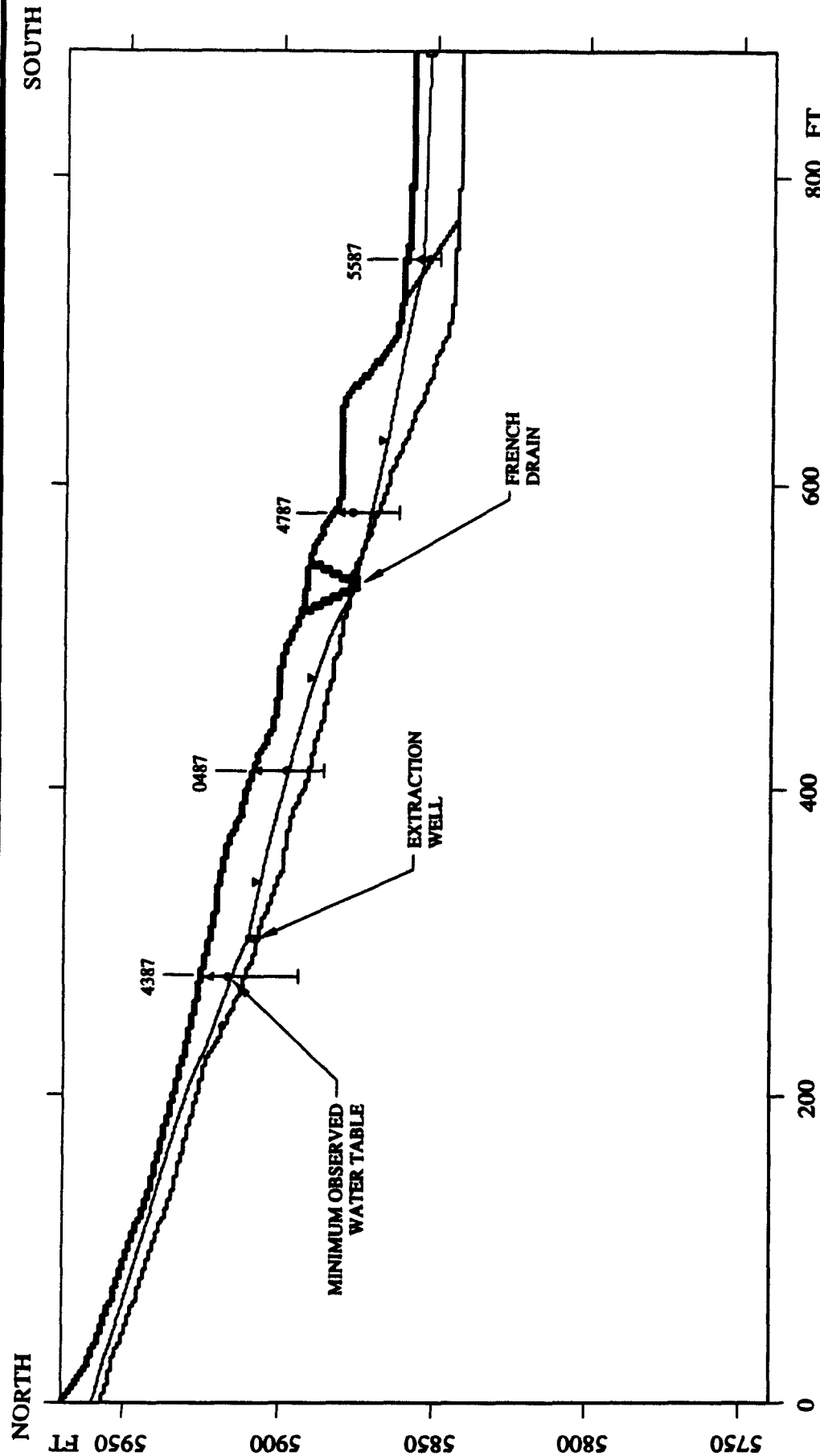


Figure B 17



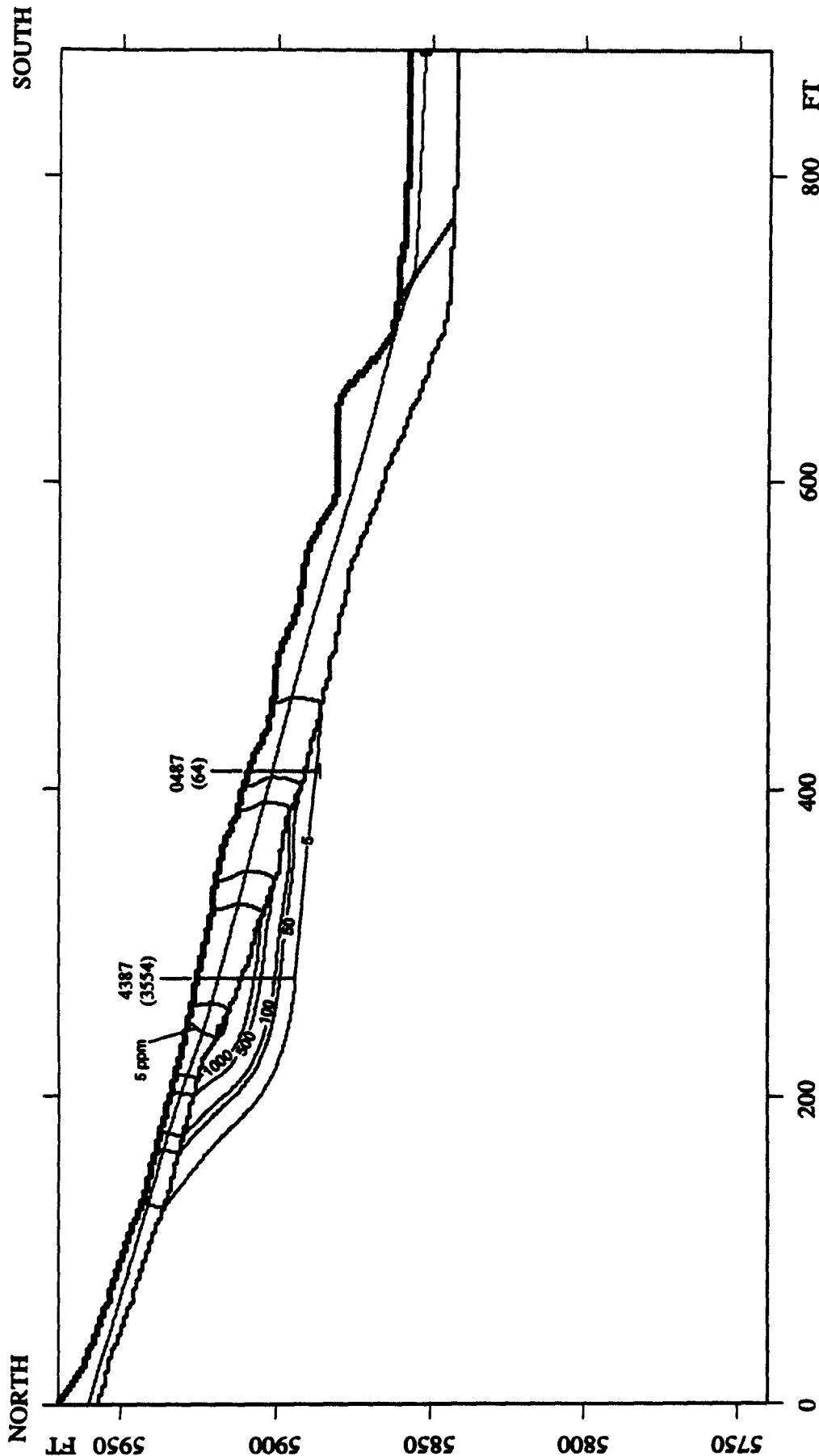
U S DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

WATER TABLE
SIMULATED VS OBSERVED
TIME = 9125 DAYS (1994)

Figure B-18



U S DEPARTMENT OF ENERGY
Rocky Flats Environmental Technology Site
Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

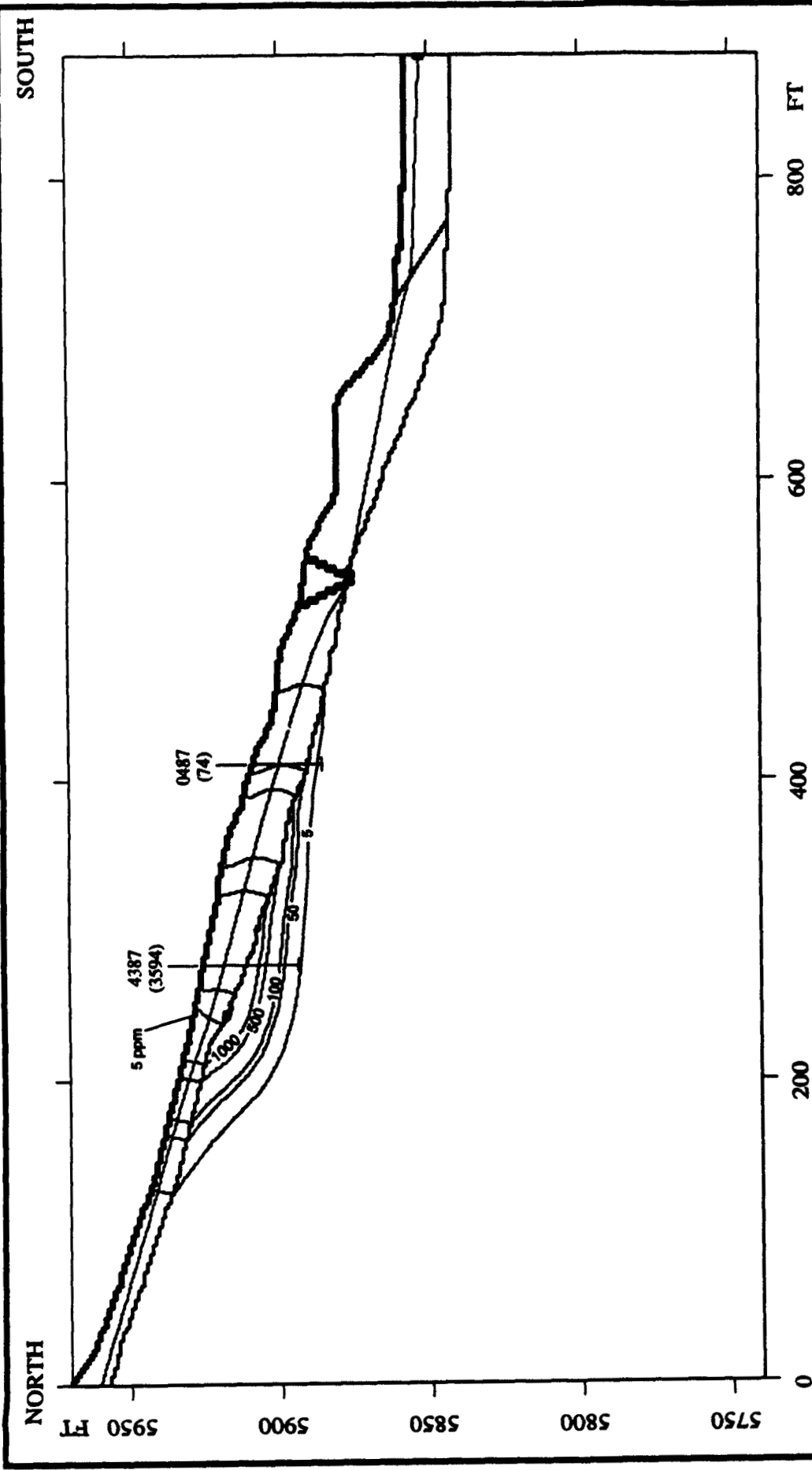
CONCENTRATION CONTOURS
OF PCE
TIME = 8395 DAYS (1992)

Figure B-19

EXPLANATION

WELL NUMBER	AVERAGE PCE CONCENTRATION IN ppb
4387 (3600)	

NOTE Concentrations in ppb unless otherwise labeled.



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Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO. 1

CONCENTRATION CONTOURS
OF PCE

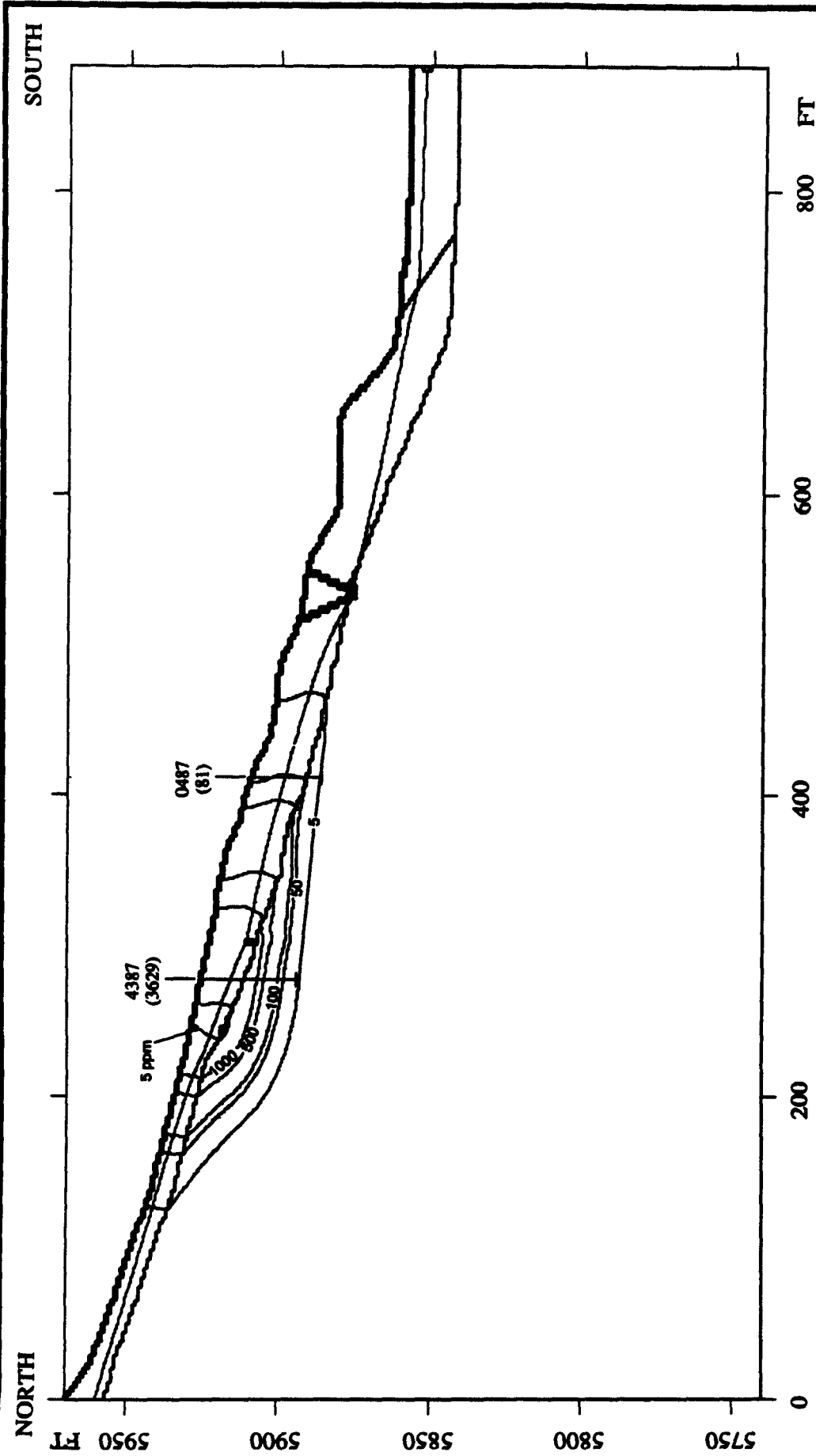
TIME = 8760 DAYS (1993)

Figure B-20

EXPLANATION

4387 (3600)	WELL NUMBER AVERAGE PCE CONCENTRATION IN ppb
----------------	---

NOTE Concentrations in ppb unless otherwise labeled.



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Rocky Flats Environmental Technology Site
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881 HILLSIDE AREA
OPERABLE UNIT NO. 1

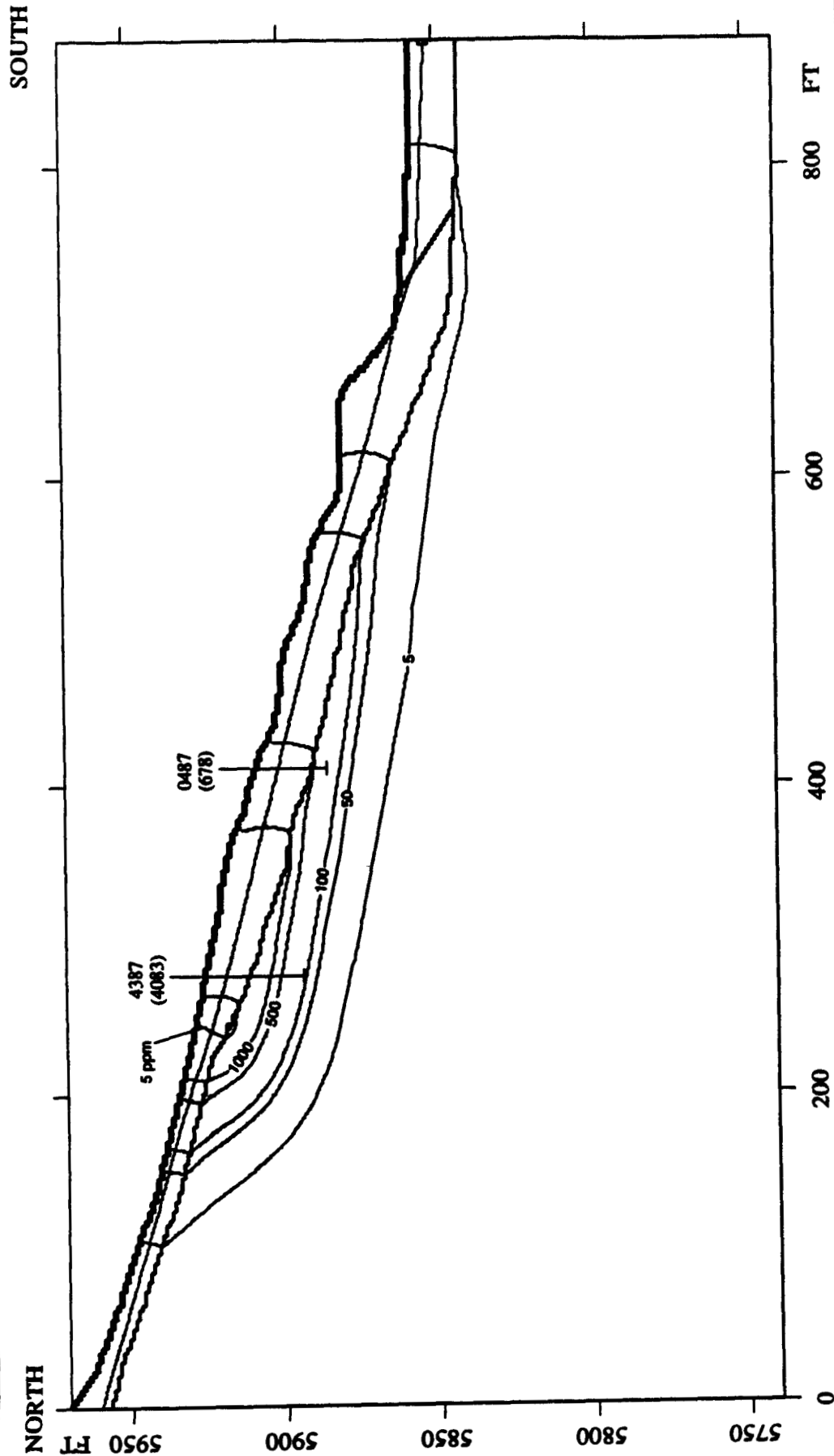
CONCENTRATION CONTOURS
OF PCE
TIME = 9125 DAYS (1994)

Figure B-21

EXPLANATION

WELL NUMBER	AVERAGE PCE CONCENTRATION IN ppb
4387	
(3600)	

NOTE: Concentrations in ppb unless otherwise labeled



U S DEPARTMENT OF ENERGY
 Rocky Flats Environmental Technology Site
 Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1
CONCENTRATION CONTOURS
OF PCE
TIME = 146000 DAYS (2369)

Figure B 22

EXPLANATION

4387 WELL NUMBER
 (3600) AVERAGE PCE CONCENTRATION IN ppb

NOTE Concentrations in ppb unless otherwise labeled

Simulated concentration of 1,1 DCE down gradient of the French drain

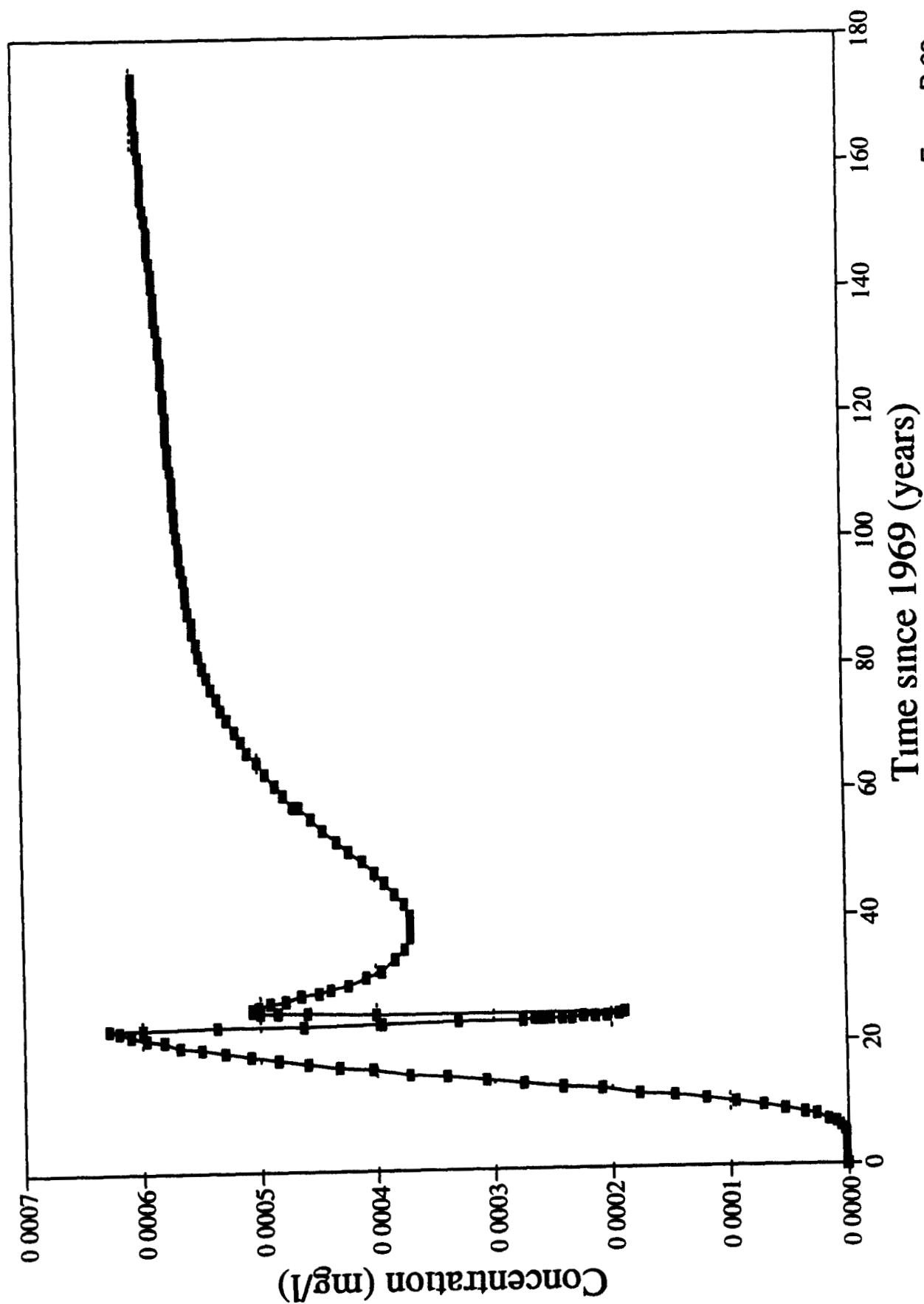


Figure B 23

Simulated concentration of 1,1 DCE at Woman creek

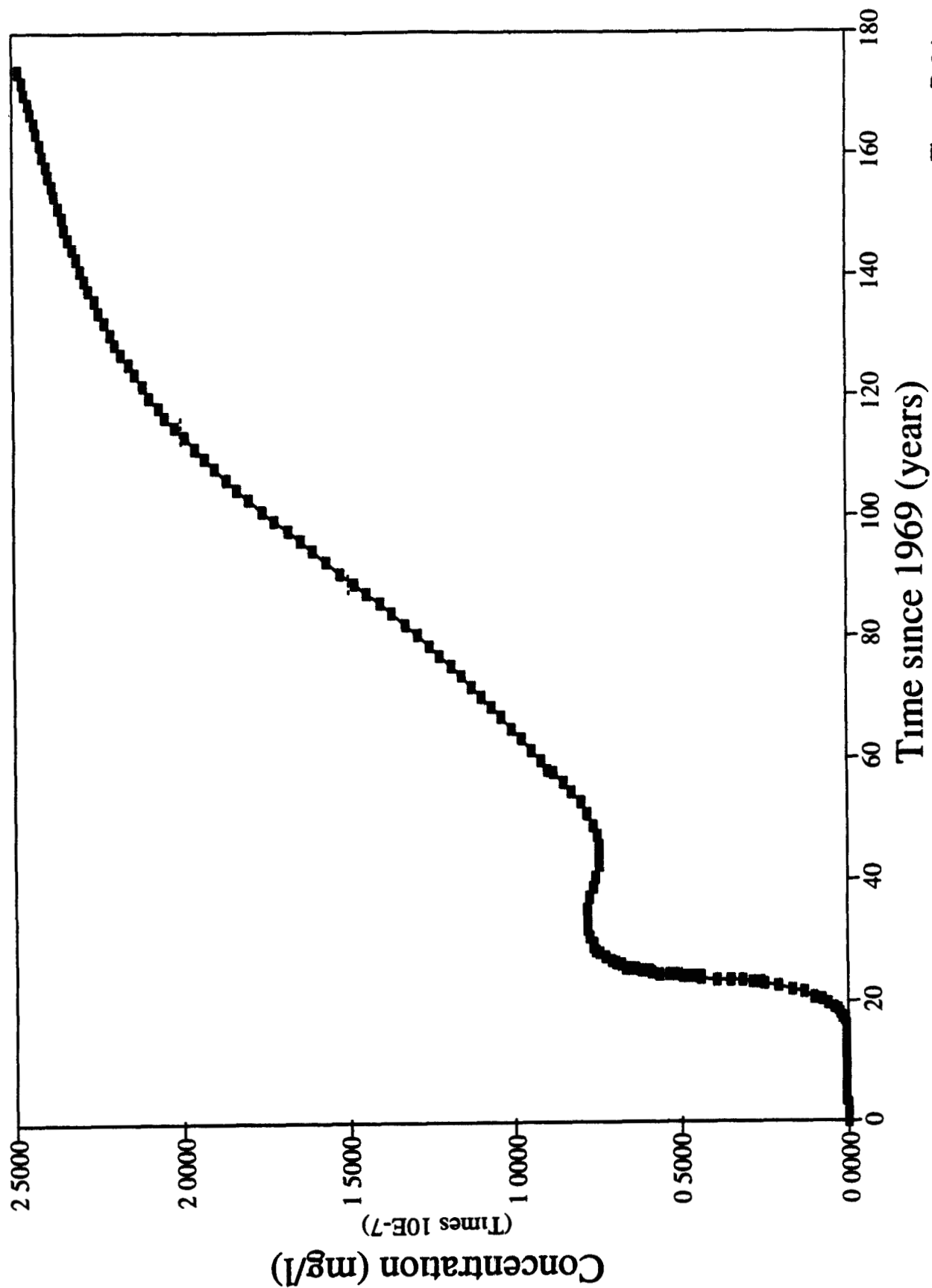


Figure B-24

Simulated concentration of PCE down gradient of the French drain

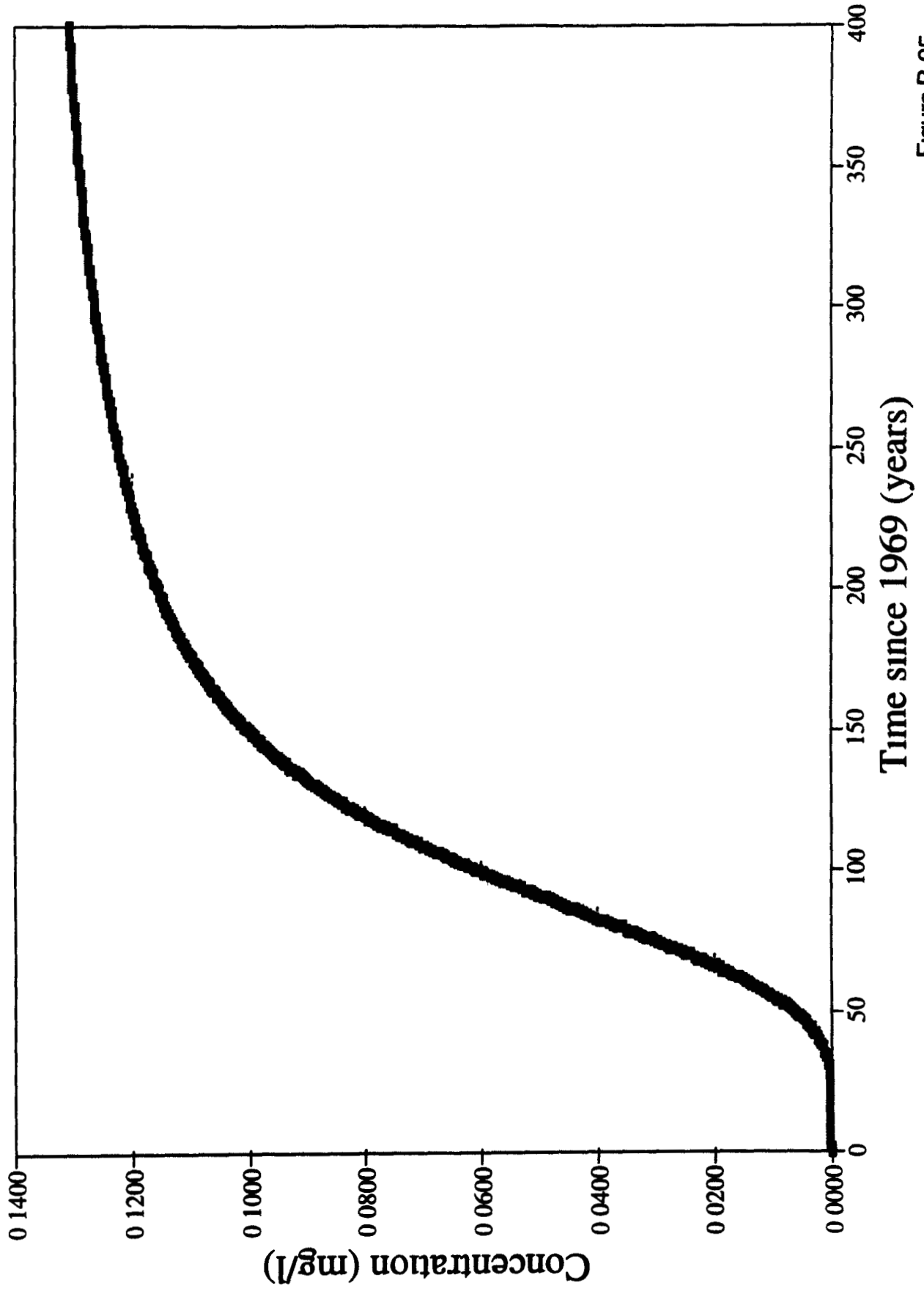


Figure B 25

Simulated concentration of PCE at Woman creek

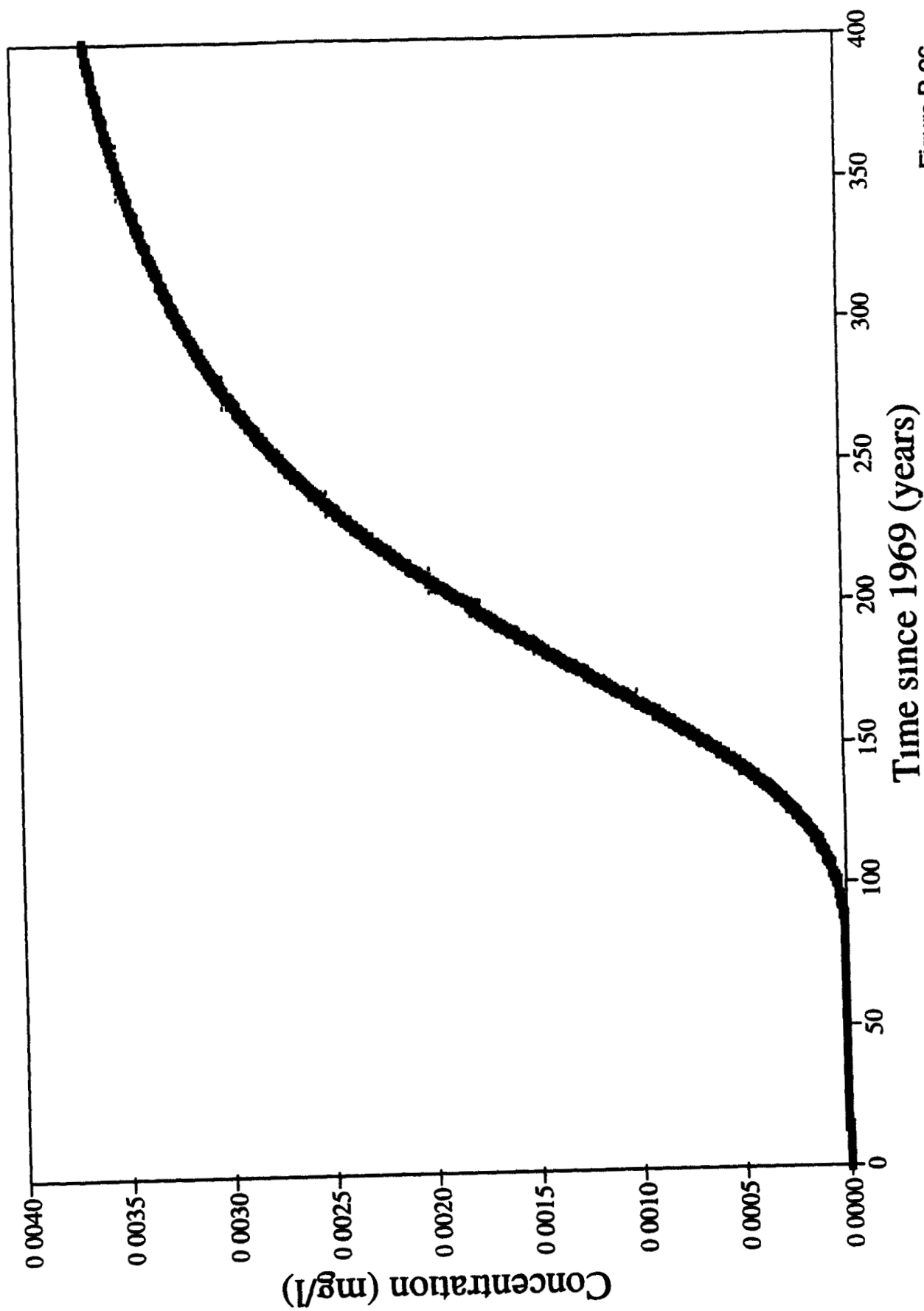


Figure B 26

Simulated concentration of 1,1,1 TCA
down gradient of the French drain

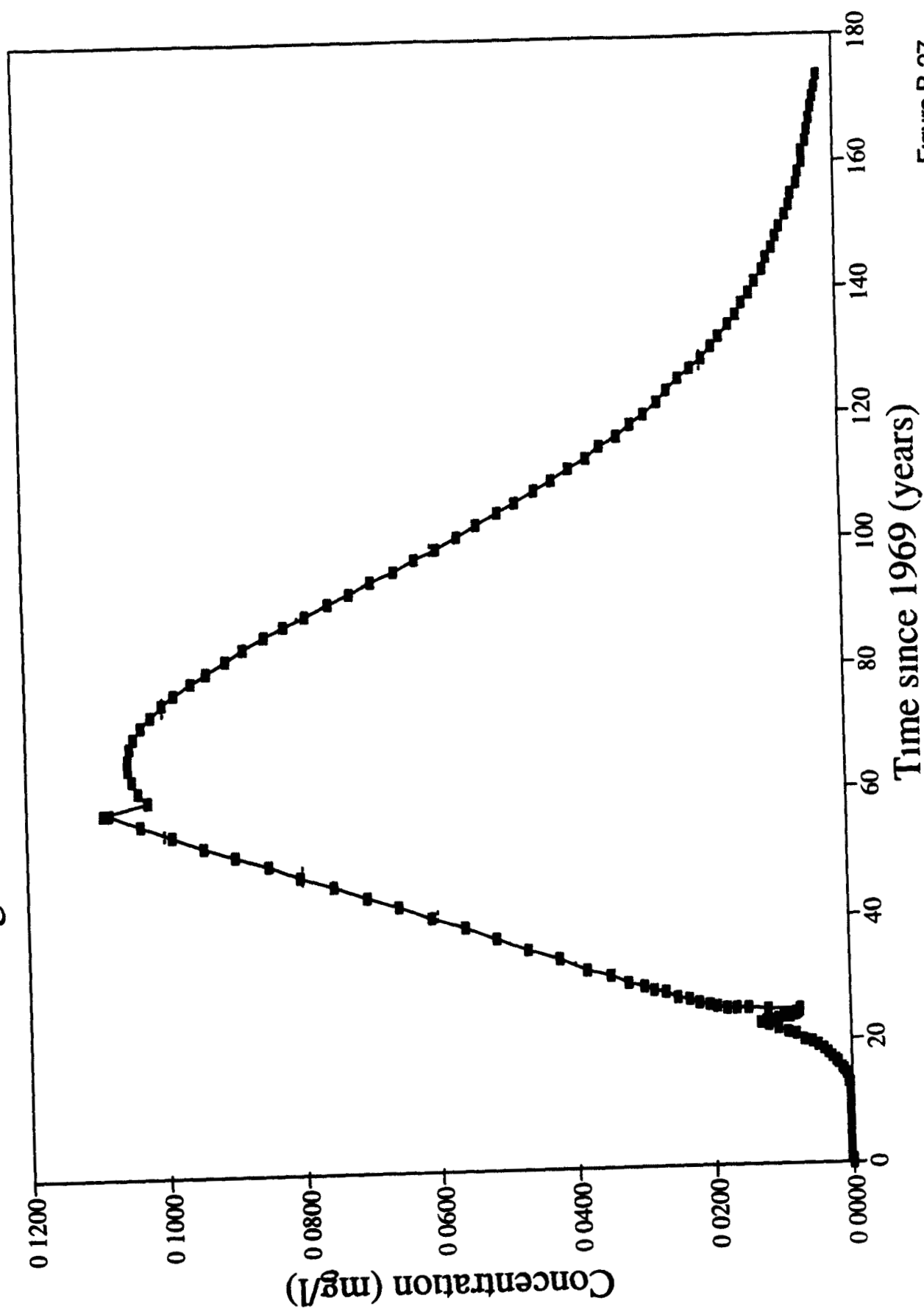


Figure B-27

Simulated concentration of 1,1,1 TCA at Woman creek

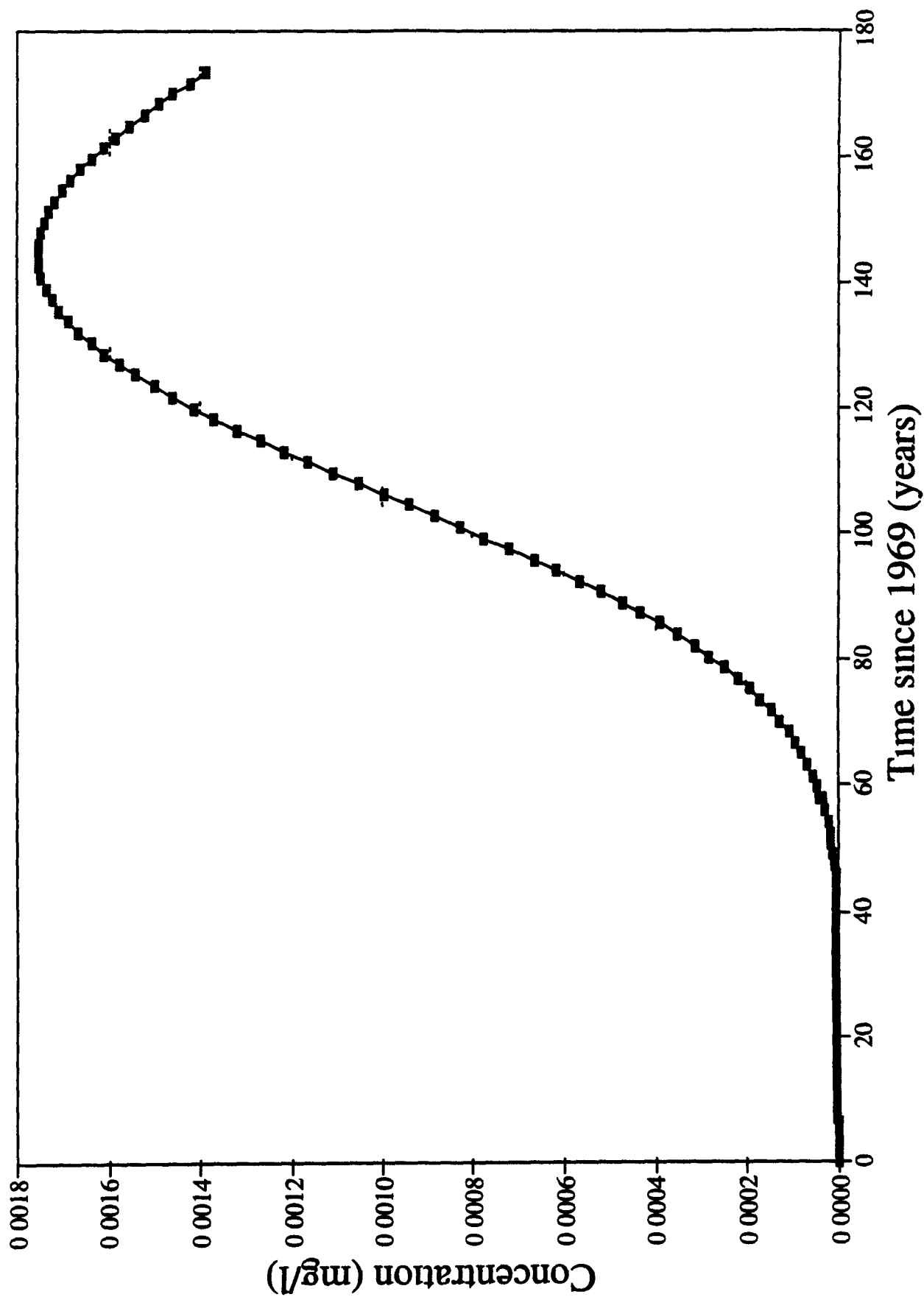


Figure B 28

Simulated concentration of CCL4 down gradient of the French drain

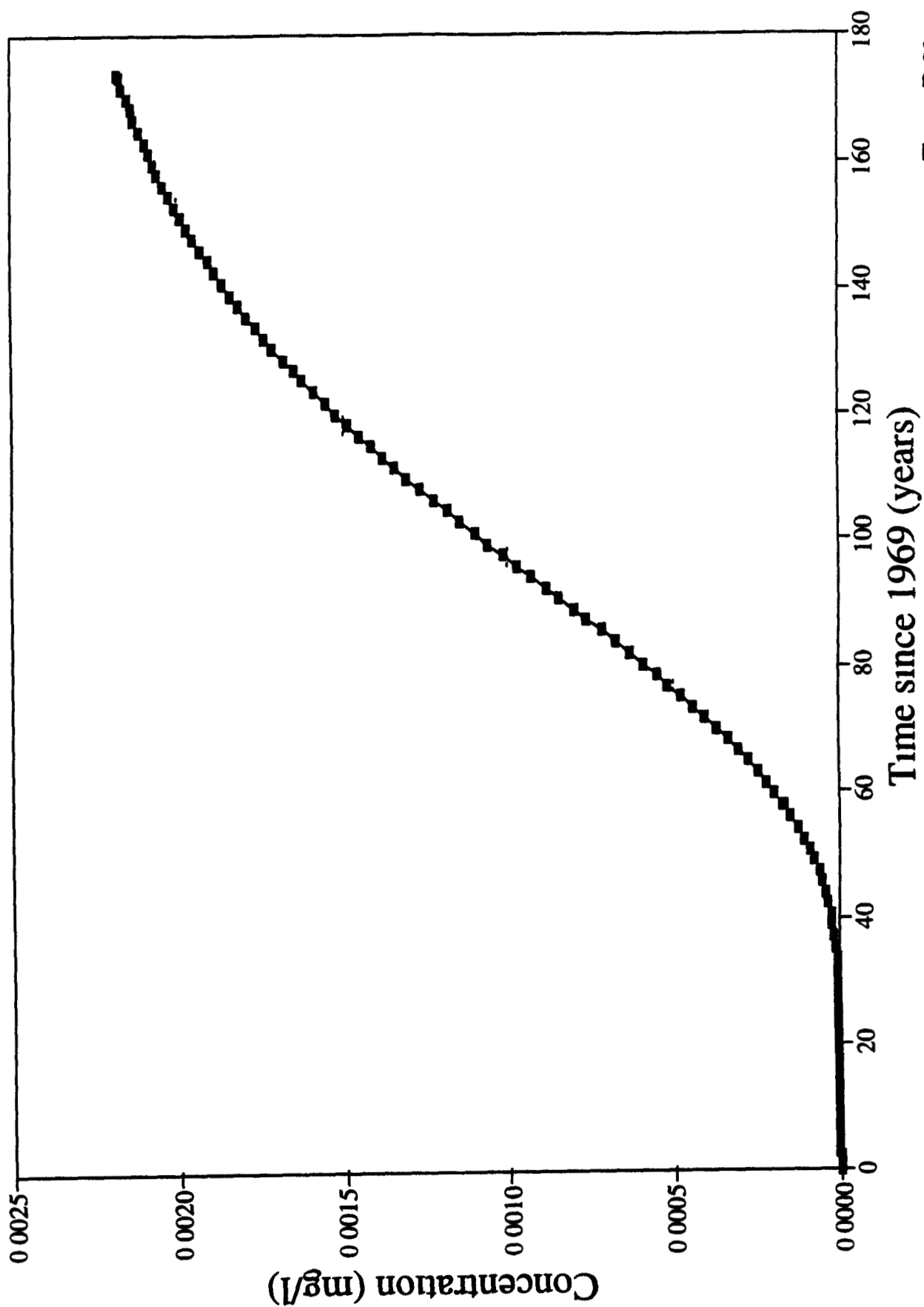


Figure B-29

Simulated concentration of CCL4 at Woman creek

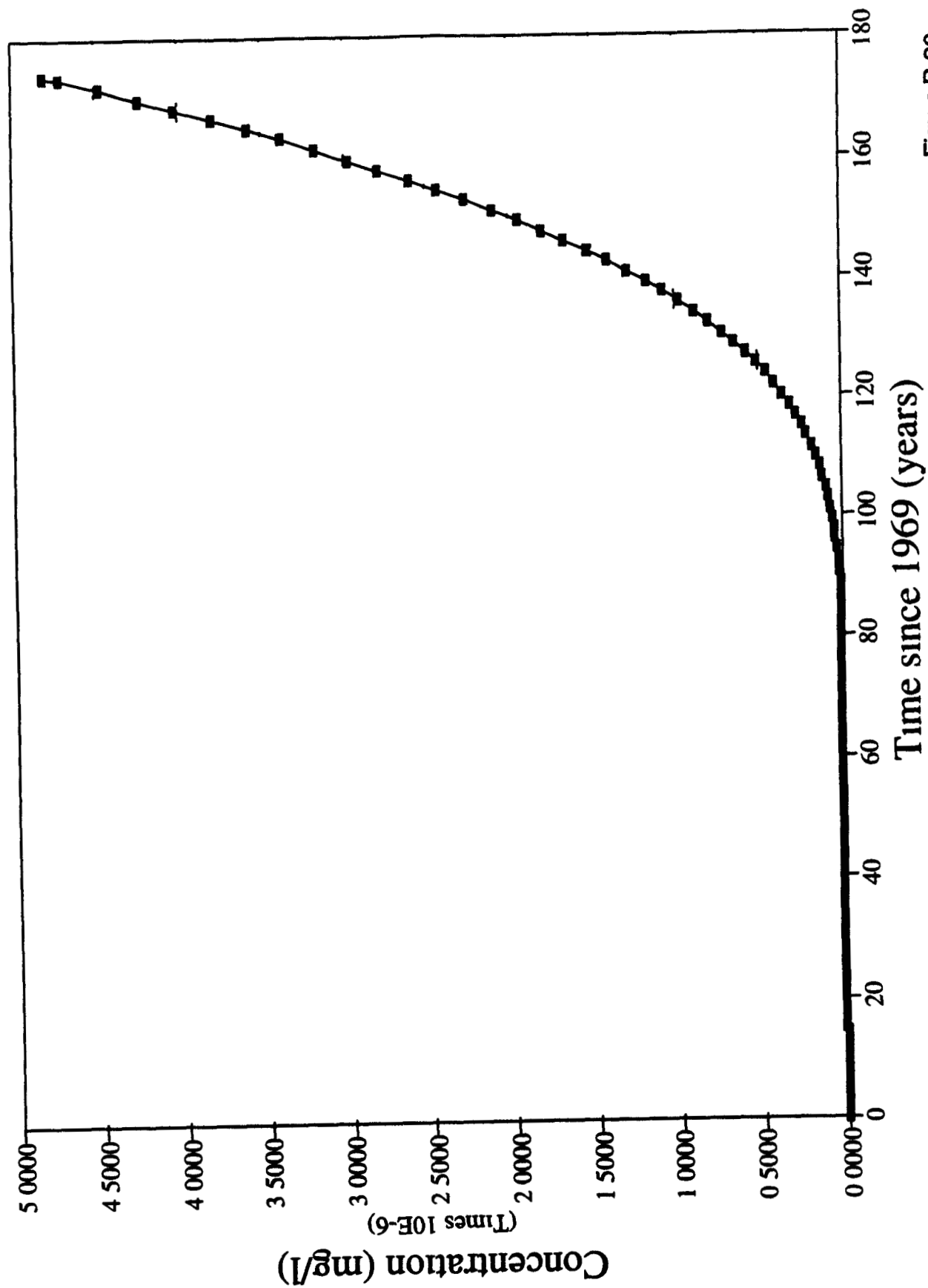


Figure B 30

Simulated concentration of Selenium down gradient of the French drain

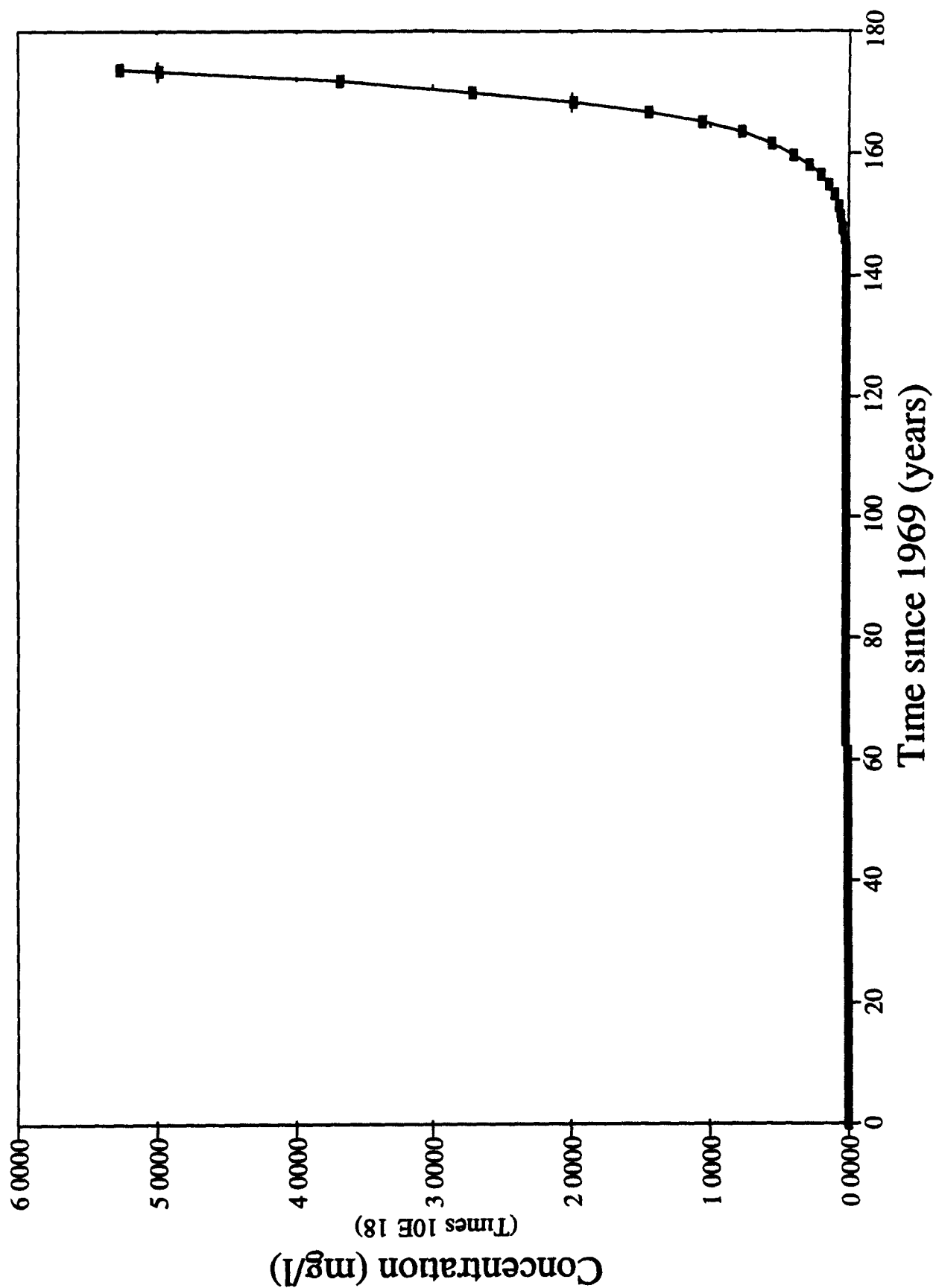


Figure B 31

Simulated concentration of Selenium at Woman creek

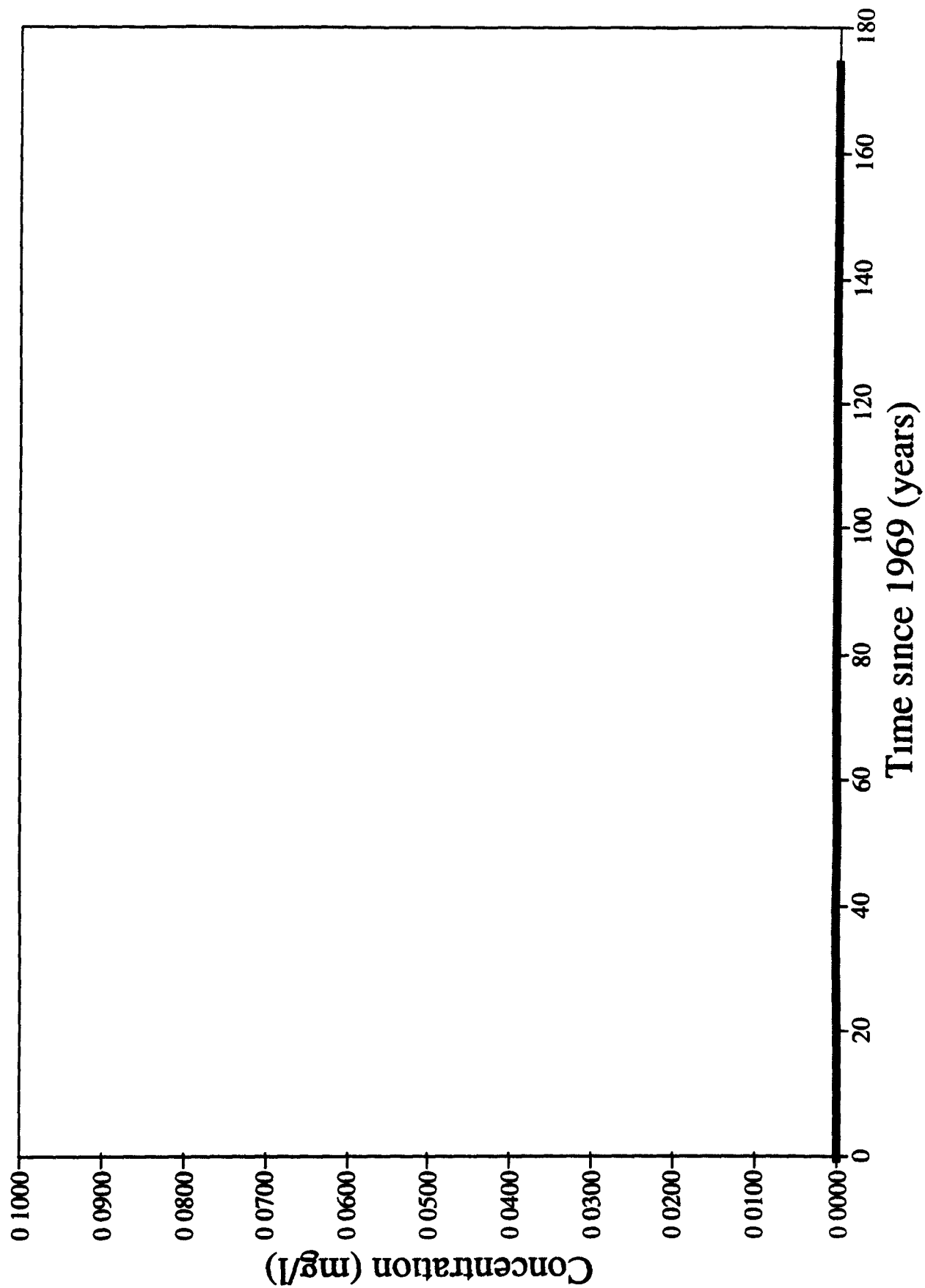
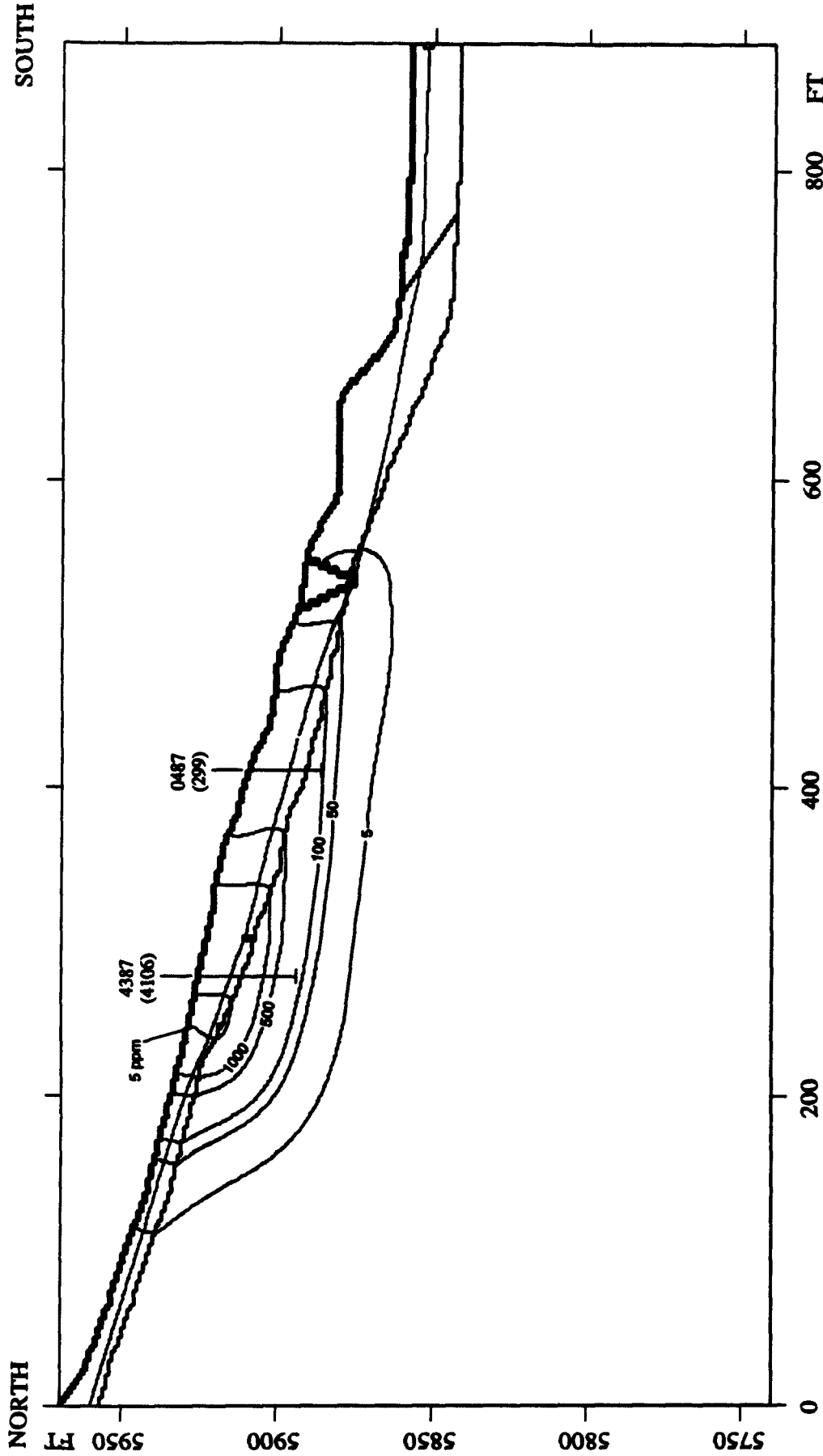


Figure B-32



U S DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO. 1

CONCENTRATION CONTOURS
OF PCE
TIME = 146000 DAYS (2369)

Figure B-33

EXPLANATION

4387	WELL NUMBER
(3600)	AVERAGE PCE CONCENTRATION IN ppb

NOTE: Concentrations in ppb unless otherwise labeled

Simulated concentration of 1,1 DCE
down gradient of the French drain

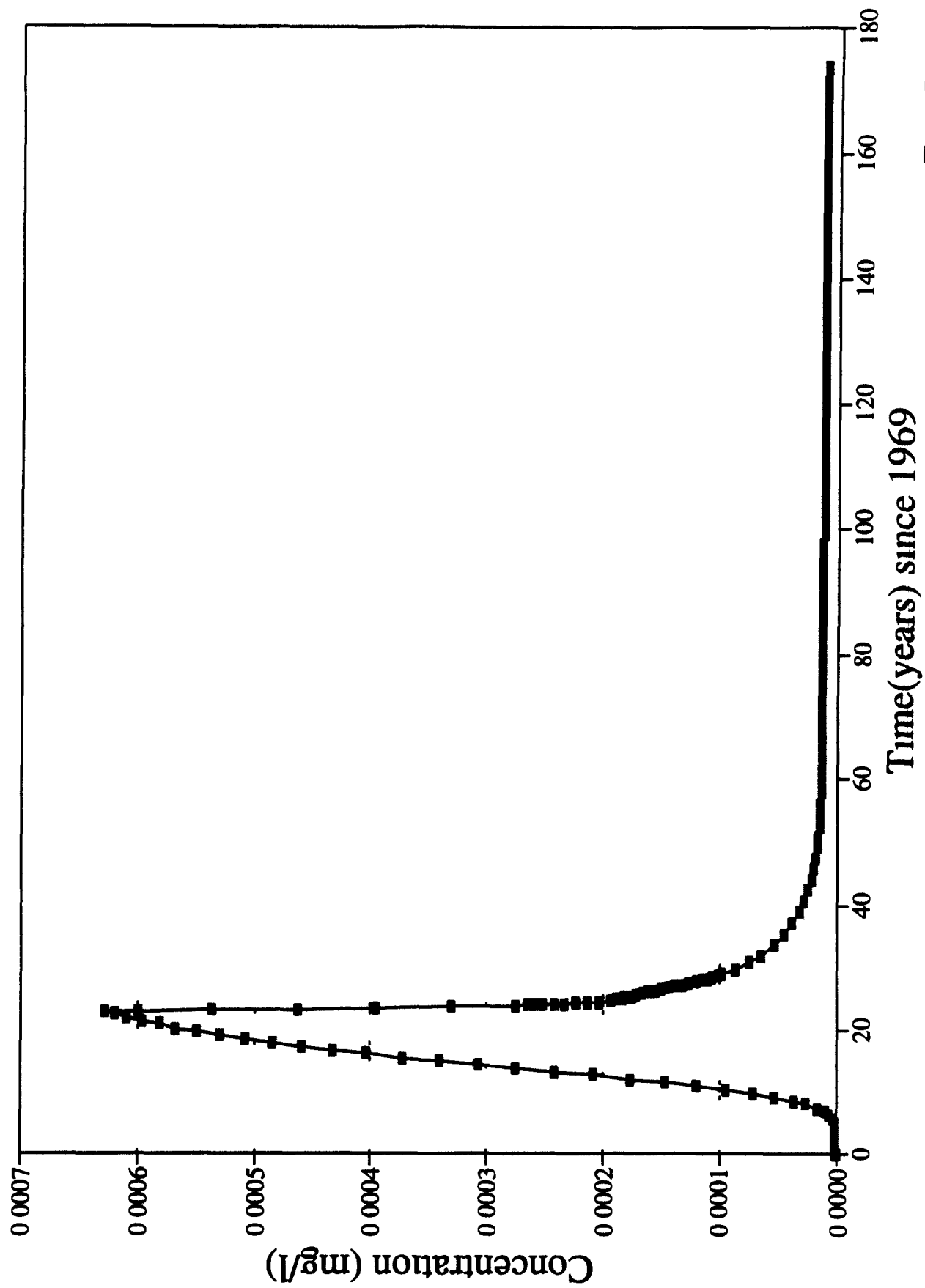


Figure B-34

Simulated concentration of 1,1 DCE at Woman creek

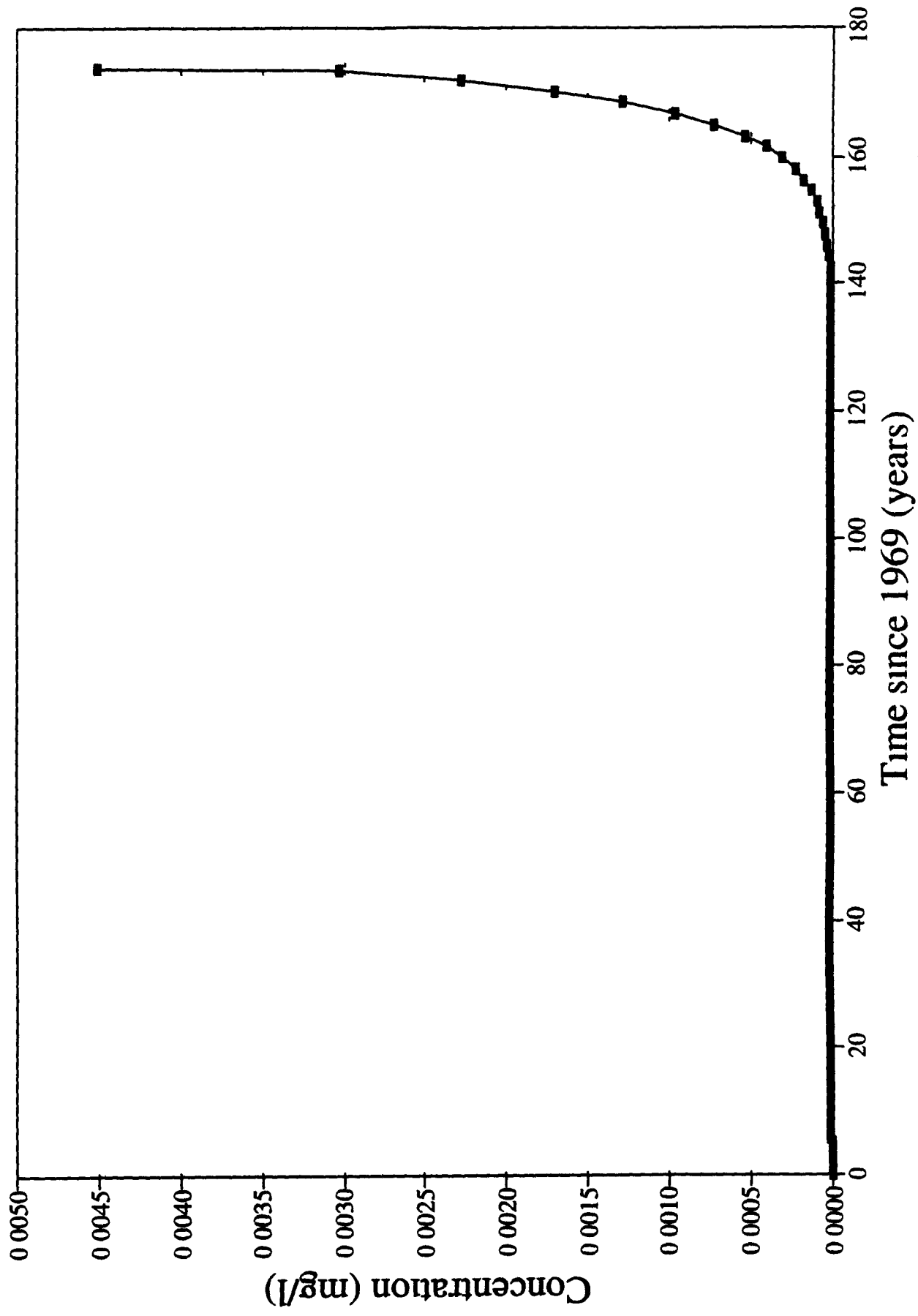


Figure B 35

Simulated concentration of PCE at Woman creek

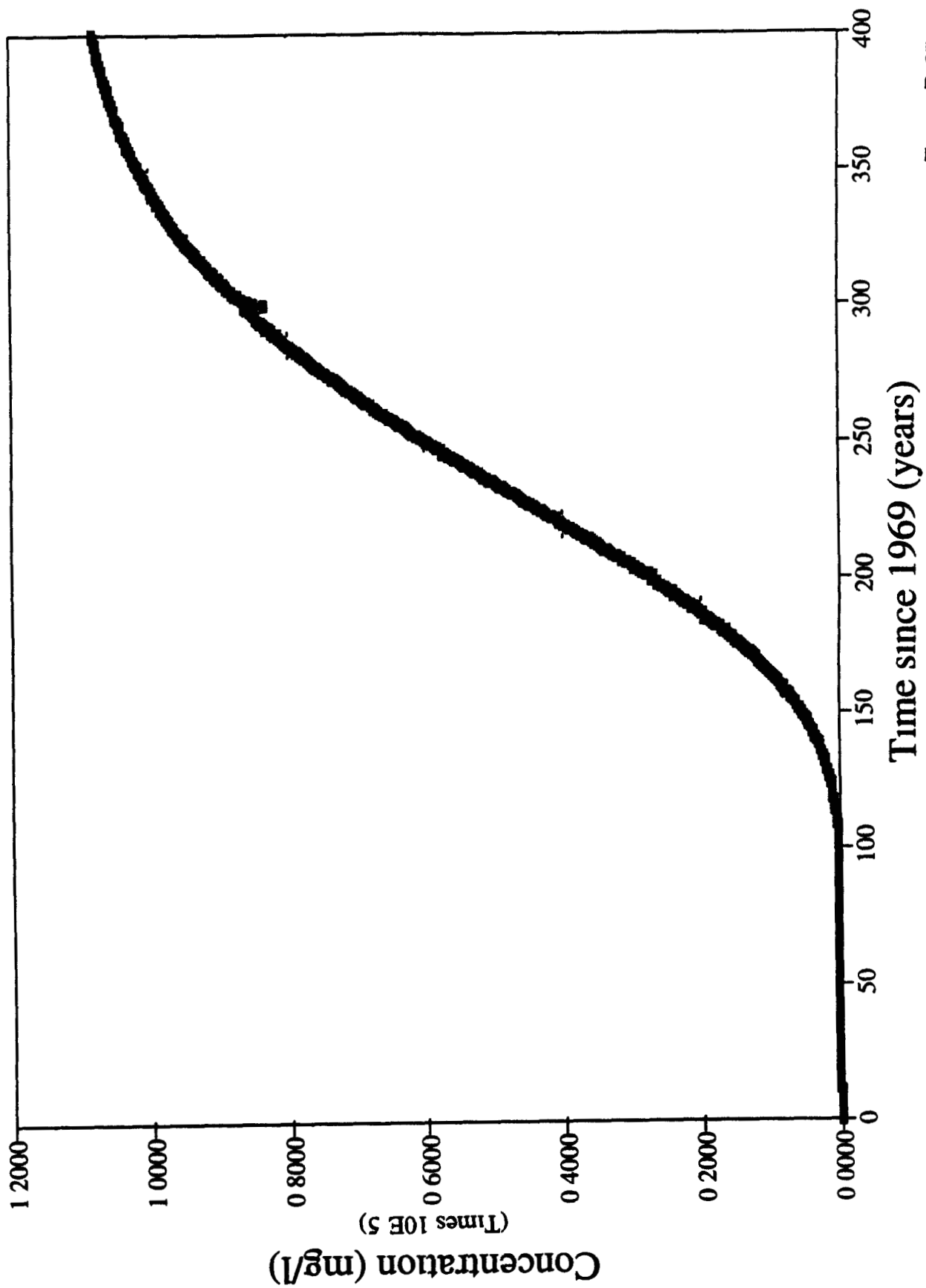


Figure B 37

Simulated concentration of 1,1,1 TCA
down gradient of the French drain

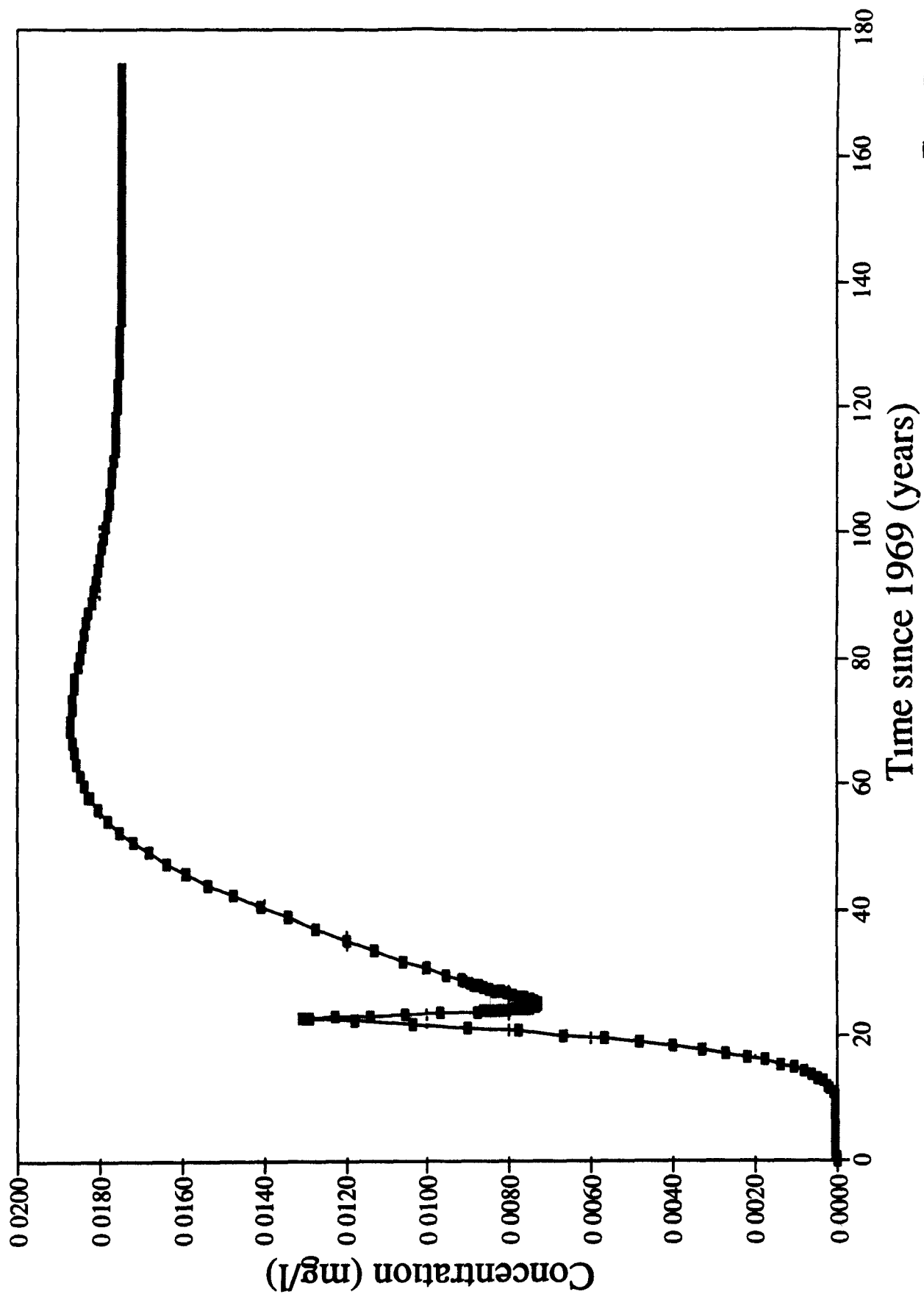


Figure B-38

Simulated concentration of 1,1,1 TCA at Woman creek

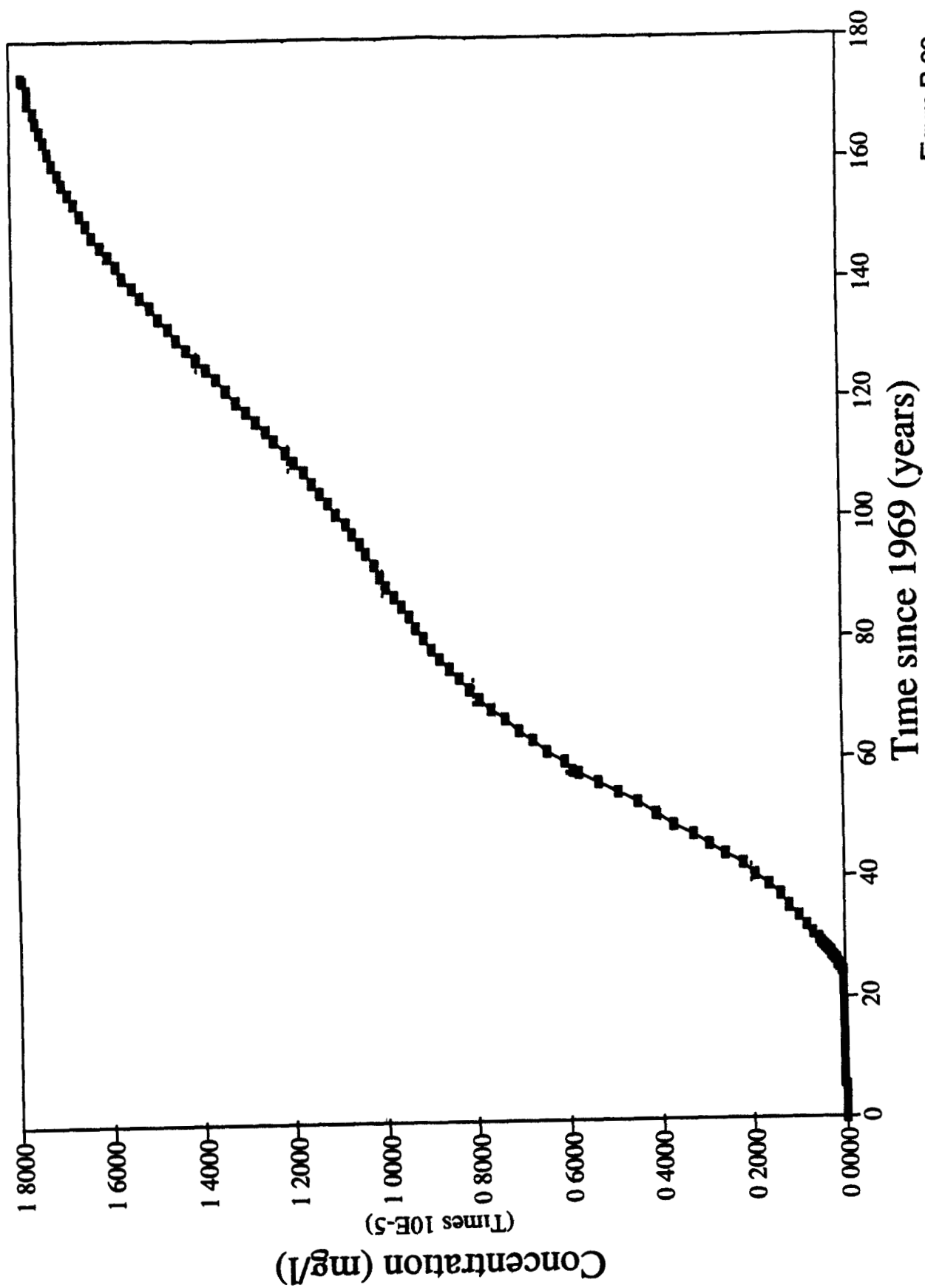


Figure B-39

Simulated concentration of CCL4
down gradient of the French drain

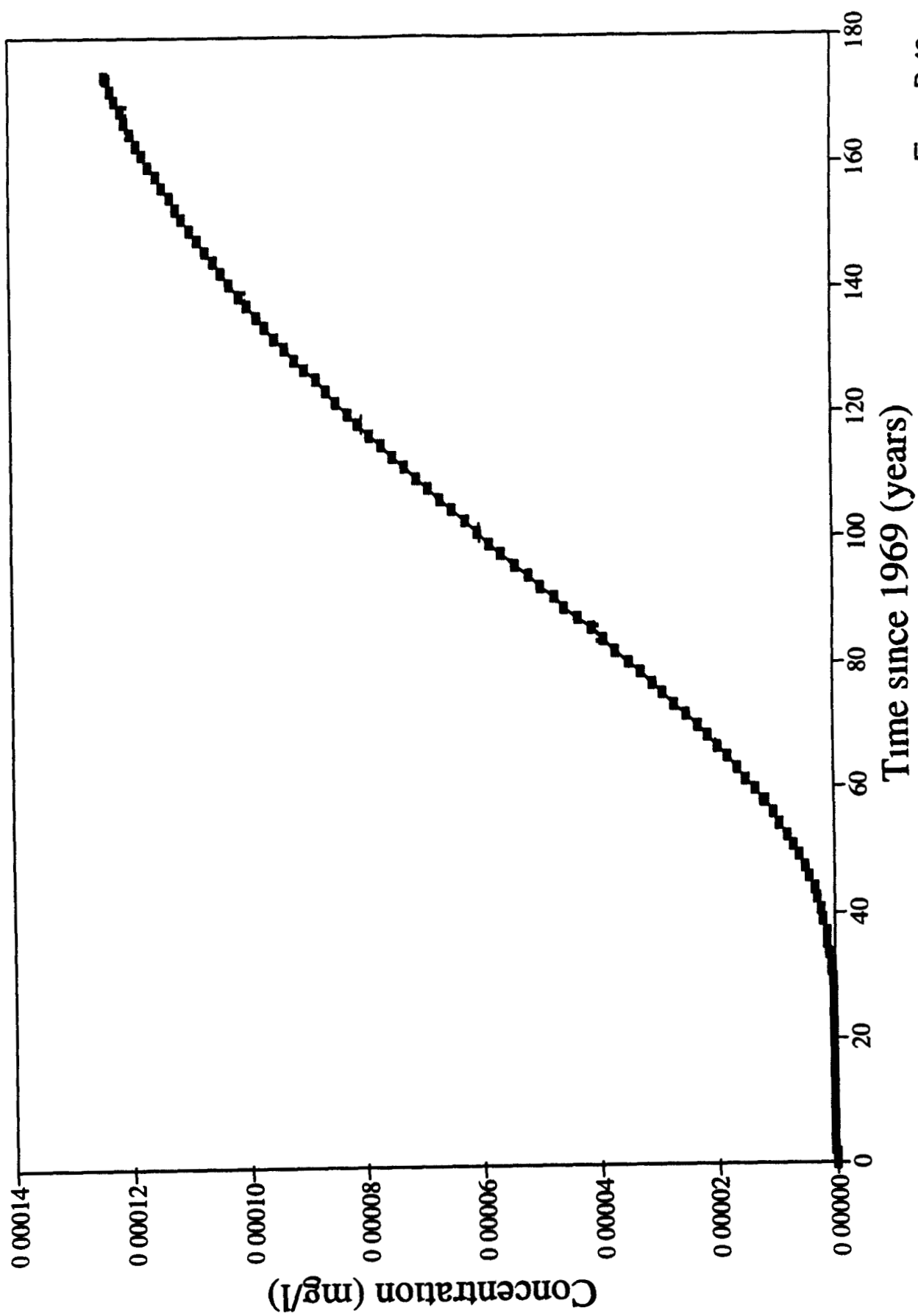


Figure B-40

Simulated concentration of CCL4 at Woman creek

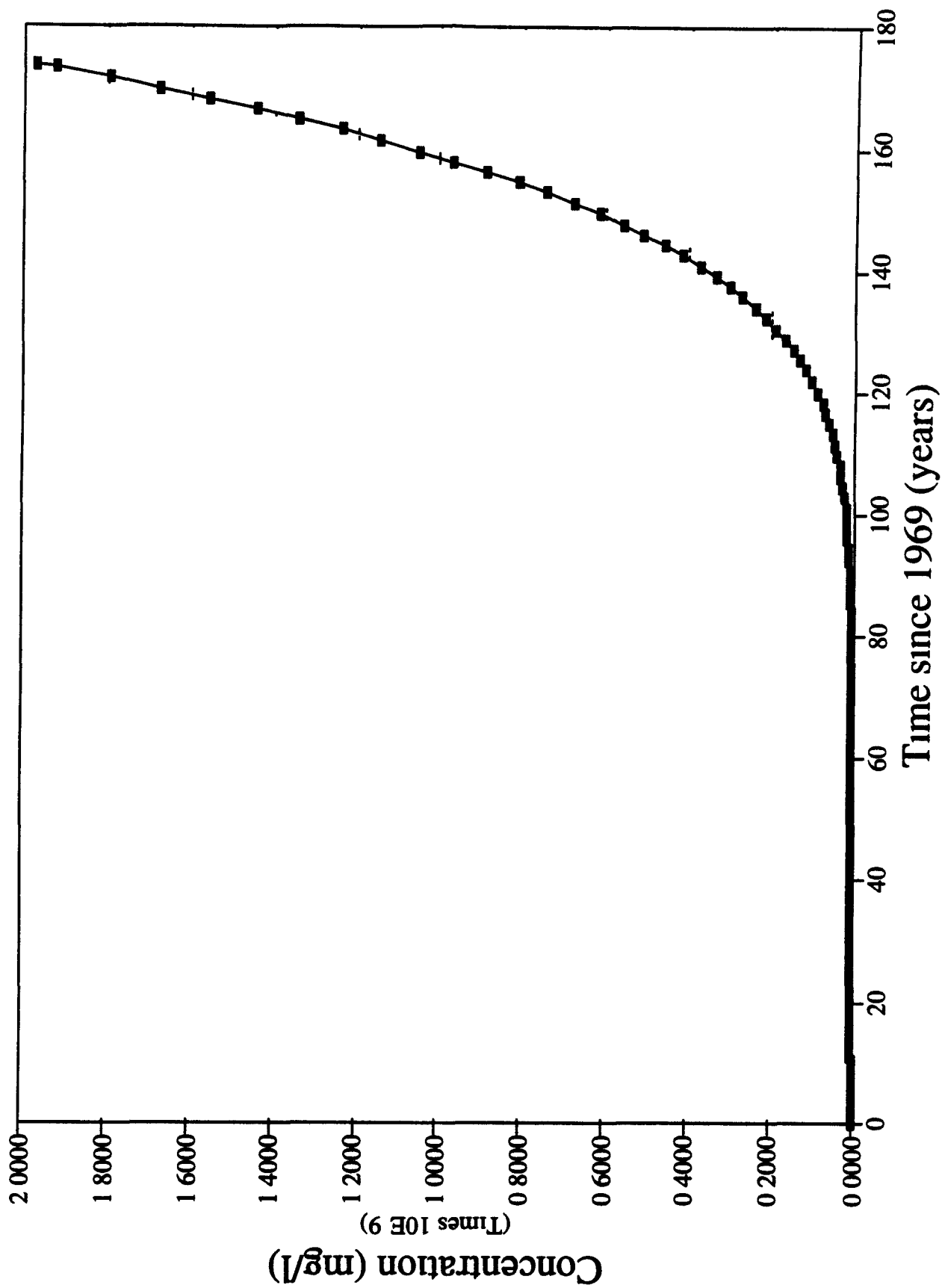


Figure B-41

Simulated concentration of Selenium down gradient of the French drain

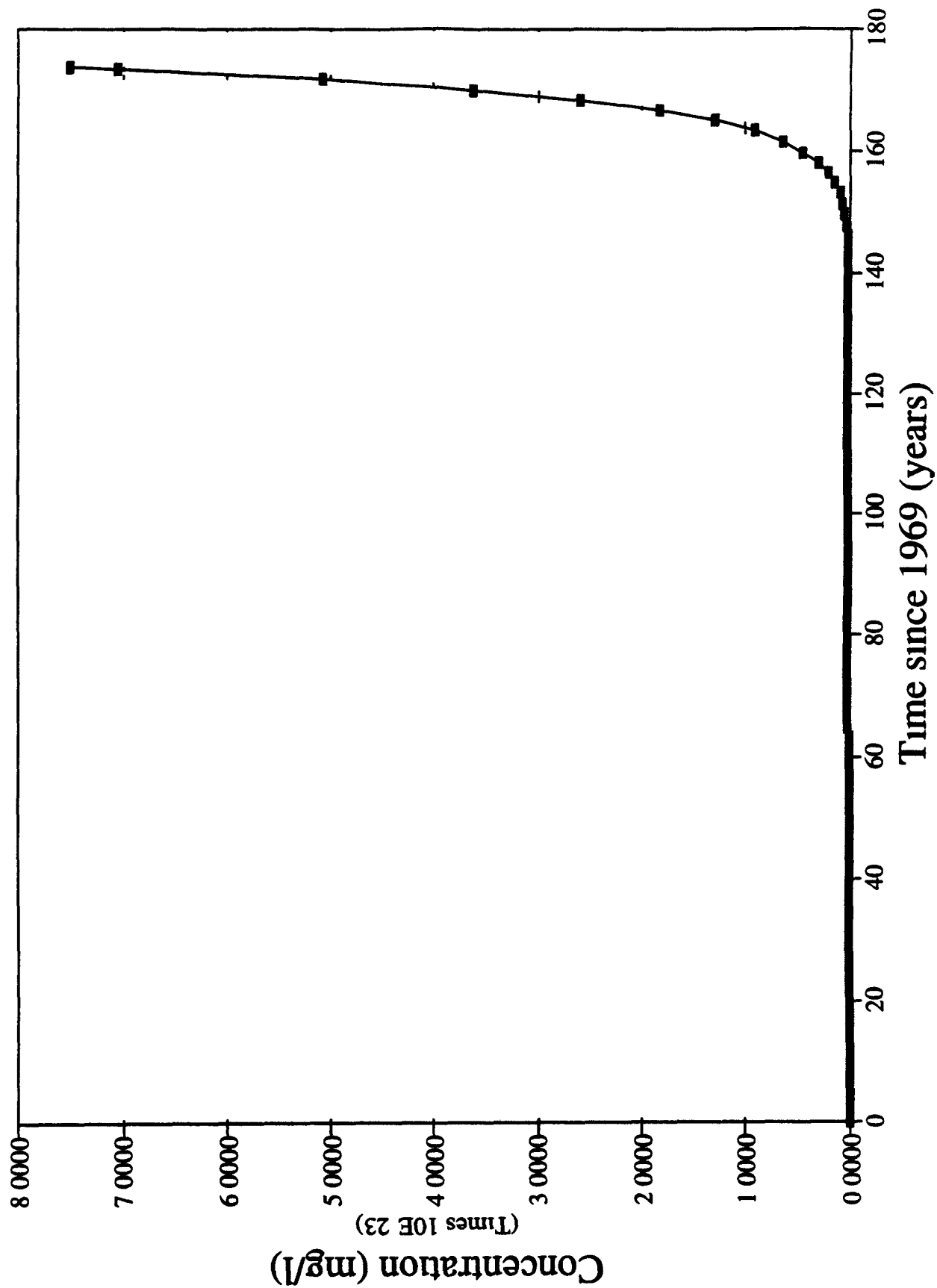


Figure B-42

Simulated concentration of Selenium at Woman creek

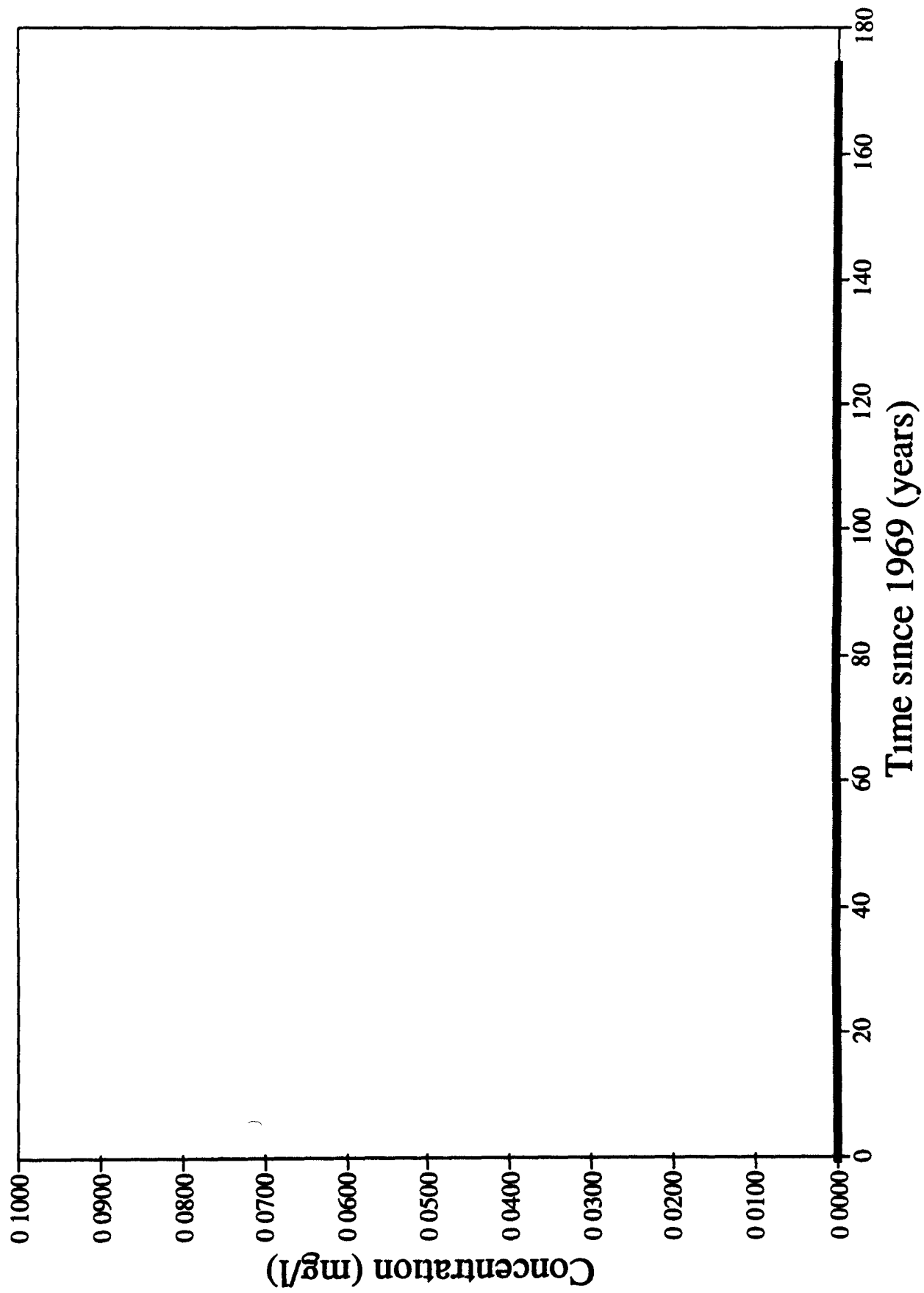
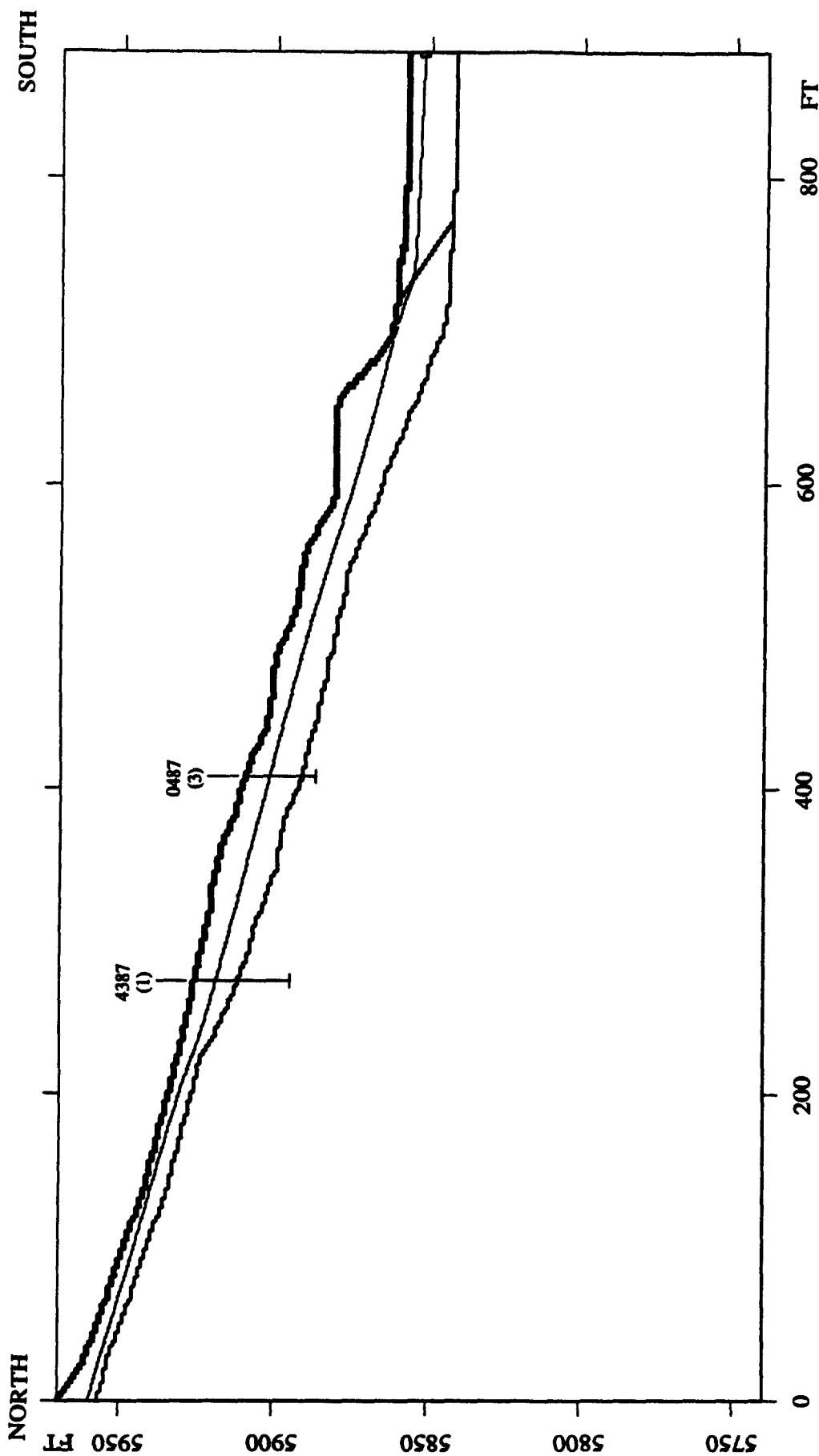


Figure B-43



U.S. DEPARTMENT OF ENERGY

Rocky Flats Environmental Technology Site
Golden, Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO. 1

CONCENTRATION CONTOURS OF PCE

TIME = 73000 DAYS (2169)

Figure B-44

EXPLANATION

WELL NUMBER	AVERAGE PCE CONCENTRATION IN PPB
4387	(3600)

NOTE: Concentrations in ppb unless otherwise labeled

Simulated concentration of 1,1 DCE
down gradient of the French drain

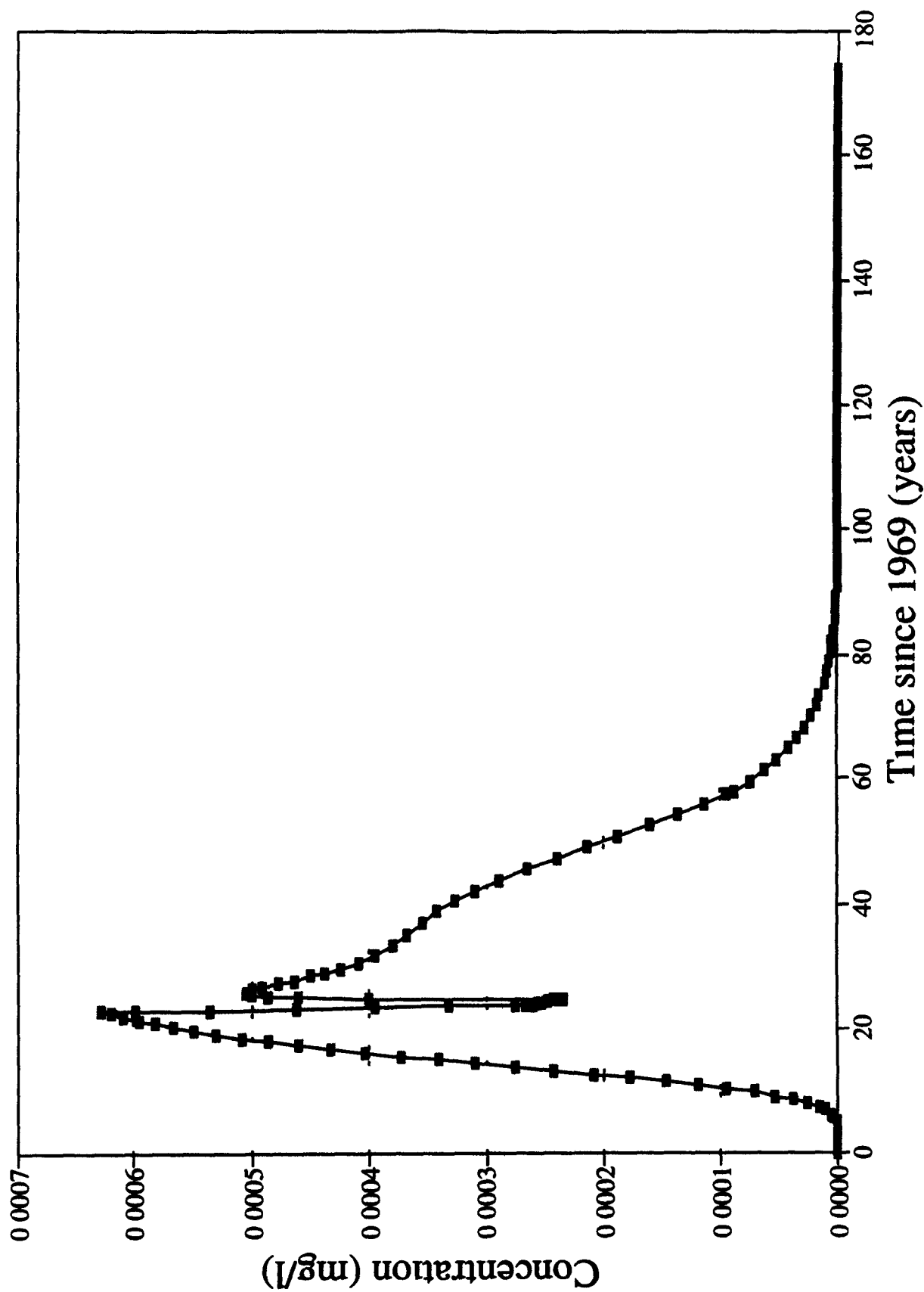


Figure B-45

Simulated concentration of 1,1 DCE at Woman creek

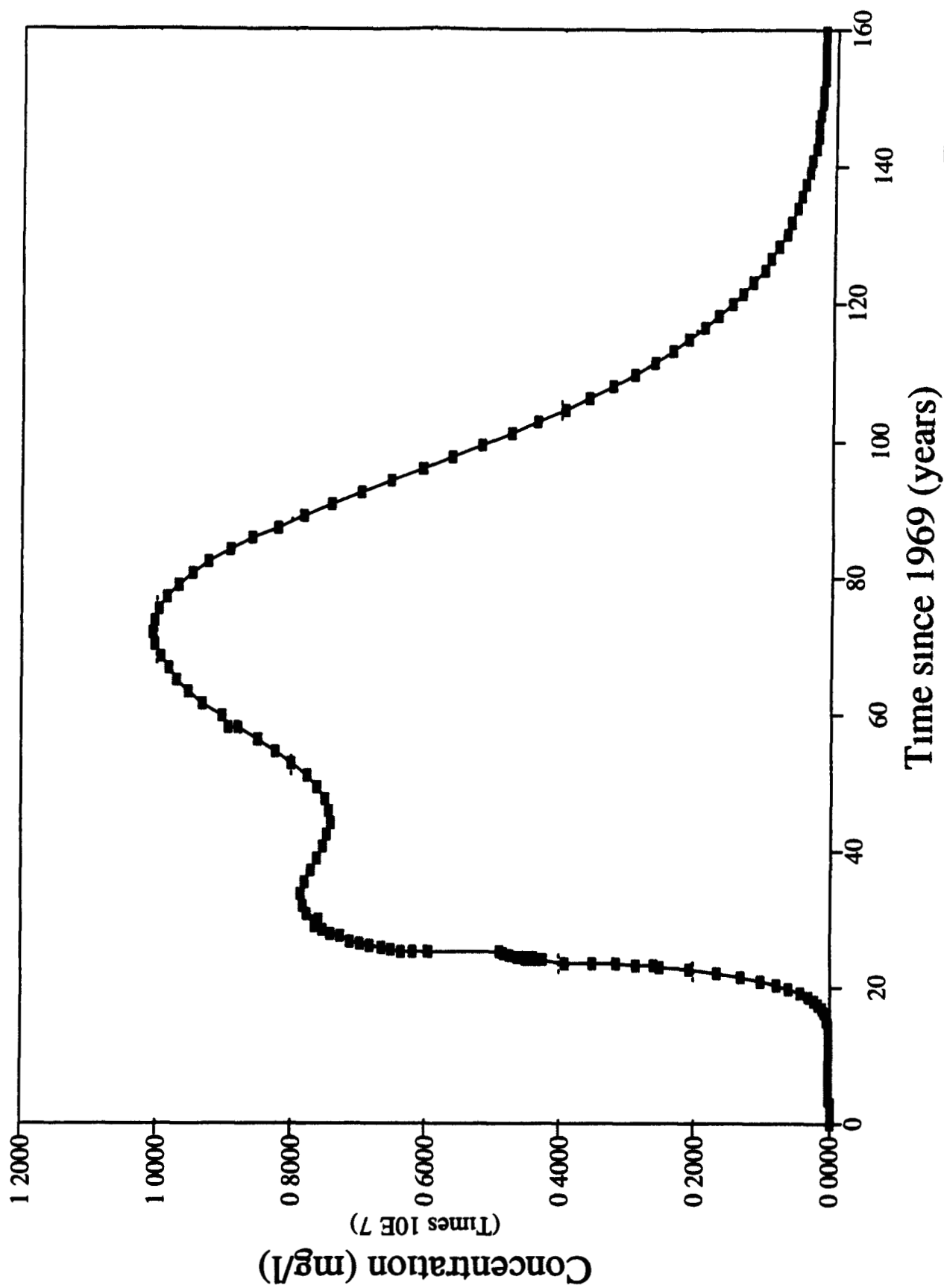


Figure B-46

Simulated concentration of PCE
down gradient of the French drain

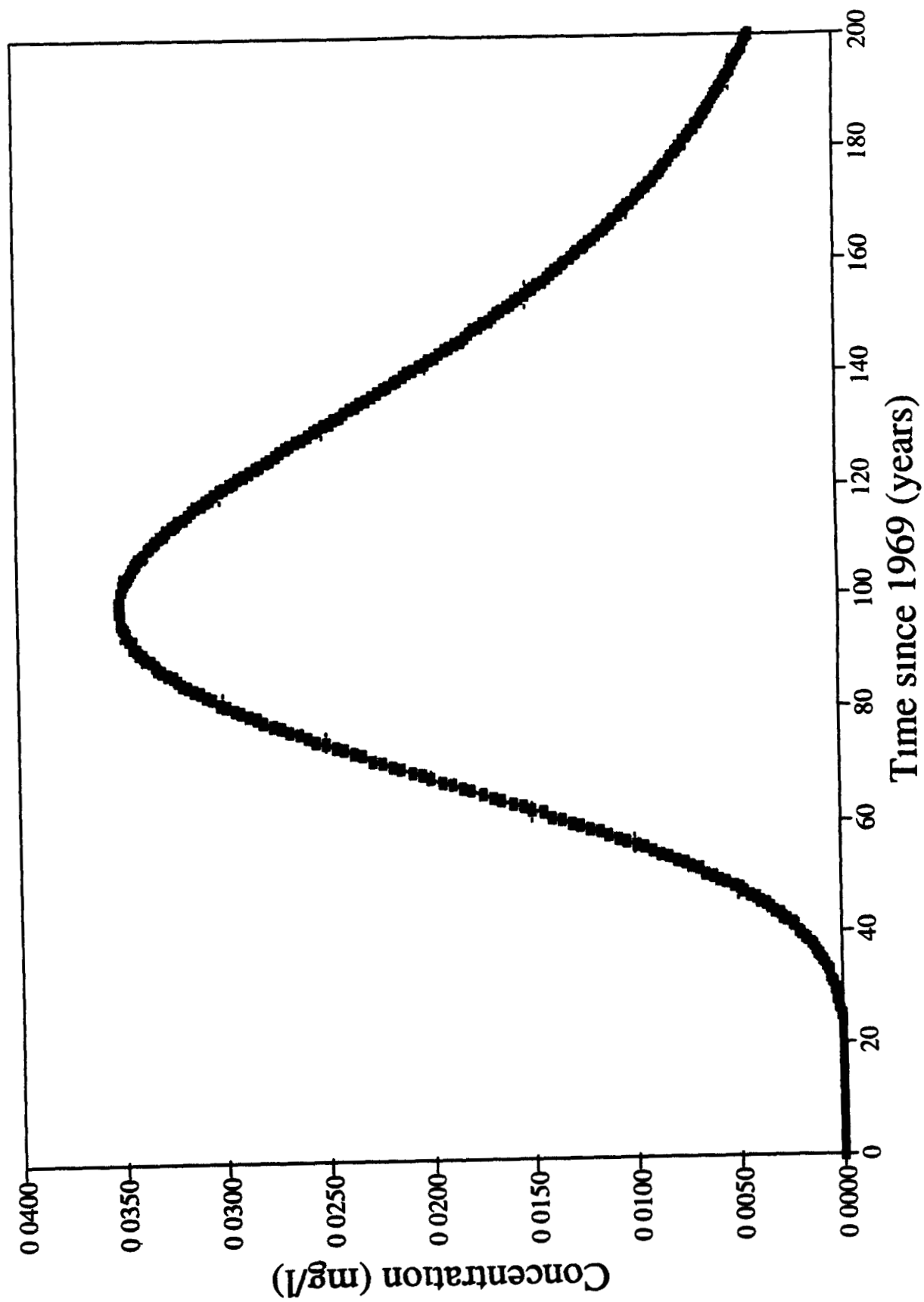


Figure B-47

Simulated concentration of PCE at Woman creek

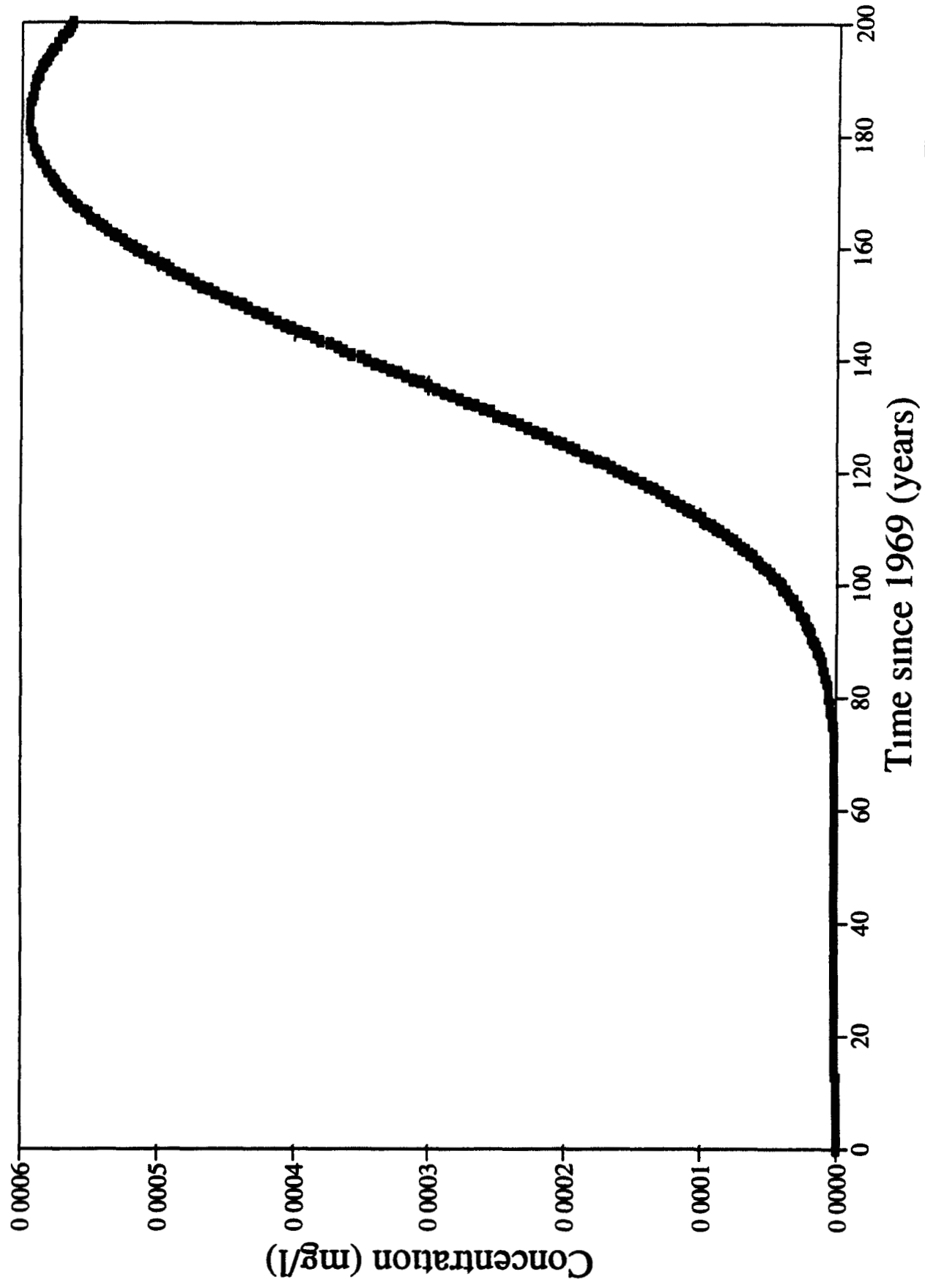


Figure B-48

Simulated concentration of 1,1,1 TCA down gradient of the French drain

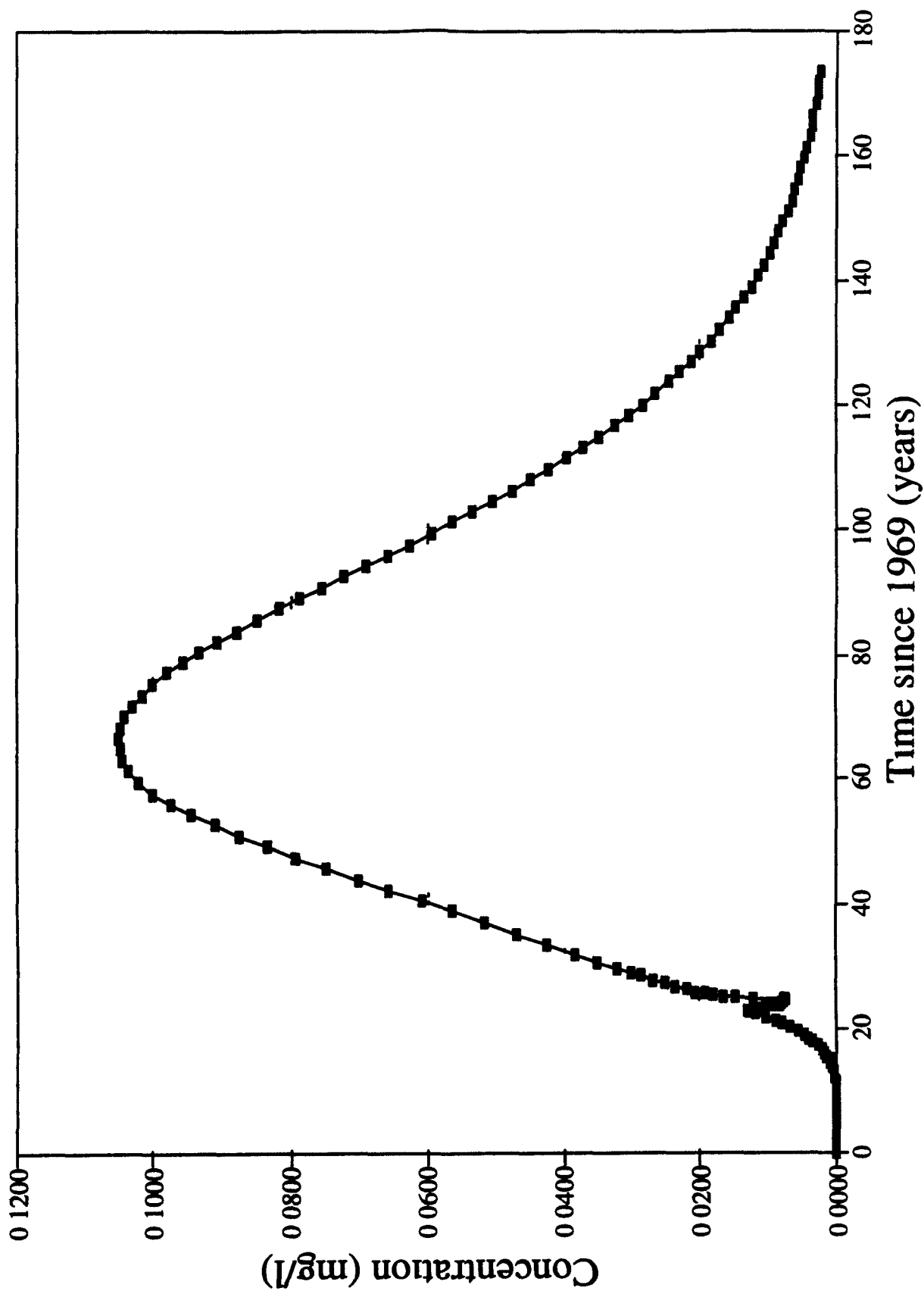


Figure B-49

Simulated concentration of 1,1,1 TCA at Woman creek

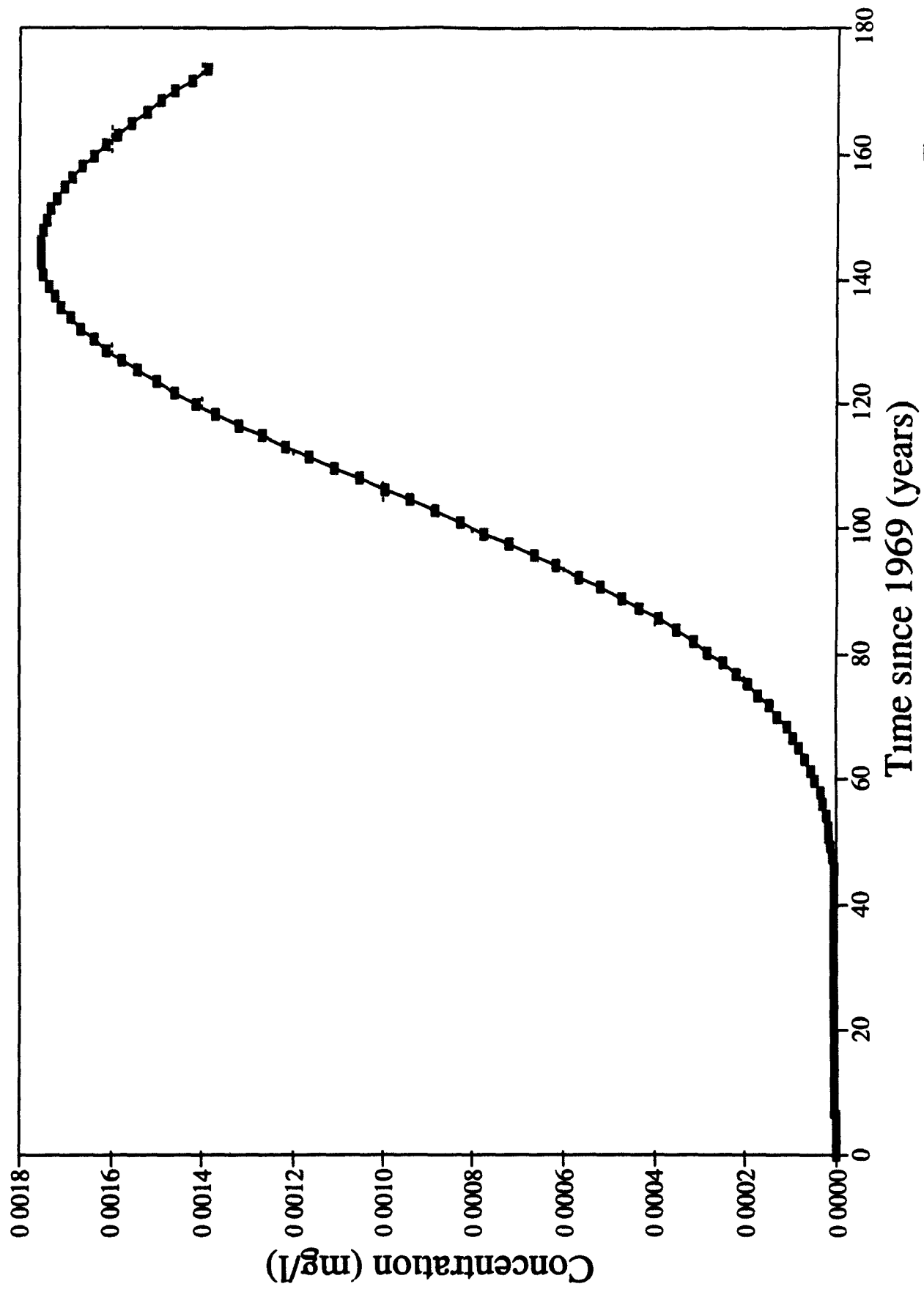


Figure B 50

Simulated concentration of CCL4 down gradient of the French drain

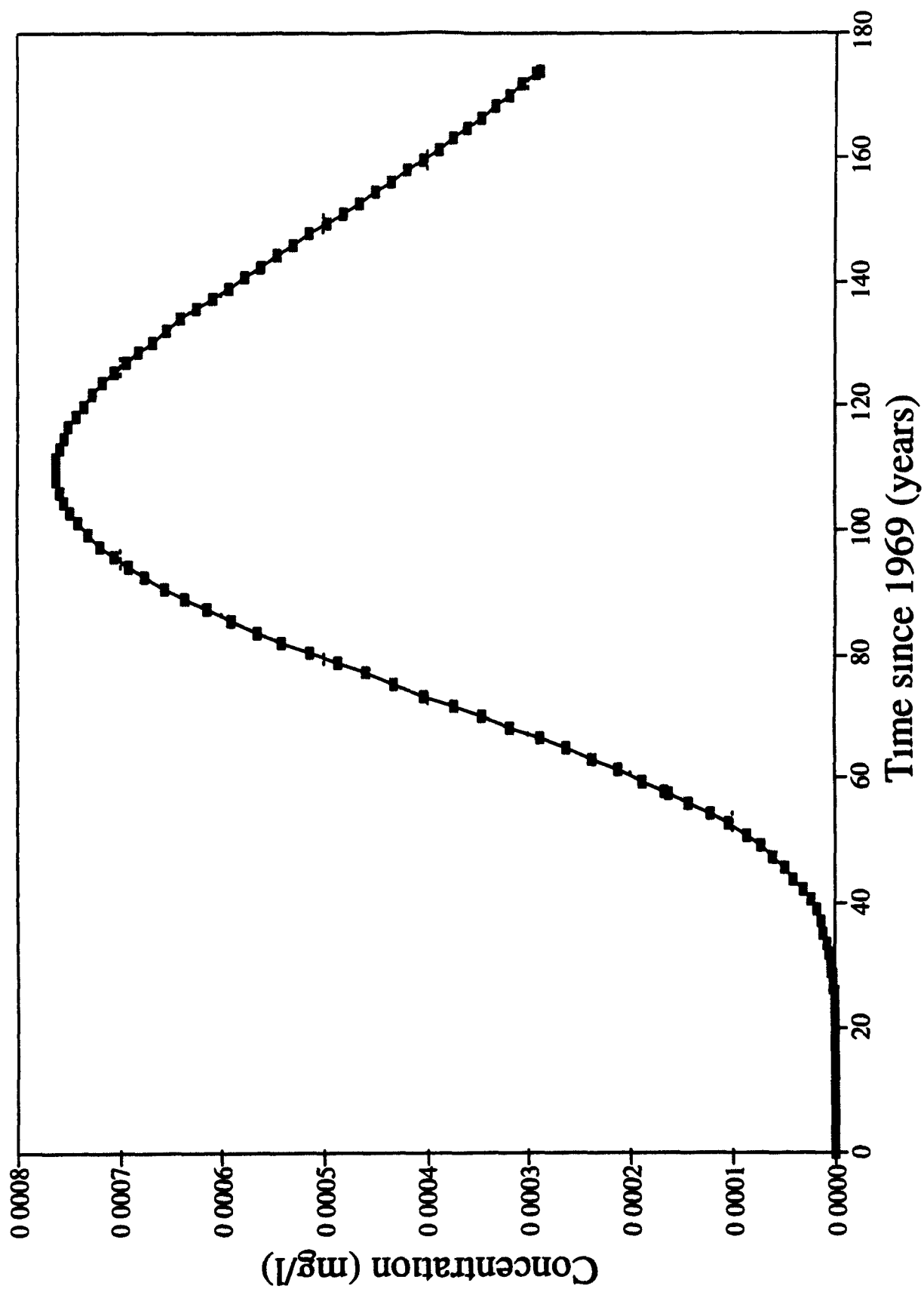


Figure B 51

Simulated concentration of CCL4 at Woman creek

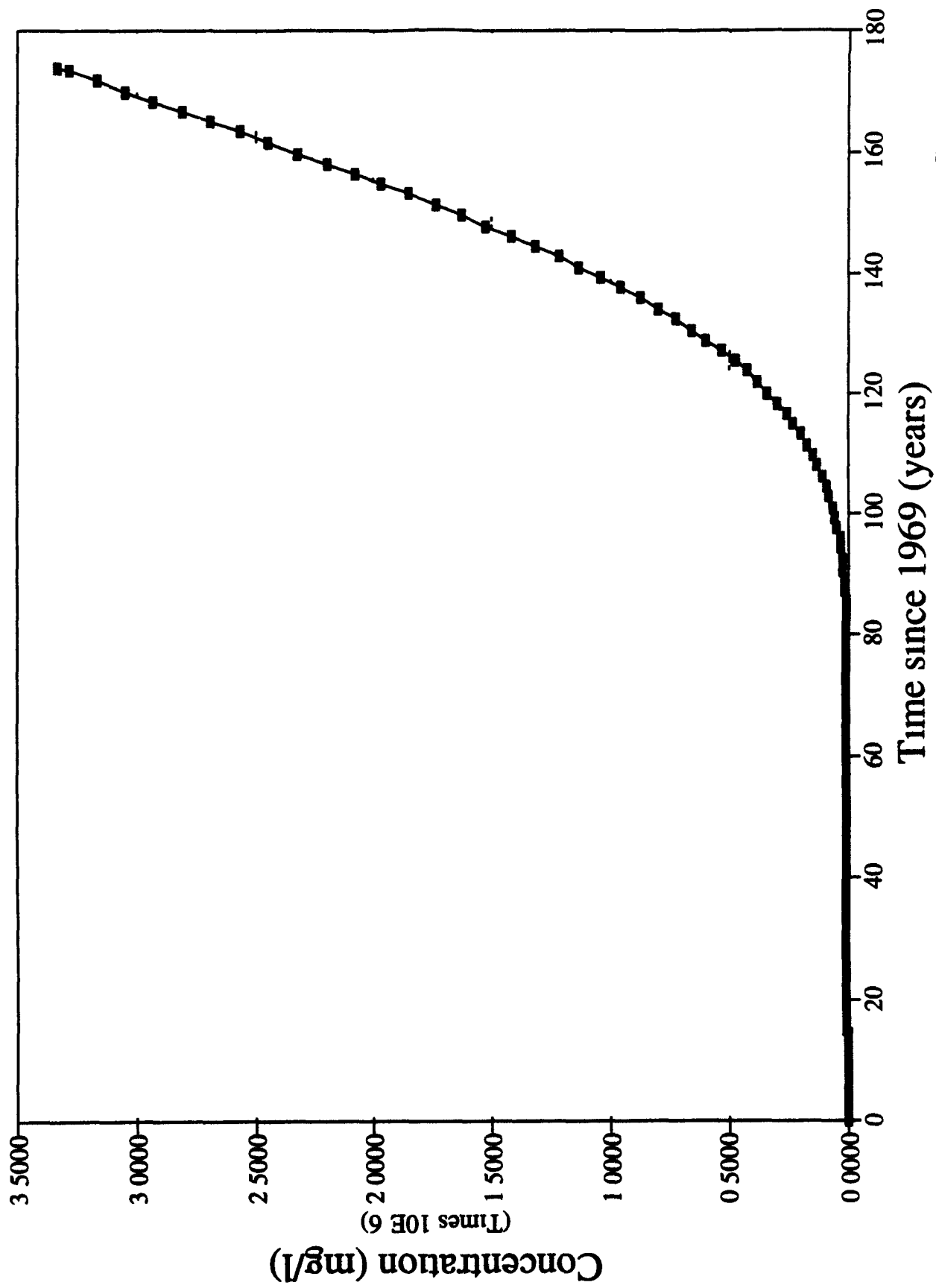


Figure B 52

Simulated concentration of Selenium down gradient of the French drain

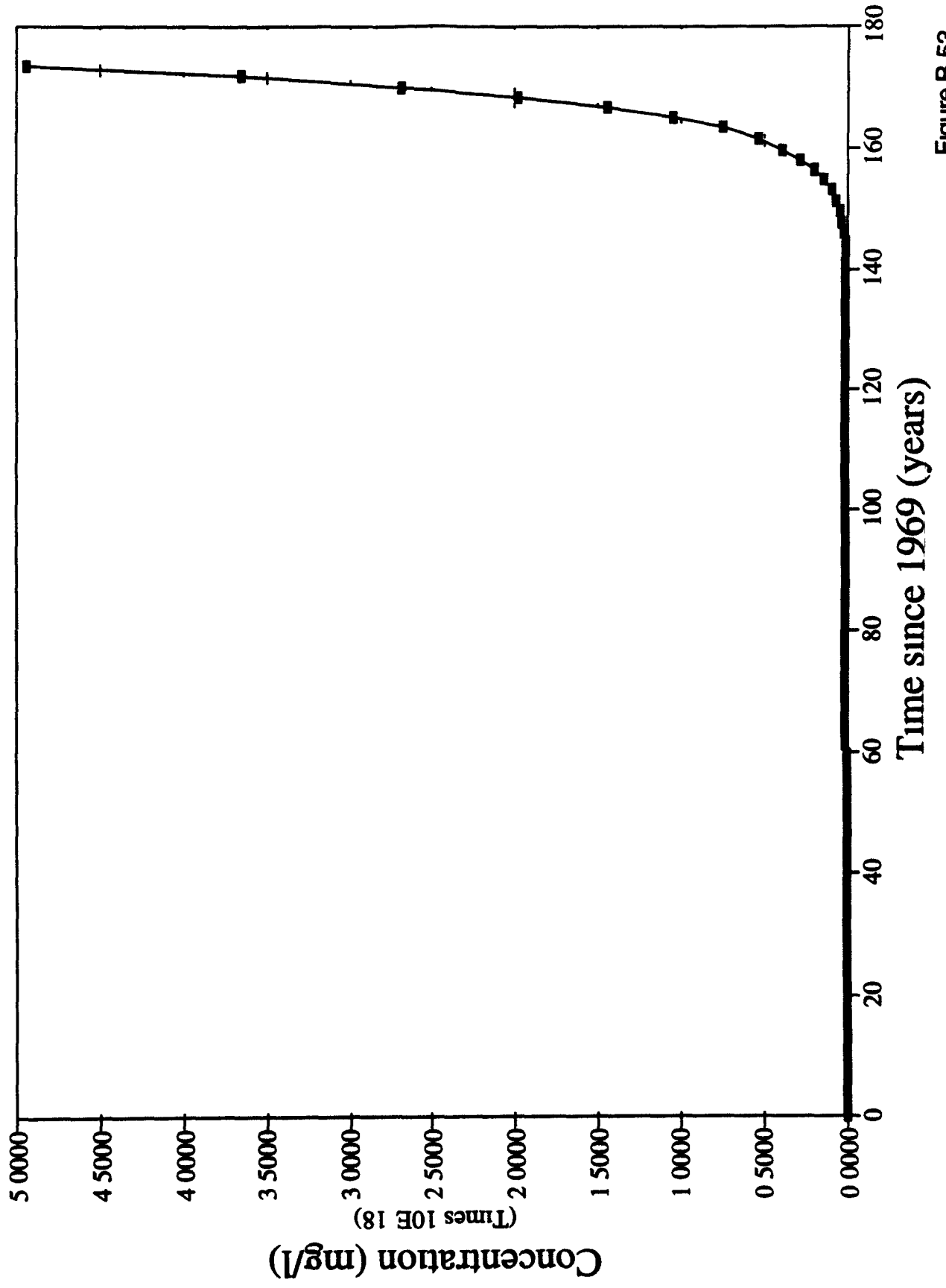


Figure B 53

Simulated concentration of Selenium at Woman creek

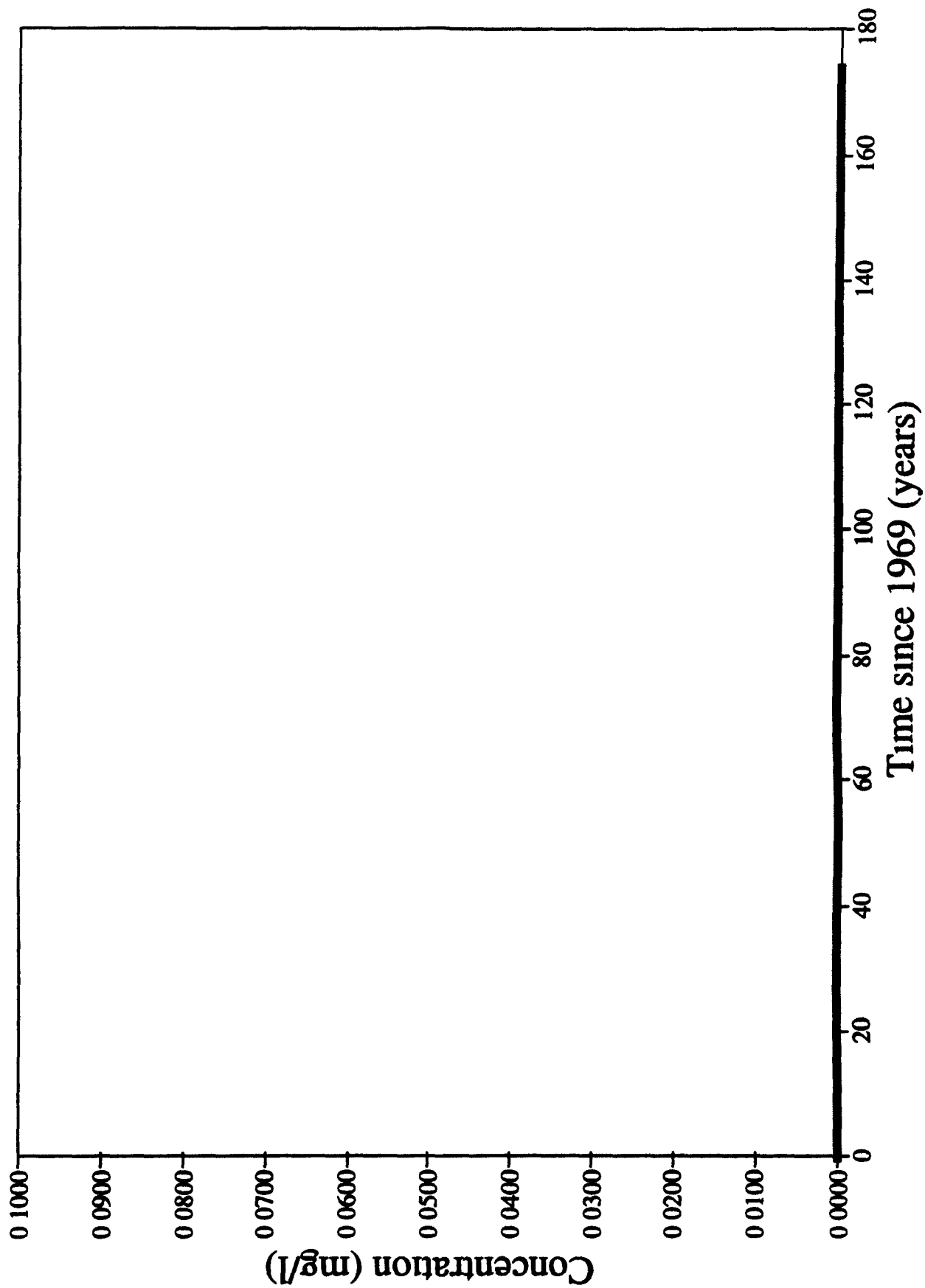


Figure B 54

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ACRONYMS

1 1 DCE	1 1 dichloroethene
CERCLA	Comprehensive Environmental Response Compensation and Liability Act
CCL ₄	carbon tetrachloride
CNS	central nervous system
DOE	Department of Energy
EE	Ecological Evaluation
FS	Feasibility Study
HI	hazard indices
HQ	hazard quotient
NOAEL	no observed adverse effect level
OU1	Operable Unit No 1
PCE	tetrachloroethene
PHE	Public Health Evaluation
RAGS	Risk Assessment Guidance for Superfund
RCRA	Resource Conservation and Recovery Act
RfD	reference dose
RFETS	Rocky Flats Environmental Technology Site
RFI/RI	RCRA Facility Investigation/Remedial Investigation
SFs	slope factors
VOCs	volatile organic compounds

C 1 0 INTRODUCTION

The Phase III Resource Conservation and Recovery Act (RCRA) Facility Investigation/Comprehensive Environmental Response Compensation and Liability Act (CERCLA) Remedial Investigation (RFI/RI) at Operable Unit No 1 (OU1) 881 Hillside Area at the Rocky Flats Environmental Technology Site (RFETS) includes a Baseline Risk Assessment (BRA) The BRA is comprised of an Ecological Evaluation (EE) and a Public Health Evaluation (PHE) The results of the complete OU1 PHE are presented in Volume X Appendix F of the Final Phase III RFI/RI dated June 1994 (DOE 1994a)

This risk assessment performed for the OU1 Feasibility Study (FS) is intended to calculate and document the human health risks associated with OU1 assuming that specified remedial actions are incorporated at the site This risk assessment considered the dominating carcinogenic risks noncarcinogenic hazards associated contaminant pathways and receptors determined in the PHE and calculated risk based on contaminant levels at the site due to incorporation of specified remedial actions

C.2 0 CONTAMINANTS OF CONCERN

The OU1 PHE (DOE 1994a) identified the future onsite adult resident receptor as having the highest potential risk values for the contaminants 1,1-dichloroethene (1,1 DCE), carbon tetrachloride (CCL_4) and tetrachloroethene (PCE). These risks were calculated assuming adequate groundwater present and available for receptor use. The total risk values in the PHE for 1,1 DCE, CCL_4 and PCE respectively are 3.8×10^{-2} , 2.5×10^{-3} and 1.1×10^{-3} with the dominating pathway being ingestion of groundwater for all three contaminants. The contaminants with the three highest calculated noncarcinogenic hazard indices (HI) in the PHE for the same receptor assuming use of groundwater are also 1,1 DCE, CCL_4 and PCE. These three contaminants also yielded the highest HIs for the future onsite residential child receptor and are of the same order of magnitude as the adult receptor. The three most dominating pathways for these contaminants are ingestion of groundwater, inhalation of volatiles and dermal contact with groundwater. These pathways are all driven by groundwater contamination and, therefore, this risk calculation focuses on groundwater associated pathways only. Groundwater modeling results are used to derive concentrations of contamination in groundwater at Woman Creek. By comparing initial modeling results with respective contaminant specific preliminary remediation goals (PRGs) for Rocky Flats (DOE 1994b), PCE was deemed the most conservative contaminant to use in this risk calculation. Detailed groundwater modeling results (refer to Appendix B) for PCE are used to calculate carcinogenic risk and noncarcinogenic HIs.

C 3 0 SCENARIOS AND PATHWAYS

Although onsite residences are not consistent with future land use plans a hypothetical future onsite resident exposure scenario is evaluated in this risk assessment. The future onsite resident is assumed to live within the OU1 study area boundary at the woman creek location. To use the most conservative scenario for direct ingestion of groundwater one of the future onsite resident scenarios assume that an adequate well water supply exists.

A future onsite worker assumed to be an office worker is also quantitatively evaluated in this risk assessment. The setting for the office worker is likely to have extensive paved areas and well maintained landscaping. It is assumed that municipal water would be supplied to the office building and therefore the future office worker will not directly access OU1 groundwater.

C 3 1 Exposure Pathways

This section discusses the potential release and transport of chemicals from OU1 and identifies exposure pathways by which the future onsite resident or future onsite office worker may potentially be exposed to site contaminants.

An exposure pathway describes a specific environmental pathway that can expose an individual to contaminants that are onsite or originate from a site. An exposure pathway includes five elements that must be present for an exposure pathway to be complete.

- Source of Chemicals
- Mechanism of Chemical Release
- Environmental Transport Medium
- Exposure Point
- Human Intake Route

An incomplete pathway means that no human exposure can occur. An exposure pathway is considered to be potentially complete and relevant if there are potential chemical release and transport mechanisms and receptors identified for that exposure pathway.

An exposure route is the pathway through which a contaminant enters or impacts an organism
There are four basic human exposure routes

- dermal absorption through contact with soil surface water or groundwater
- inhalation of volatile organic compounds (VOCs) or airborne particulates
- ingestion of soil surface water or groundwater
- external irradiation if radionuclides are present

Chemicals that volatilize from groundwater and/or site soils and are released to indoor air also represent a potentially complete inhalation pathway for the future onsite resident and office worker

As documented in the PHE the pathways that dominated the human health risk are associated with groundwater contamination Therefore the pathways considered in this risk assessment will only consider groundwater contamination associated with the potential remedial actions The following paragraphs describe the potential exposure pathways

Receptors that were quantitatively evaluated in the PHE were

- current offsite residents,
- future onsite residents,
- current onsite workers
- future onsite workers and
- future onsite ecological researcher

Of these potential receptors only the future onsite residents and the future onsite workers could be significantly exposed to contaminants in the groundwater Future onsite residents could be exposed to direct ingestion of groundwater dermal contact with groundwater and inhalation of volatiles that have diffused through the house foundation and from indoor use of groundwater such as showering Future onsite workers could be exposed to volatiles that have diffused through the office building foundation Since groundwater will not be used in an office building no direct exposure to groundwater is anticipated for the future onsite worker These two receptors and potential scenarios are considered conservative since neither receptor could be

exposed until the RFETS has been released for unrestricted use The remaining receptors evaluated in the PHE do not have significant exposure to groundwater and therefore were not evaluated in this risk assessment

C 3 1 1 Future Onsite Resident

Contaminants that volatilize from site groundwater and are released to indoor air through the house foundation represent a potentially complete inhalation pathway to future onsite residents Assuming that site groundwater is used within the household inhalation of VOCs from indoor water use represents another potentially complete inhalation pathway Inhalation of outdoor VOCs is considered insignificant due to expected dispersal and dilution of the VOCs

Assuming that site groundwater will be used within the future onsite residential household direct ingestion of groundwater contamination represents a potentially complete pathway Future onsite residents also could physically contact contaminated groundwater Therefore dermal absorption of contaminants from contact with contaminated groundwater represents a potentially complete pathway

The location of the groundwater contamination for the future onsite resident is assumed to be woman creek

C 3 1 2 Future Onsite Office Worker

Since the direct use of groundwater is not considered credible for this receptor the only remaining exposure pathway is volatilization of contaminants from site groundwater and release to indoor air through the office building foundation The inhalation pathway is then potentially complete for the future onsite office worker Similar to the future onsite resident scenario the inhalation of outdoor VOCs is considered incomplete due to expected dispersal and dilution of the VOCs As with the future onsite resident the location of the contamination for the future onsite office worker is assumed to be woman creek

C 4 0 EXPOSURE ASSESSMENT AND INTAKE EQUATIONS

Pathway specific exposures or intakes are quantified through the use of intake equations exposure parameters and exposure concentrations Intake equations are pathway specific while exposure parameters and exposure concentrations are scenario-specific and pathway specific Exposure concentrations for this risk assessment have been modeled using groundwater modeling techniques The generalized intake equations associated with each pathway and the non chemical specific parameters that are used in the equations are presented in this section

C 4 1 Ingestion of Water

Equation 1 was used to calculate direct ingestion or intake of contaminated water The ingestion rate was adjusted in accordance with the scenario

$$\text{Intake (mg/kg/day)} = \frac{\text{CW} \times \text{IR} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}} \quad (1)$$

where

- CW = Chemical concentration in water (mg/liter)
- IR = Ingestion rate (liter/day)
- EF = Exposure frequency (days/year)
- ED = Exposure duration (years)
- BW = Body weight (kg)
- AT = Averaging time (period over which exposure is averaged in days)

The chemical concentration in water is a modeled value and the modeling techniques are described in the PHE (DOE 1994a) Some parameters vary between adult and child receptors such as ingestion rates exposure durations and body weights The adult and child ingestion rates are 2 liters and 1 liter per day respectively Exposure frequency for residential receptors is 350 days/year The exposure durations for adult and child receptors are 24 and 6 years respectively The adult and child body weights are 70 kilograms and 15 kilograms, respectively The averaging time for a carcinogen is 25 550 days or 70 years

C 4 2 Dermal Contact With Water

Equation 2 was used to calculate absorbed dose through the skin or intake for the future onsite resident. This is the only receptor that potentially can contact contaminated groundwater. This equation calculates the actual absorbed dose, not the amount of chemical that comes in contact with the skin.

$$\text{Absorbed Dose (mg/kg/day)} = \frac{CW \times SA \times PC \times ET \times EF \times ED \times CF}{BW \times AT} \quad (2)$$

where

- CW = Chemical concentration in water (mg/liter)
- SA = Skin surface area available for contact (cm²)
- PC = Chemical specific dermal permeability constant (cm/hr)
- ET = Exposure time (hours/day)
- EF = Exposure frequency (days/year)
- ED = Exposure duration (years)
- CF = Volumetric conversion factor for water (1 liter/1000 cm³)
- BW = Body weight (kg)
- AT = Averaging time (period over which exposure is averaged in days)

The chemical concentration in water is a modeled value as described in the PHE. Some parameters vary between adult and child receptors, such as skin surface areas, exposure durations, and body weights. The adult and child skin surface areas are 23,200 cm² and 9,180 cm², respectively. The dermal permeability constants are chemical specific and their origination is discussed in the PHE. Adult and child exposure times for dermal contact with groundwater are 0.2 hours/day. Exposure frequency for a residential adult and child is 350 days/year. Adult and child exposure durations are 24 years and 6 years, respectively. The volumetric conversion factor for water is 0.001 liters/cm³. Adult and child body weights are 70 kilograms and 15 kilograms, respectively. The averaging time for a carcinogen is 25,550 days or 70 years.

C 4 3 Inhalation of Airborne Contaminants

Exposure scenarios involving the residential adult residential child and office worker include intake of airborne contaminants The contaminants are in the vapor phase and originate from groundwater contaminants volatilizing and diffusing through either a home foundation or office building foundation as applicable Assuming well water is used within the home the residential receptor can also inhale contaminants volatilized during in home water use Dermal absorption of vapor phase contaminants is considered to be a negligible portion of inhalation intakes and therefore is disregarded in accordance with Risk Assessment Guidance for Superfund (RAGS) (EPA 1991) Equation 3 was used to calculate inhalation intakes for residential and office worker receptors

$$\text{Intake (mg/kg/day)} = \frac{\text{CA} \times \text{IR} \times \text{EF} \times \text{ED}}{\text{BW} \times \text{AT}} \quad (3)$$

where

- CA = Contaminant concentration in air (mg/m³)
- IR = Inhalation rate (m³/day)
- EF = Exposure frequency (days/year)
- ED = Exposure duration (years)
- BW = Body weight (kg)
- AT = Averaging time (period over which exposure is averaged in days)

Both residential and office worker receptors have the potential to inhale volatilized contamination that has diffused through the foundation of either a home or an office building, as applicable It is assumed that groundwater would not service onsite office buildings, therefore only a residential receptor could inhale volatilized contamination due to indoor water use The chemical concentrations in indoor air (volatilized through a foundation and volatilized due to indoor water use) are modeled values as described in the PHE Some parameters vary between the onsite office worker adult and child receptors such as inhalation rates exposure frequencies exposure durations, body weights, and averaging times The inhalation rate is 15 m³/day for a residential adult (assuming indoor activities) and 20 m³/day for both a residential child and office worker The exposure frequency is 350 days/year for a residential adult and

child and 250 days/year for an office worker The exposure duration is 24 years for a residential adult 6 years for a residential child and 25 years for an office worker The body weight is 70 kilograms for a residential adult and office worker and 15 kilograms for a residential child

C 5 0 TOXICITY ASSESSMENT

This section provides the toxicity constants used for risk characterization purposes and summarizes toxicological information. Specific derivation of toxicity constants and respective sources is discussed in the PHE. For this risk assessment, toxicity information summarized for two categories of potential effects: noncarcinogenic and carcinogenic effects. These two categories were selected because of the slightly differing methodologies for estimating potential health risks associated with exposures to carcinogens and noncarcinogens.

C 5 1 Tetrachloroethene

Tetrachloroethene, also known as perchloroethylene (PCE), has widespread use in the dry cleaning and textile industries. It is also used in the cold cleaning and vapor degreasing of metals, as a chemical intermediate in the synthesis of fluorocarbons, as a component of aerosol laundry treatment products, as a solvent for silicones, as the insulating fluid and cooling gas in electrical transformers, and in typewriter correction fluid. PCE is not known to occur naturally but contributes to water pollution through leaching from vinyl liners in asbestos-cement water pipelines and as wastewater from metal finishing, laundries, aluminum-forming, organic chemical/plastics manufacturing, and municipal treatment plants. Air contamination is the result of emissions and vaporization losses from dry cleaning and industrial metal cleaning (ATSDR 1992).

The effects discussed below are due to occupational exposure levels which are much higher than the expected environmental levels. Primarily, exposure occurs through inhalation of contaminated air or ingestion of contaminated water. PCE can cause lightheadedness, dizziness, euphoria, blindness, cardiac arrhythmias, hypotension, cyanosis, respiratory depression, pulmonary hemorrhages, and central nervous system (CNS) depression in acute dosages. When chronically dosed, trigeminal nerve impairment, liver injury, and chapped skin can occur. PCE is metabolized and excreted very slowly. Individuals with diseases of the heart, liver, kidneys,

and lungs are the most vulnerable to PCE poisoning. It has also been known to cause jaundice in newborns from PCE excretion in the breast milk (ATSDR 1992)

Historically few acute or chronic industrial toxicity problems have arisen from the use of this solvent although researchers have reported both hepatotoxicity and CNS effects. Ingested or inhaled PCE is mostly excreted by the lungs. The metabolism of PCE is very slow a very low percentage is excreted in the urine as metabolites. Currently no inhalation RfD is available for PCE. Oral RfDs have been calculated based on research with rodents. Primary effects associated with PCE exposure include liver and kidney damage and CNS depression. The oral RfD for chronic exposures is $1\text{E-}2$ mg/kg/day with an uncertainty factor of 1000. There is medium confidence in this RfD because no one study combined the features required for deriving a high confidence RfD. Confidence in the principle study is low because it lacked complete histopathological examination at the no observed adverse effect level (NOAEL) and corroborative studies on its teratogenic and reproductive impacts are lacking (EPA 1994)

PCE is listed as a probable group B2 carcinogen in IRIS has an oral SF of $5.2\text{E-}2$ and an inhalation SF of $2.03\text{E-}3$. This classification was based on studies performed on rodents where inhalation produced both leukemia and tumors of the liver. PCE is for the most part nonmutagenic and has not been shown to cause reproductive toxicity.

Table C 5 1 summarizes chemical specific constants for PCE

C 5 2 Concentrations of Contamination

Groundwater modeling was used to calculate the expected contamination in groundwater at various locations downgradient of IHSS 119 1. The concentrations were modeled to include the specific remediation scenarios starting in 1969 and continuing in time steps. The scenarios that were modeled are no action, continued use of the french drain and extraction well (institutional controls) and remediating the contamination at the source (remediation). The no action scenario was modeled out to the year 2369 (400 years) the continued operation of the french drain and

**Table C 5-1
Chemical Specific Constants**

Chemical	Tetrachloroethene
Metal or Organic	Organic
Weight of Evidence	B2
SF Ingestion (mg/kg/day)	5 20E 2
SFi Inhalation (mg/kg/day)	2 03E 3
Target System	Liver/Hepatic Lesions
RfD Ingestion (mg/kg/day)	1 0E 2
RfD Inhalation (mg/kg/day)	n/a
Dermal Permeability (cm/hr)	4 80E-02
Additional Notes	RfD Inhalation no data

extraction well scenario was modeled to year 2269 (300 years) and the remediation scenario was modeled to year 2169 (200 years). The concentrations of PCE at the end of the modeling runs for the no action and continued operation of the french drain and extraction well scenarios were still rising slightly however the peak is expected to occur within a short time frame and at a concentration that is not significantly higher than the last concentration result. Therefore the highest concentrations of PCE for these scenarios was conservatively used to calculate carcinogenic risk and noncarcinogenic hazard effects. The highest concentration of PCE at woman creek for the remediation scenario occurred during the year 2152. Therefore the concentration for this scenario is assumed to be the 30-year average concentration centered around the year 2152. The calculated groundwater concentrations were then used in the Johnson and Ettinger (1991) soil gas model which considers chemical specific parameters such as Henry's law constant and air diffusion coefficients to calculate a vapor concentration inside a building refer to the PHE for further details. To calculate the concentration in indoor air from groundwater use the conservatively modeled groundwater concentrations were multiplied by the volatilization fraction of 0.065 mg/m³ air per mg/l water. This conservative approach is consistent with Andelman (1990) and is discussed further in the PHE. The concentrations of PCE and associated scenarios are summarized in Table C 5 2.

C 5 3 Contaminant Intakes

The intake equations discussed in section 4 0 use the nonchemical specific parameters chemical specific parameters chemical concentrations and appropriate scenarios to calculate respective chemical intakes. Tables C 5 3 through C 5 8 summarize the carcinogenic and noncarcinogenic intakes by scenario receptor and pathway.

C 5 4 Risk and Hazard Quotient Calculation

Potential carcinogenic risks are expressed as an estimated probability of an individual developing cancer from lifetime exposure to the carcinogen. This probability is based on projected intakes and chemical specific dose-response data called cancer slope factors (SFs). Cancer SFs and the

**Table C 5-2
PCE Concentrations at Woman Creek**

Scenario	Indoor Air Volatiles Diffusing through the Foundation (mg/m³)	Groundwater (mg/l)	Indoor Air from Groundwater Use (mg/m³)
Unchanged contamination discontinued FD and extraction well operations	9 69E 10	3 60E-03	2 34E-04
Unchanged contamination continued FD and extraction well operations	2 32E 12	8 62E-06	5 60E-07
Remediated contamination discontinued FD and extraction well operations	1 57E 10	5 84E-04	3 80E-05

Table C 5 3
Carcinogenic Intakes at Woman Creek, No Action Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Resident with Groundwater	6 83E 11	3 38E-05	3 77E-06	1 65E-05
Future Onsite Resident without Groundwater	6 83E 11	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	6 77E 11	N/A	N/A	N/A

Table C 5-4
Carcinogenic Intakes at Woman Creek, French Drain and Extraction Well Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Resident with Groundwater	1.63E-13	8.10E-08	9.02E-09	3.95E-08
Future Onsite Resident without Groundwater	1.63E-13	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	1.62E-13	N/A	N/A	N/A

Table C 5 5
Carcinogenic Intakes at Woman Creek, Source Remediation Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Resident with Groundwater	1 11E 11	5 49E-06	6 11E-07	2 67E-06
Future Onsite Resident without Groundwater	1 11E 11	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	1 10E 11	N/A	N/A	N/A

Table C 5 6
Noncarcinogenic Intakes at Woman Creek, No Action Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Adult Resident with Groundwater	N/A	9 86E-05	1 10E-05	N/A
Future Onsite Child Resident with Groundwater	N/A	2 30E-04	2 03E-05	N/A
Future Onsite Adult Resident without Groundwater	1 99E 10	N/A	N/A	N/A
Future Onsite Child Resident without Groundwater	1 24E-09	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	1 90E 10	N/A	N/A	N/A

Table C 5-7
Noncarcinogenic Intakes at Woman Creek, French Drain and Extraction Well Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Adult Resident with Groundwater	N/A	2.36E-07	2.63E-08	N/A
Future Onsite Child Resident with Groundwater	N/A	5.51E-07	4.86E-08	N/A
Future Onsite Adult Resident without Groundwater	4.77E-13	N/A	N/A	N/A
Future Onsite Child Resident without Groundwater	2.97E-12	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	4.54E-13	N/A	N/A	N/A

Table C 5 8
Noncarcinogenic Intakes at Woman Creek, Source Remediation Scenario
(mg/kg/day)

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater
Future Onsite Adult Resident with Groundwater	N/A	1 60E-05	1 78E-06	N/A
Future Onsite Child Resident with Groundwater	N/A	3 73E-05	3 29E-06	N/A
Future Onsite Adult Resident without Groundwater	3 23E 11	N/A	N/A	N/A
Future Onsite Child Resident without Groundwater	2 01E 10	N/A	N/A	N/A
Future Onsite Office Worker without Groundwater	3 07E 11	N/A	N/A	N/A

estimated daily intake of a compound averaged over a lifetime of exposure is used to estimate the incremental risk that an individual exposed to that compound may develop cancer. Potential carcinogenic risks are estimated from the following equation:

$$\text{Risk} = \text{Intake} \times \text{SF} \quad (4)$$

where

Risk	=	Potential lifetime excess cancer risk (unitless)
SF	=	Slope factor for chemicals (mg/kg/day) ⁻¹
Intake	=	Chemical intake (mg/kg/day)

Potential health effects of chronic exposure to noncarcinogenic compounds is assessed by calculating a hazard quotient (HQ) which is derived by dividing the estimated daily intake by a chemical specific RfD as shown in the following equation:

$$\text{HQ} = \text{Intake} / \text{RfD} \quad (5)$$

where

HQ	=	Noncancer hazard quotient (unitless)
Intake	=	Chemical intake (mg/kg/day)
RfD	=	Reference dose (mg/kg/day)

A HQ greater than 1.0 indicates that exposure to that contaminant, (at the concentrations and for the duration and frequencies of exposure estimated in the exposure assessment) may cause adverse health effects in exposed populations. However, the level of concern associated with exposure to noncarcinogenic compounds does not increase linearly as HQ values exceed 1.0. In other words, HQ values do not represent a probability or a percentage. For example, an HQ of 10 does not indicate that adverse health effects are 10 times more likely to occur than an HQ value of 1.0, but that potential adverse health effects are of greater concern.

C 6 0 RISK CHARACTERIZATION

Risk characterization involves estimating the magnitude of potential adverse effects summarizing the nature of the threats to public health and considering the nature and weight of evidence supporting these risk estimates and the degree of uncertainty surrounding the estimates. Specifically, risk characterization involves combining the results of the exposure and toxicity assessments to provide numerical estimates of health risk. These estimates are comparisons of exposure levels with appropriate RfDs or estimates of the lifetime cancer risk with a given intake.

Generally, to quantify the health risks, the intakes are first calculated as identified in section 4 0 for each applicable scenario. The intakes were calculated from the concentrations discussed in section 5 2 and the methodology documented in the EPA RAGS (1989). The specific intakes calculated in section 5 3 were then compared to the applicable chemical specific toxicological data presented in section 5 1 to determine the health risk.

The health risks from PCE were calculated to determine potential carcinogenic and noncarcinogenic effects as discussed in Sections 6 1 and 6 2 respectively.

C 6 1 Carcinogenic Effects

Carcinogenic risks from exposure to PCE were calculated for a future onsite resident using groundwater, using public water, and for a future onsite office worker using public water. The source of contamination considered (1) maintaining the current groundwater contamination level and removing the french drain and extraction well, (2) maintaining the current groundwater contamination level and continuing the french drain and extraction well operations, and (3) remediating the contamination source and removing the french drain and extraction well. These receptors and scenarios considered PCE contamination at woman creek. Tables C 6-1 through C 6-3 summarize the results of the risk calculations by scenario, receptor, and pathway.

The three highest carcinogenic risks at woman creek are associated with the future onsite resident using groundwater for household use. The risks for the future onsite resident without groundwater and the future office worker without groundwater are negligible (in the 10^{-13} to 10^{-16} range)

The scenario that yielded the maximum calculated carcinogenic risk assumed current PCE groundwater contamination and removal of the french drain and extraction well (no action scenario). The total calculated risk for the future onsite resident with this exposure is 1.99×10^{-6} with the dominating pathway of ingestion of groundwater with a risk of 1.76×10^{-6} (see Table C 6-1)

The next highest calculated carcinogenic risk assumed remediation of the contamination and discontinuing the operation of the french drain and extraction well. The total calculated risk for the future on site resident with this exposure is 3.22×10^{-7} with the dominating pathway of ingestion of groundwater with a risk of 2.85×10^{-7} (see Table C 6-3)

The third highest calculated carcinogenic risk assumed current PCE groundwater contamination and continued operation of the french drain and the extraction well (Institutional controls). The total calculated risk for the future on site resident with this exposure is 4.76×10^{-9} with the dominating pathway of ingestion of groundwater with a risk of 4.21×10^{-9} (see Table C 6-2)

C 6 2 Noncarcinogenic Effects

The receptors and pathways used to evaluate carcinogenic effects were also used to evaluate noncarcinogenic effects. The hazard indices for PCE are the summed HQs for each exposure pathway. If the hazard index exceeds unity there may be a concern for potential health effects and the exposure should be evaluated more closely. Tables C 6-4 through C 6-6 summarize the results of the HQ and hazard indices calculations by scenario, receptor, and pathway.

Table C 6-1
Carcinogenic Risk at Woman Creek, No Action Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Resident with Groundwater	1 39E 13	1 76E-06	1 96E-07	3 35E-08	1 99E-06
Future On site Resident without Groundwater	1 39E 13	N/A	N/A	N/A	1 39E 13
Future On site Office Worker without Groundwater	1 37E 13	N/A	N/A	N/A	1 37E 13

Table C 6-2
Carcinogenic Risk at Woman Creek, French Drain and Extraction Well Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Resident with Groundwater	3 32E 16	4 21E-09	4 69E 10	8 01E 11	4 76E-09
Future On site Resident without Groundwater	3 32E 16	N/A	N/A	N/A	3 32E 16
Future On site Office Worker without Groundwater	3 29E 16	N/A	N/A	N/A	3 29E 16

Table C 6-3
Carcinogenic Risk at Woman Creek, Source Remediation Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Resident with Groundwater	2.25E-14	2.85E-07	3.18E-08	5.43E-09	3.22E-07
Future On site Resident without Groundwater	2.25E-14	N/A	N/A	N/A	2.25E-14
Future On site Office Worker without Groundwater	2.23E-14	N/A	N/A	N/A	2.23E-14

Table C 6-4
Noncarcinogenic Hazard Indices at Woman Creek, No Action Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Adult Resident with Groundwater	N/A	9 86E-03	1 10E-03	N/A	1 10E-02
Future On site Child Resident with Groundwater	N/A	2 30E-02	2 03E-03	N/A	2 50E-02
Future On site Adult Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Child Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Office Worker without Groundwater	N/A	N/A	N/A	N/A	N/A

Table C 6-5
Noncarcinogenic Hazard Indices at Woman Creek, French Drain and
Extraction Well Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Adult Resident with Groundwater	N/A	2.36E-05	2.63E-06	N/A	2.62E-05
Future On site Child Resident with Groundwater	N/A	5.51E-05	4.86E-06	N/A	6.00E-05
Future On site Adult Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Child Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Office Worker without Groundwater	N/A	N/A	N/A	N/A	N/A

Table C 6-6
Noncarcinogenic Hazard Indices at Women Creek, Source Remediation Scenario

Receptor	Inhalation of Volatiles Diffusing through Foundation	Ingestion of Groundwater	Dermal Contact with Groundwater	Inhalation of Volatiles from Indoor use of Groundwater	TOTAL
Future On site Adult Resident with Groundwater	N/A	1 60E-03	1 78E-04	N/A	1 78E-03
Future On site Child Resident with Groundwater	N/A	3 73E-03	3 29E-04	N/A	4 06E-03
Future On site Adult Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Child Resident without Groundwater	N/A	N/A	N/A	N/A	N/A
Future On site Office Worker without Groundwater	N/A	N/A	N/A	N/A	N/A

The calculation of HQs and respective hazard indices did not yield a significant noncarcinogenic hazard (i.e. did not approach unity). The highest hazard index is 2.50×10^{-2} for a future onsite child resident with groundwater assuming PCE contamination and discontinuing the french drain and extraction well operations (no action scenario) (see Table C-6-4). The dominating pathway for this receptor is ingestion of groundwater with a HQ of 2.30×10^{-2} . The remaining hazard indices ranged from 1.10×10^{-2} to 2.62×10^{-5} . HQs were not calculated for receptors that do not have access to groundwater because the only applicable pathway for these receptors is inhalation of volatiles diffusing through the foundation and the PCE inhalation RfD is not available.

C 7 0 SUMMARY

These residual risk calculations discussed in this risk assessment were intended to develop a quantitative assessment of the risk associated with appropriate receptors and scenarios after specific remedial action alternatives have been implemented. Based on information from the PHE, the most conservative contamination scenarios, receptors, and pathways were evaluated. Concentrations of contaminants were modeled using groundwater modeling techniques and then receptor intakes were calculated. The intakes were combined with toxicological data in risk and HQ equations to calculate potential probabilities for carcinogenic risk and noncarcinogenic HQs. The carcinogenic risks and hazard quotients were then summed by scenario to yield total potential carcinogenic and noncarcinogenic effects.

The maximum calculated carcinogenic risk is for the no action scenario. The total risk to the future onsite resident with groundwater is 1.99×10^{-6} .

The hazard indices calculated for the scenarios and receptors were not significant (i.e., did not approach unity). The maximum hazard index is 2.50×10^{-2} for a future onsite child resident with groundwater, assuming the current levels of PCE contamination, and discontinuing operations of the french drain and extraction well.

C 8 0 REFERENCES

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APPENDIX D

POTENTIAL APPLICABLE OR RELEVANT AND APPROPRIATE REQUIREMENTS (ARARs)

Table E-1 Alternative 0- No Action

Activity	Resource Description	QTY	Unit	Base Costs Per Unit				Base Costs				Total Costs	
				Mat'l	Equip.	Labor	Sub-contract	Mat'l	Equip.	Labor	Sub-contract		
Direct Capital Costs													
Decommission French Drains	Backhoe		1 day		\$2,200.00	\$302.00			\$0	\$2,200	\$302	\$0	\$2,502
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam. @ 20'/depth	4	ea				\$26,000.00		\$0	\$0	\$0	\$100,000	\$100,000
Field Personnel	Surveyor	8	hr			\$60.00			\$0	\$0	\$480	\$0	\$480
	Health & Safety	8	hr			\$68.00			\$0	\$0	\$520	\$0	\$520
	Geologist	40	hr			\$66.00			\$0	\$0	\$2,200	\$0	\$2,200
Subtotal Direct Capital Costs													
Indirect Capital Costs													
Misc. Labor & Materials	10% of direct labor & \$1.50 in materials cost for each direct labor hour								\$0	\$2,200	\$3,482	\$0	\$106,682
Permits	6% of direct materials, equipment, & labor								\$64	\$0	\$348	\$0	\$432
Construction Management	10% of direct materials, equipment, & labor								\$0	\$110	\$174	\$0	\$284
Project Management	10% of direct materials, equipment, & labor								\$0	\$220	\$348	\$0	\$568
Overhead, Profit & Bond	26.3% of direct materials, equipment, & labor								\$0	\$220	\$348	\$0	\$568
Subcontractor Fee	10% of subcontractor costs								\$0	\$557	\$881	\$0	\$1,438
Subtotal Indirect Capital Costs													
Contingency	30% of direct and indirect capital costs								\$64	\$1,107	\$2,100	\$10,000	\$13,269
Total Capital Costs													
									\$109	\$4,299	\$7,256	\$149,000	\$154,684

Annual O&M Direct Costs												
Subtotal O&M Direct Costs												
Total O&M Costs												

Annual Post Closure Direct Costs												
Soil Sampling	Collect Groundwater Samples	15	ea				\$1,500.00	\$0	\$0	\$0	\$18,000	\$18,000
Analytical Work	Sample Analysis for VOCs & Inorganics	14	ea				\$4,100.00	\$0	\$0	\$0	\$57,400	\$57,400
Subtotal Post Closure Direct Costs												
Annual Post Closure Indirect Costs												
Project Management	10% of post closure direct materials, equipment, & labor costs							\$0	\$0	\$0	\$0	\$0
Subcontractor Fee	10% of post closure subcontractor costs							\$0	\$0	\$0	\$7,540	\$7,540
Subtotal Post Closure Indirect Costs												
Contingency	30% of total post closure direct and indirect costs							\$0	\$0	\$0	\$24,882	\$24,882
Total Annual Post Closure Costs												
								\$0	\$0	\$0	\$107,822	\$107,822
Total Post Closure Costs (30 Yrs @ 5% discount rate)												
								\$0	\$0	\$0	\$1,740,363	\$1,740,363

Total Cost Of Alternative												
								\$109	\$4,299	\$7,256	\$1,883,363	\$1,895,027

Table E-2 Alternative 1 Institutional Controls Without the French Drain

Activity	Resource Description	Qty.	Unit	Bare Costs Per Unit			Source	Bare Costs			Sub-contract	Total Costs
				Mat'l	Equip.	Labor		Equip.	Labor			
Direct Capital Costs												
Decommission French Drain	Backhoe	1	day		\$2,500.00	\$382.00	Means Ref.	\$0	\$2,500	\$382	\$0	\$2,882
Drill Monitoring Wells	Drill & Case 4 walls, 6" diam. & 30' depth	4	ea			\$25,000.00	Vendor Quote	\$0	\$0	\$0	\$100,000	\$100,000
Field Personnel	Surveyor	8	hr			\$50.00	Prof. Judgment	\$0	\$0	\$400	\$0	\$400
	Health & Safety	8	hr			\$65.00	Prof. Judgment	\$0	\$0	\$520	\$0	\$520
	Geologist	48	hr			\$65.00	Prof. Judgment	\$0	\$0	\$3,200	\$0	\$3,200
Subtotal Direct Capital Costs												
Indirect Capital Costs												
Misc. Labor & Materials	10% of direct labor & \$1.50 in materials cost for each direct labor hour						Facil. Eng. 009	\$04	\$0	\$348	\$0	\$432
Permits	5% of direct materials, equipment, & labor						Prof. Judgment	\$0	\$110	\$174	\$0	\$284
Construction Management	10% of direct materials, equipment, & labor						Prof. Judgment	\$0	\$250	\$348	\$0	\$598
Project Management	10% of direct materials, equipment, & labor						EO&G Cost Ref.	\$0	\$250	\$348	\$0	\$598
Overhead, Profit & Bond	35.9% of direct materials, equipment, & labor						Facil. Eng. 009	\$0	\$457	\$681	\$0	\$1,138
Subcontractor Fee	10% of subcontractor costs						Facil. Eng. 009	\$0	\$0	\$0	\$10,000	\$10,000
Subtotal Indirect Capital Costs												
								\$044	\$1,107	\$2,100	\$10,000	\$13,250
Contingency	30% of direct and indirect capital costs						Facil. Eng. 009	\$35	\$403	\$1,874	\$33,000	\$35,312
Total Capital Costs												
								\$109	\$4,309	\$7,356	\$143,000	\$154,664

Annual O&M Direct Costs												
Subtotal O&M Direct Costs								\$0	\$0	\$0	\$0	\$0
Total O&M Costs								\$0	\$0	\$0	\$0	\$0

Annual Post Closure Direct Costs												
Soil Sampling	Groundwater Samples	15	ea			\$1,500.00	Prof. Judgment	\$0	\$0	\$0	\$18,000	\$19,500
Analytical Work	Sample Analysis for VOCs & Inorganics	14	ea			\$4,100.00	Vendor Quote	\$0	\$0	\$0	\$57,400	\$61,500
Subtotal Post Closure Direct Costs								\$0	\$0	\$0	\$75,400	\$75,400
Annual Post Closure Indirect Costs												
Project Management	10% of post closure direct materials, equipment, & labor costs						EO&G Cost Ref.	\$0	\$0	\$0	\$0	\$0
Subcontractor Fee	10% of post closure subcontractor costs						Facil. Eng. 009	\$0	\$0	\$0	\$7,540	\$7,540
Subtotal Post Closure Indirect Costs								\$0	\$0	\$0	\$7,540	\$7,540
Contingency	30% of total post closure direct and indirect costs						Facil. Eng. 009	\$0	\$0	\$0	\$24,863	\$24,863
Total Annual Post Closure Costs								\$0	\$0	\$0	\$107,803	\$107,803
Total Post Closure Costs (50 yrs @ 8% discount rate)								\$0	\$0	\$0	\$1,740,363	\$1,740,363

Total Cost Of Alternative								\$109	\$4,309	\$7,356	\$1,893,363	\$1,896,027
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Table E-3 Alternative 2 Institutional Controls With the French Drain

Activity	Resource Description	Qty.	Unit	Base Costs Per Unit				Base Costs				Total Costs
				Mat'l	Equip.	Labor	Sub- contract	Mat'l	Equip.	Labor	Sub- contract	
Direct Capital Costs												
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam. & 20' depth	4	ea				\$25,000.00			\$0	\$0	\$100,000
Field Personnel	Surveyor	6	hr			\$60.00				\$400	\$0	\$400
	Health & Safety	6	hr			\$65.00				\$390	\$0	\$390
	Geologist	40	hr			\$65.00				\$2,600	\$0	\$2,600
Subtotal Direct Capital Costs												
										\$0	\$3,120	\$103,120
Indirect Capital Costs												
Min. Labor & Materials	10% of direct labor & \$1.50 in materials cost for each direct labor hour							\$44		\$312	\$0	\$356
Permits	5% of direct materials, equipment, & labor							\$0		\$156	\$0	\$156
Construction Management	10% of direct materials, equipment, & labor							\$0		\$312	\$0	\$312
Project Management	10% of direct materials, equipment, & labor							\$0		\$312	\$0	\$312
Overhead, Profit & Bond	25.2% of direct materials, equipment, & labor							\$0		\$789	\$0	\$789
Subcontractor Fee	10% of subcontractor costs							\$0		\$10,000	\$0	\$10,000
Subtotal Indirect Capital Costs												
								\$44		\$1,881	\$0	\$11,881
Contingency	30% of direct and indirect capital costs							\$35		\$1,500	\$0	\$34,536
Total Capital Costs												
								\$109		\$6,502	\$0	\$149,611
Annual O&M Direct Costs												
Groundwater Treatment	UV/Peroxide AOX Treatment System	1	yr				\$676,000.00	(1)		\$0	\$0	\$676,000
Subtotal O&M Direct Costs												
										\$0	\$0	\$676,000
Annual O&M Indirect Costs												
Min. Labor & Materials	10% of direct labor & \$1.50 in material cost for each direct labor hour									\$0	\$0	\$0
Project Management	10% of direct materials, equipment, & labor costs									\$0	\$0	\$0
Overhead, Profit & Bond	25.2% of direct materials, equipment, & labor costs									\$0	\$0	\$0
Subcontractor Fee	10% of subcontractor costs									\$0	\$67,000	\$67,000
Subtotal O&M Indirect Costs												
										\$0	\$67,000	\$67,000
Contingency	30% of total direct and indirect O&M costs									\$0	\$203,000	\$233,000
Total Annual O&M Cost												
										\$0	\$269,000	\$269,000
Total O&M Costs (\$30,770 @ 8% discount rate)												
										\$0	\$15,003,263	\$15,003,263
Annual Post Closure Direct Costs												
Soil/Sediment Sampling	Collect Groundwater Samples	15	ea				\$1,500.00			\$0	\$0	\$15,000
Analytical Work	Sample Analysis for VOCs & Inorganics	16	ea				\$4,100.00			\$0	\$0	\$67,400
Subtotal Post Closure Direct Costs												
										\$0	\$0	\$75,400
Annual Post Closure Indirect Costs												
Project Management	10% of post closure direct materials, equipment, & labor costs									\$0	\$0	\$0
Subcontractor Fee	10% of post closure subcontractor costs									\$0	\$7,540	\$7,540
Subtotal Post Closure Indirect Costs												
										\$0	\$7,540	\$7,540
Contingency	30% of total post closure direct and indirect costs									\$0	\$24,882	\$24,882
Total Annual Post Closure Costs												
										\$0	\$30,422	\$30,422

Activity	Bioscience Investigations	OSY	Total	Bioscience Costs Per Unit				Bioscience Costs				Total Costs
				Materials	Equipment	Labor	Sub-contract	Materials	Equipment	Labor	Sub-contract	
Total Post Closure Costs (30 yrs @ 5% discount rate)												
								\$0	\$0	\$0	\$0	\$1,740,963

Total Cost Of Alternative												
								\$108	\$0	\$6,502	\$17,496,616	\$17,496,616

(1) Costs represent annual operating costs as presented in the Phase I Preliminary Plan For Future Utilization of Existing Water Treatment Facilities At Rocky Flats Plant, Draft Report (June 15 1994)

Table E-4 Alternative 3 Modified French Drain With Additional Extraction Wells

Activity	Resource Description	Qty.	Unit	Base Costs Per Unit				Base Costs				Total Costs
				Mat'l	Equip.	Labor	Sub-contractor	Mat'l	Equip.	Labor	Sub-contractor	
Direct Capital Costs												
Additional Extraction Wells	Drill 4 Extraction Wells, 6" diam, 30 ft depth	4	ea				\$25,000.00		\$0	\$0	\$100,000	\$100,000
	Sump Pumps	4	ea		\$300.00				\$0	\$1,200	\$0	\$1,200
	PVC Piping to French Drain Sump, 2.5"	800	lf	\$1.70		\$2.19			\$1,360	\$0	\$1,752	\$3,112
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam, & 30' depth	4	ea				\$65,000.00		\$0	\$0	\$100,000	\$100,000
	Field Personnel								\$0	\$0	\$0	\$0
	Health & Safety	16	hr			\$60.00			\$0	\$960	\$0	\$960
		12	hr			\$68.00			\$0	\$816	\$0	\$816
	Geologist	68	hr			\$66.00			\$0	\$3,300	\$0	\$3,300
Subtotal Direct Capital Costs								\$1,360	\$1,200	\$6,053	\$200,000	\$208,193
Indirect Capital Costs												
Misc. Labor & Materials	10% of direct labor & \$1.00 in materials cost for each direct labor hour											
	Profits	Prof. Judgment										
	Construction Management	Prof. Judgment										
	Project Management	Prof. Judgment										
	Overhead, Profit & Bond	BLAG Cost Est.										
	Subcontractor Fee	Facil. Eng. 000										
		Facil. Eng. 000										
Subtotal Indirect Capital Costs								\$0	\$0	\$0	\$20,000	\$20,000
Contingency								\$0	\$0	\$0	\$20,000	\$20,419
											\$0	\$0
											\$0	\$0
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Annual O&M Direct Costs	UV/Peroxide & IX Treatment System	1	yr					\$0	\$0	\$0	\$0	\$0
Groundwater Treatment												
Subtotal O&M Direct Costs								\$0	\$0	\$0	\$0	\$0
Annual O&M Indirect Costs												
Misc. Labor & Materials												
10% of direct labor & \$1.00 in material cost for each direct labor hour								\$0	\$0	\$0	\$0	\$0
Profits												
5% of direct materials, equipment, & labor costs								\$0	\$0	\$0	\$0	\$0
Construction Management												
10% of direct materials, equipment, & labor costs								\$0	\$0	\$0	\$0	\$0
Project Management												
10% of direct materials, equipment, & labor costs								\$0	\$0	\$0	\$0	\$0
Overhead, Profit & Bond												
25.3% of direct materials, equipment, & labor costs								\$0	\$0	\$0	\$0	\$0
Subcontractor Fee												
10% of subcontractor costs								\$0	\$0	\$0	\$0	\$0
Subtotal O&M Indirect Costs								\$0	\$0	\$0	\$0	\$0
Contingency												
30% of total direct and indirect O&M costs								\$0	\$0	\$0	\$0	\$0
Total Annual O&M Cost								\$0	\$0	\$0	\$0	\$0
Total O&M Costs (50 yrs @ 5% discount rate)								\$0	\$0	\$0	\$0	\$0

Annual Post Closure Direct Costs	Collect Groundwater Samples	13	ea					\$0	\$0	\$0	\$0	\$0
Soil Sampling												
Analytical Work	Sample Analysis for VOCs & Inorganics	14	ea					\$0	\$0	\$0	\$0	\$0
Subtotal Post Closure Direct Costs								\$0	\$0	\$0	\$0	\$0
Annual Post Closure Indirect Costs												
Project Management												
10% of post closure direct materials, equipment, & labor costs								\$0	\$0	\$0	\$0	\$0
Subcontractor Fee												
10% of post closure subcontractor costs								\$0	\$0	\$0	\$0	\$0
Subtotal Post Closure Indirect Costs								\$0	\$0	\$0	\$0	\$0
Total Post Closure Costs								\$0	\$0	\$0	\$0	\$0

Activity	Measure Description	Qty	Base Costs Per Unit					Base Costs				
			Mat	Equip	Labor	Sub- contract	Seismic	Mat	Equip	Labor	Sub- contract	Total Costs
Contingency	50% of total post closure direct and indirect costs						Feed. Eng. 000	00	00	00	00	\$24,832
Total Annual Post Closure Costs								00	00	00	\$107,832	\$107,832
Total Post Closure Costs (\$30.775 @ 50% discount rate)								00	00	00	\$1,740,363	\$1,740,363

Total Cost Of Alternative			\$2,838	\$2,945	\$13,820	\$17,639,616	\$17,649,610
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(1) Costs represent annual operating costs as presented in the Phase I Preliminary Plan For Future Utilization of Existing Water Treatment Facilities At Rocky Flats Plant, Draft Report (June 16 1994)

Table E-5 Alternative 4 Groundwater Pumping With Soil Vapor Extraction

Activity	Resource Description	Qty	Unit	Direct Costs (Per Unit)				Sub-contract	Indirect Costs				Total Costs
				Mat'l	Equip.	Labor	Sub-		Mat'l	Equip.	Labor	Sub-contract	
Direct Capital Costs													
Soil Gas Survey	Geologist	120	hr			\$65.00			Prof. Judgment	\$0	\$0	\$0	\$6,000
	Field Technician	80	hr			\$35.00			Prof. Judgment	\$0	\$0	\$0	\$2,800
	Portable GC	2	wk		\$1,000.00				Prof. Judgment	\$0	\$0	\$0	\$2,000
	Probe, Pump, and Misc. Equipment	100	ea		\$116.00				Prof. Judgment	\$0	\$1,500	\$0	\$11,500
Dewatering	Drill Extraction Wells, 6" diam, 20 ft depth	2	ea			\$25,000.00			Vendor Quote	\$0	\$0	\$0	\$50,000
	10 gpm submersible pumps	2	ea						Vendor Quote	\$0	\$600	\$0	\$600
	PVC Piping to French Drain Sump, 2.5"	400	ft	\$1.70		\$2.10			Means Ref.	\$400	\$0	\$0	\$1,566
SVE System	Drill & Install Casing for Vapor Extraction Wells, 4" diam & 30' depth	16	ea						Vendor Quote	\$0	\$0	\$0	\$375,000
	Vapor Extraction System	2	ea			\$5,500.00			Vendor Quote	\$0	\$0	\$0	\$10,500
	4" PVC Piping including Fittings	200	ft	\$3.41		\$2.77			Means Ref.	\$1,000	\$0	\$0	\$1,864
	6" PVC	40	ft	\$5.36		\$3.81			Means Ref.	\$210	\$0	\$0	\$322
	4" Butterfly Valves, PVC	2	ea	\$300.00		\$100.00			Vendor Quote	\$400	\$0	\$0	\$1,200
	Manholes	16	ea			\$1,500.00			Vendor Quote	\$0	\$0	\$0	\$22,500
	Vacuum Gages	16	ea			\$100.00			Vendor Quote	\$0	\$0	\$0	\$2,560
	Flow Element/Level Indicator	2	ea			\$2,000.00			Vendor Quote	\$0	\$0	\$0	\$4,000
	Shed Housing SVE Pumps & Carbon Adsorption Equipment	200	sq ft						Vendor Quote	\$0	\$0	\$0	\$4,000
	Electric Heater	1	ea		\$350.00				Prof. Judgment	\$0	\$0	\$0	\$4,000
Installation of Equipment	Installation Mechanical & Electrical	60	hrs		\$400.00				Prof. Judgment	\$0	\$0	\$0	\$2,240
	Materials	1	lb						Prof. Judgment	\$400	\$0	\$0	\$400
Off-gas Treatment	Carbon Adsorption System	2	ea						Vendor Quote	\$0	\$10,436	\$0	\$10,436
	Initial Granular Activated Carbon (3,000 lbs)	1	lb	\$2,340.00					Vendor Quote	\$2,340	\$0	\$0	\$3,240
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam. & 20' depth	4	ea						Vendor Quote	\$0	\$0	\$0	\$100,000
Additional Field Personnel	Surveyor	40	hr			\$60.00			Prof. Judgment	\$0	\$0	\$0	\$2,000
	Health & Safety	16	hr			\$65.00			Prof. Judgment	\$0	\$1,040	\$0	\$1,040
	Geologist	80	hr			\$65.00			Prof. Judgment	\$0	\$0	\$0	\$5,200
SVE Well Closures (1)	Labor	120	hr			\$35.00			Prof. Judgment	\$0	\$0	\$0	\$3,271
	Bottomless Great Wells	1	CT	\$64.00					Vendor Quote	\$60	\$0	\$0	\$60
Decommission French Drains (1)	Backhoe	1	day		\$2,200.00				Means Ref.	\$0	\$1,713	\$0	\$1,996
Subtotal Direct Capital Costs													
Indirect Capital Costs													
Eng. Design & Inspection	10% of direct materials, equipment, & labor								Prof. Eng. 000	\$977	\$2,490	\$3,713	\$0
	Misc. Labor & Materials								Prof. Eng. 000	\$780	\$0	\$2,476	\$0
	Permits								Prof. Judgment	\$336	\$336	\$1,226	\$0
	Construction Management								Prof. Judgment	\$651	\$1,090	\$2,476	\$0
	Project Management								ED&I	\$749	\$1,909	\$2,547	\$0
	Overhead, Profit & Bond								Prof. Eng. 000	\$1,648	\$4,300	\$6,262	\$0
	Subcontractor Fee								Prof. Eng. 000	\$0	\$0	\$0	\$67,431
Subtotal Indirect Capital Costs													
Contingency	50% of direct and indirect capital costs								Prof. Eng. 000	\$3,493	\$9,503	\$13,129	\$0
Total Capital Costs													
										\$15,124	\$35,904	\$56,601	\$821,263

Activity	Resource Description	Qty	Unit	Direct Costs Per Unit			Material	Equipment	Labor	Sub-contract			Total Costs	
				Material	Equipment	Labor				Material	Equipment	Labor		
Annual O&M Direct Costs														
Operations	Replacement Granular Activated Carbon	9,000	lb	\$1.06		\$1.26				Vendor Quote	\$6,720	\$0	\$11,260	\$20,970
	Transportation of Spent GAC	4	loads			\$2,700.00				Vendor Quote	\$0	\$0	\$10,800	\$10,800
	Disposal of Granular Activated Carbon	12,000	lb			\$0.34				Vendor Quote	\$0	\$0	\$4,080	\$4,080
	Electrical Costs of Vacuum Pumps	36	wh	\$47.00						Prof. Judgment	\$1,692	\$0	\$1,692	\$1,692
	Electrical Costs of Heater	34	wh	\$23.00						Prof. Judgment	\$857	\$0	\$857	\$857
	Confirmatory Sampling	80	ea			\$150.00				Prof. Judgment	\$0	\$0	\$12,000	\$12,000
	Groundwater Treatment @ IM/BA Treat.													
	Facil.	1	yr			\$676,000.00				(2)	\$0	\$0	\$676,000	\$676,000
Maintenance	Labor	180	hr			\$35.00				Prof. Judgment	\$0	\$0	\$6,300	\$6,300
	Materials & Parts	12	mo	\$200.00						Prof. Judgment	\$2,400	\$0	\$0	\$2,400
Personnel	Operator	2,400	hr			\$65.00				Prof. Judgment	\$0	\$0	\$156,000	\$156,000
	EM&S	96	hr			\$65.00				Prof. Judgment	\$0	\$0	\$6,240	\$6,240
	Geologist	180	hr			\$65.00				Prof. Judgment	\$0	\$0	\$11,700	\$11,700
Subtotal O&M Direct Costs														
\$14,570														
Annual O&M Indirect Costs														
Misc. Labor & Materials	10% of direct labor & \$1.50 in material cost for each direct labor hour									Facil. Eng. 000	\$2,540	\$0	\$2,540	\$2,540
Project Management	10% of direct materials, equipment, & labor costs									EM&S Cost Est.	\$1,488	\$0	\$1,488	\$1,488
Overhead, Profit & Bond	55.9% of direct materials, equipment, & labor costs									Facil. Eng. 000	\$3,933	\$0	\$3,933	\$3,933
Subcontractor Fee	10% of subcontractor costs									Facil. Eng. 000	\$0	\$0	\$0	\$0
Subtotal O&M Indirect Costs														
\$4,916														
Contingency	50% of total direct and indirect O&M costs									Facil. Eng. 000	\$0.000	\$0	\$0.000	\$0.000
Subtotal O&M Costs														
\$19,486														
Total Annual O&M Cost														
\$19,486														
Total O&M Costs (expenditures occur in yrs 2-5 @ 5% discount rate) (2)														
\$19,486														

Annual Post Closure Direct Costs													
Settlement Sampling	Collect Groundwater Samples	12	ea							Prof. Judgment	\$1,500.00	\$0	\$1,500.00
Analytical Work	Sample Analysis for VOCs & Inorganics	14	ea							Vendor Quote	\$4,100.00	\$0	\$4,100.00
Subtotal Post Closure Direct Costs													
											\$0	\$0	\$5,600.00
Annual Post Closure Indirect Costs													
Project Management	10% of post closure direct materials, equipment, & labor costs									EM&S Cost Est.	\$0	\$0	\$0
Subcontractor Fee	10% of post closure subcontractor costs									Facil. Eng. 000	\$0	\$0	\$0
Subtotal Post Closure Indirect Costs													
											\$0	\$0	\$0
Contingency	50% of total post closure direct and indirect costs									Facil. Eng. 000	\$0	\$0	\$0
											\$0	\$0	\$0
Total Annual Post Closure Costs													
											\$0	\$0	\$0
Total Post Closure Costs (continues for 20 yrs after completion of remedial action @ 5% discount rate)													
											\$0	\$0	\$0
											\$0	\$0	\$0

Total Cost Of Alternative													
											\$122,517	\$36,994	\$777,369
											\$0	\$0	\$7,204,517
											\$0	\$0	\$7,204,517

- (1) Future capital cost that takes place upon completion of treatment. (yr 6 @ 5% discount rate total cost in 1994 dollars)
- (2) Costs represent annual operating costs as presented in the Phase I Preliminary Plan For Future Utilization of Existing W for Treatment Facilities At Rocky Flats Plant, Draft Report (June 15 1994)
- (3) Operational expenses for the UV/peroxide water treatment plant during year 1 have also been included in the total cost.

Table E-6 Alternative 5 Groundwater Pumping and Soil Vapor Extraction With Thermal Enhancement

Activity	Resource Description	Qty	Unit	Base Costs Per Unit				Base Costs				Total Costs
				Mat'l	Equip.	Labor	Subcontract	Mat'l	Equip.	Labor	Subcontract	
Direct Capital Costs												
Soil Gas Survey	Geologist	120	hr			\$55.00			\$0	\$0	\$6,600	\$6,600
	Field Technician	80	hr			\$25.00			\$0	\$0	\$2,000	\$2,000
	Portable GC	2	wk		\$1,000.00				\$0	\$2,000	\$0	\$2,000
	Probes, Pump, and Misc. Equipment	100	ea		\$15.00				\$0	\$1,500	\$0	\$1,500
Dewatering	Drill Extraction Wells, 6" diam, 20 ft depth	2	ea						\$0	\$0	\$60,000	\$60,000
	10 gpm submersible pumps	2	ea		\$300.00				\$0	\$0	\$0	\$0
	PVC Piping to French Drain Sump, 2.5"	400	lf	\$1.70		\$2.19			\$400	\$276	\$0	\$1,556
SVE System	Drill & Install Casing for Vapor Extraction Wells, 4" diam & 20' depth	15	ea						\$0	\$0	\$375,000	\$375,000
	Vapor Extraction System	2	ea						\$0	\$0	\$10,500	\$10,500
	6" PVC Piping Including Fittings	300	lf	\$3.41		\$2.77			\$1,023	\$831	\$0	\$1,854
	6" PVC	40	lf	\$5.35		\$2.81			\$210	\$112	\$0	\$322
	4" Butterfly Valves, PVC	2	ea	\$300.00		\$100.00			\$600	\$300	\$0	\$1,200
	Manholes	15	ea						\$0	\$0	\$22,500	\$22,500
	Vacuum Gages	15	ea						\$0	\$0	\$2,250	\$2,250
	Flow Element/Local Indicator	2	ea						\$0	\$0	\$4,000	\$4,000
	Shed Housing SVE Pumps & Carbon Adsorption Equipment	200	sq ft						\$0	\$0	\$4,000	\$4,000
	Electric Heater	1	ea		\$350.00				\$0	\$0	\$0	\$0
									\$0	\$350	\$0	\$350
	RF Heating Unit	Setup, Startup, & Tuning	1	ln						\$0	\$0	\$80,000
	Equipment Rental	60	wk						\$0	\$0	\$500,000	\$500,000
Installation of Equipment	Installation Mechanical & Electrical	64	hrs			\$25.00			\$0	\$2,240	\$0	\$2,240
	Materials	1	ln	\$400.00					\$400	\$0	\$0	\$400
Off-gas Treatment	Carbon Adsorption System	2	ea		\$5,210.00				\$0	\$10,420	\$0	\$10,420
	Initial Granular Activated Carbon (5,000 lbs)	1	ln	\$3,240.00					\$3,240	\$0	\$0	\$3,240
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam. & 20' depth	4	ea						\$0	\$0	\$100,000	\$100,000
Additional Field Personnel	Surveyor	40	hr			\$50.00			\$0	\$2,000	\$0	\$2,000
	Health & Safety	16	hr			\$65.00			\$0	\$1,040	\$0	\$1,040
	Geologist	80	hr			\$65.00			\$0	\$5,200	\$0	\$4,000
Well Closure (1)	Labor	120	hr			\$25.00			\$0	\$3,015	\$0	\$3,015
	Backhoe Great Wells	1.40	CT	\$54.00					\$65	\$0	\$0	\$65
Decommission French Drain (1)									\$0	\$0	\$0	\$0
	Backhoe	1	day		\$2,200.00	\$302.00			\$0	\$1,894	\$312	\$2,205
Subtotal Direct Capital Costs												
								\$6,616	\$16,760	\$28,196	\$1,214,310	\$1,265,734
Indirect Capital Costs												
Eng. Design & Inspection	15% of direct materials, equipment, & labor								\$978	\$2,517	\$3,789	\$7,264
Misc Labor & Materials	10% of direct labor & \$1.50 in materials cost for each direct labor hour								\$780	\$0	\$2,515	\$3,293
Permits	5% of direct materials, equipment, & labor								\$328	\$839	\$1,266	\$2,421
Construction Management	10% of direct materials, equipment, & labor								\$403	\$1,678	\$2,133	\$4,142
Project Management	10% of direct materials, equipment, & labor								\$403	\$1,678	\$2,133	\$4,142
Overhead, Profit & Bond	25.3% of direct materials, equipment, & labor								\$1,649	\$4,245	\$5,357	\$12,251
Subcontractor Fee	10% of subcontractor costs								\$0	\$0	\$121,431	\$121,431

Activity	Remedial Investigation	QTY	Unit	Base Costs Per Unit				Base Costs			Total Costs
				Mat'l	Equip.	Labor	Sub-material	Mat'l	Equip.	Labor	
Subtotal Indirect Capital Costs								\$5,194	\$11,500	\$10,397	\$187,071
Contingency	50% of direct and indirect capital costs							\$5,400	\$8,300	\$13,337	\$435,941
Total Capital Costs								\$10,594	\$19,800	\$23,734	\$1,045,745

Annual O&M Direct Costs											
Operations											
Replacement Granular Activated Carbon	9,000	lb		\$1.08		\$1.36		\$9,720	\$0	\$11,360	\$39,070
Transportation of Spent GAC	4	loads					\$2,700.00	\$0	\$0	\$0	\$10,000
Disposal of Granular Activated Carbon	13,000	lb					\$0.14	\$0	\$0	\$0	\$4,000
Electrical Costs of Vacuum Pumps	20	wk		\$47.00				\$1,000	\$0	\$0	\$1,000
Electrical Costs of Blower	24	hrs		\$22.00				\$528	\$0	\$0	\$528
Electrical Costs of IF Unit (400 KW Unit)	1	yr					\$117,000.00	\$0	\$0	\$0	\$117,000
Groundwater Sampling	80	ea					\$150.00	\$0	\$0	\$0	\$12,000
Groundwater Treatment @ IM/BA Treat.							\$675,000.00	\$0	\$0	\$0	\$675,000
Facil.	1	yr						\$0	\$0	\$0	\$675,000
Maintenance								\$0	\$0	\$0	\$6,720
Labor	150	hr		\$35.00				\$5,250	\$0	\$0	\$5,250
Materials & Parts	15	ea		\$200.00				\$3,000	\$0	\$0	\$3,000
Personnel								\$0	\$0	\$0	\$145,000
Operator	4,100	hr		\$35.00				\$143,500	\$0	\$0	\$143,500
HAAS	90	hr		\$65.00				\$5,850	\$0	\$0	\$5,850
Geologist	150	hr		\$65.00				\$9,750	\$0	\$0	\$9,750
Subtotal O&M Direct Costs								\$14,270	\$0	\$190,370	\$1,015,230

Annual O&M Indirect Costs											
Misc Labor & Materials								\$4,000	\$0	\$18,037	\$24,037
Project Management								\$1,438	\$0	\$18,037	\$19,475
Overhead, Profit & Bond								\$3,000	\$0	\$45,004	\$48,004
Subcontractor Fee								\$0	\$0	\$0	\$0
Subtotal O&M Indirect Costs								\$12,000	\$0	\$81,081	\$175,791
Contingency	50% of total direct and indirect O&M costs							\$7,500	\$0	\$78,833	\$237,304
Total Annual O&M Cost								\$24,330	\$0	\$159,914	\$1,445,337
Total O&M Costs (expenditures occur in Years 2-5 @ 5% discount rate)								\$24,330	\$0	\$159,914	\$1,445,337

Annual Post Closure Direct Costs											
Subsistent Sampling											
Analytical Work											
Subtotal Post Closure Direct Costs											
Annual Post Closure Indirect Costs											
Project Management											
Subcontractor Fee											
Subtotal Post Closure Indirect Costs											
Contingency	50% of total post closure direct and indirect costs										
Total Annual Post Closure Costs											
Total Post Closure Costs (continues for 50 yrs after completion of remedial action @ 5% discount rate)											

Activity	Maximum Discharge/Week	CY	Unit	Base Costs Per Unit				Base Costs				Sub-contract	Total Costs
				Mat'l	Equip.	Labor	Sub-contract	Mat'l	Equip.	Labor			
Total Cost Of Alternative													

- (1) Future Capital Cost that takes place upon completion of treatment. (Y 3 @ 5% discount rate total cost is in 1994 dollars)
- (2) Costs represent annual operating costs as presented in the Phase I Preliminary Plan For Future Utilization of Existing Water Treatment Facilities At Rocky Flats Plant, Draft Report (June 16 1994)
- (3) Operational expenses for the UV/peroxide water treatment plant during year 1 have been included in the total cost.

Table E-7 Alternative 6 Hot Air Injection with Mechanical Mixing

Activity	Resource Description	QTY	Unit	Base Costs Per Unit				Base Costs				Total Costs
				Mat'l	Equip.	Labor	Sub- contract	Mat'l	Equip.	Labor	Sub- contract	
Direct Capital Costs												
Soil Gas Survey	Geologist	120	hr			\$65.00		\$0	\$0	\$6,000	\$0	\$6,000
	Field Technician	80	hr			\$65.00		\$0	\$0	\$2,800	\$0	\$2,800
	Portable GC	2	wk	\$1,000.00				\$0	\$2,000	\$0	\$0	\$2,000
	Probes, Pump, and Misc. Equipment	100	ea	\$18.00				\$0	\$1,500	\$0	\$0	\$1,500
Dewatering	Drill Extraction Wells, 4 diam, 20 ft depth	5	ea				\$25,000.00					
	5 gal submersible pumps	5	ea	\$500.00				\$0	\$0	\$0	\$125,000	\$125,000
	PVC Piping to French Drain Sump, 2.5"	1000	ft	\$1.70		\$2.15		\$1,700	\$0	\$2,150	\$0	\$3,850
Treatment of Soils	Mechanical Mixing Tool (1)	4,000	cy				\$150.00		\$0	\$0	\$675,000	\$675,000
	Surveyor	40	hr			\$60.00		\$0	\$0	\$2,400	\$0	\$2,400
Additional Field Personnel	Health & Safety	16	hr			\$65.00		\$0	\$0	\$1,040	\$0	\$1,040
	Geologist	80	hr			\$65.00		\$0	\$0	\$4,400	\$0	\$4,400
Drill Monitoring Wells	Drill & Case 4 wells, 6" diam. & 20' depth	4	ea				\$25,000.00		\$0	\$0	\$100,000	\$100,000
	Labor	50	hr			\$35.00		\$0	\$0	\$1,554	\$0	\$1,554
Well Closure (2)	Bentonite Grout Walls	0.40	CY	\$54.00				\$20	\$0	\$0	\$0	\$20
	Backhoe	1	day	\$2,500.00		\$300.00		\$0	\$1,001	\$328	\$0	\$2,319
Subtotal Direct Capital Costs								\$1,720	\$8,901	\$20,941	\$600,000	\$630,562
Indirect Capital Costs												
Eng, Design & Inspection	15% of direct materials, equipment, & labor							\$258	\$1,049	\$3,141	\$0	\$4,448
	10% of direct labor & \$1.50 in materials cost for each direct labor hour							\$279	\$0	\$2,084	\$0	\$2,373
	5% of direct materials, equipment, & labor							\$95	\$350	\$1,047	\$0	\$1,493
	10% of direct materials, equipment, & labor							\$172	\$689	\$2,084	\$0	\$2,935
	10% of direct materials, equipment, & labor							\$198	\$792	\$2,408	\$0	\$3,410
	25.3% of direct materials, equipment, & labor							\$425	\$1,709	\$5,295	\$0	\$7,429
	10% of subcontractor costs							\$0	\$0	\$0	\$0	\$0
Subtotal Indirect Capital Costs								\$1,428	\$4,570	\$18,088	\$60,000	\$112,181
Contingency												
	50% of direct and indirect capital costs							\$644	\$2,408	\$11,107	\$287,000	\$312,550
Total Capital Costs								\$4,001	\$16,159	\$48,132	\$1,267,000	\$1,364,352

Annual O&M Direct Costs												
Groundwater Treatment												
	Facil	1	yr					\$0	\$0	\$0	\$675,000	\$675,000
Subtotal O&M Direct Costs												
								\$0	\$0	\$0	\$675,000	\$675,000
Annual O&M Indirect Costs												
Misc. Labor & Materials												
	10% of direct labor & \$1.50 in materials cost for each direct labor hour							\$0	\$0	\$0	\$0	\$0
Project Management												
	10% of direct materials, equipment, & labor costs							\$0	\$0	\$0	\$0	\$0
Overhead, Profit & Bond												
	25.3% of direct materials, equipment, & labor costs							\$0	\$0	\$0	\$0	\$0
Subcontractor Fee												
	10% of subcontractor costs							\$0	\$0	\$0	\$67,500	\$67,500

Table E-8 Alternative 7 Soil Excavation and Groundwater Removal With Sump Pumps

Activity	Resource Description	Qty	Unit	Base Costs Per Unit				Source	Base Costs				Total Costs
				Mat'l	Equip.	Labor	Sub-contract		Mat'l	Equip.	Labor	Sub-contract	
Direct Capital Costs													
Construct Staging Area	Tiered Shovelfoot	404	CY		\$0.70	\$0.30		Means Ref.	\$0	\$283	\$61	\$0	\$364
	Base Course	1008	SY	\$2.12	\$0.20	\$0.22		Means Ref.	\$2,136	\$201	\$221	\$0	\$2,558
	Reinforced Concrete Slab, 6" Thick	104	CY	\$46.28	\$0.00	\$72.83		Means Ref.	\$9,101	\$0	\$7,542	\$0	\$17,669
	Submersible Pump, 5 gpm	2	ea		\$300.00			Vendor Quote	\$0	\$600	\$0	\$0	\$600
	2" PVC Piping (including fittings)	200	W	\$1.44		\$2.04		Means Ref.	\$288	\$0	\$408	\$0	\$696
Scrape Top Soil & Stockpile	Towed Scraper	322	CY		\$3.46	\$0.81		Means Ref.	\$0	\$1,114	\$282	\$0	\$1,376
Dust Control	Water Truck, 5,000 gal capacity	1,440	hr		\$21.22	\$28.00		Means Ref.	\$0	\$30,701	\$38,000	\$0	\$68,701
Excavate Soil, Haul to Staging Area	Dumper	22,630	CY		\$2.91	\$0.64		Means Ref.	\$0	\$65,063	\$14,483	\$0	\$80,337
	Backhoe	1,122	CY		\$0.08	\$0.27		Means Ref.	\$0	\$1,109	\$419	\$0	\$1,528
	Front End Loader	21,406	CY		\$0.06	\$0.26		Means Ref.	\$0	\$14,189	\$7,526	\$0	\$21,714
	Dump Trailer	22,630	CY		\$1.18	\$0.48		Means Ref.	\$0	\$26,026	\$10,410	\$0	\$36,434
	20 gpm Suction Pumps	2	ea		\$400.00			Vendor Quote	\$0	\$800	\$0	\$0	\$800
Decontaminating	2.5" PVC Piping (including fittings)	200	W	\$1.70		\$2.19		Means Ref.	\$340	\$0	\$438	\$0	\$778
	Corrugated Metal Piping	12	W	\$4.18		\$1.40		Means Ref.	\$50	\$0	\$17	\$0	\$67
	Pan Gravel	20	CY	\$17.86				Means Ref.	\$357	\$0	\$0	\$0	\$357
	Health & Safety Specialist	1,400	hr			\$64.00		EC&G Rad Eng.	\$0	\$0	\$90,640	\$0	\$90,640
	Monitoring Equipment Maintenance	100	hr			\$68.00		EC&G Rad Eng.	\$0	\$0	\$6,800	\$0	\$6,800
Rad Screening of Soils	Thermal Desorption Unit	22,630	CY			\$75.00		Vendor Quote	\$0	\$0	\$1,697,250	\$0	\$1,697,250
	Thermal Desorption Unit Mobilization	1	h			\$4,000.00		Vendor Quote	\$0	\$0	\$4,000	\$0	\$4,000
	Thermal Desorption Unit Demobilization	1	h			\$1,500.00		Vendor Quote	\$0	\$0	\$1,500	\$0	\$1,500
	Wheel Mounted Front End Loader	45,260	CY		\$0.66	\$0.35		Means Ref.	\$0	\$29,872	\$15,841	\$0	\$45,713
	Transportation to Disposal Facility (90 mi)	22,630	CY			\$33.00		Vendor Quote	\$0	\$0	\$1,199,390	\$0	\$1,199,390
Transportation/Disposal of Soil	Disposal at Licensed Facility	22,630	CY			\$123.00		Vendor Quote	\$0	\$0	\$2,783,490	\$0	\$2,783,490
	Soil Samples	1,866	ea			\$250.00		Prof. Judgment	\$0	\$0	\$471,458	\$0	\$471,458
	UV/Precipitate & IX Treatment System	1	yr			\$676,000.00		(1)	\$0	\$0	\$676,000	\$0	\$676,000
	Pit-Ram PMA/Gravel, 5 m haul	22,630	CY	\$3.57	\$4.86	\$1.83		Means Ref.	\$80,789	\$109,982	\$41,413	\$0	\$232,184
	Tiered Shovelfoot, 12" lifts	22,630	CY		\$0.35	\$0.09		Means Ref.	\$0	\$7,921	\$2,037	\$0	\$9,957
Decontaminate French Drain	Revegetation	2,904	SY	\$0.22	\$0.06	\$0.06		Means Ref.	\$639	\$174	\$174	\$0	\$987
	Backhoe	1	day		\$2,200	\$382.00		Means Ref.	\$0	\$2,200	\$382	\$0	\$2,582
	Drill & Case 4 wells, 6" diam. & 20' depth	4	ea			\$35,000.00		Vendor Quote	\$0	\$0	\$100,000	\$0	\$100,000
	Sr. Geologist	110	hrs			\$75.00		Prof. Judgment	\$0	\$0	\$8,250	\$0	\$8,250
	Surveyor	100	hrs			\$50.00		Prof. Judgment	\$0	\$0	\$5,000	\$0	\$5,000
Confidentiality Sampling	Soil Samples From Excavation Site	25	ea			\$250.00		Prof. Judgment	\$0	\$0	\$6,250	\$0	\$6,250
Subtotal Direct Capital Costs													
Indirect Capital Costs													
Engineering, Design & Inspection	10% of direct materials, equipment, & labor							Facil. Eng. 000	\$14,091	\$43,808	\$38,213	\$0	\$96,113
Miscellaneous Labor & Materials	10% of direct labor & \$1.60 in materials cost for each direct labor hour							Facil. Eng. 000	\$8,403	\$0	\$24,142	\$0	\$32,545

Activity	Resources Description	QTY	Unit	Direct Costs Per Unit				Indirect Costs				Total Costs
				Mat'l	Equip.	Labor	Sub-contractor	Mat'l	Equip.	Labor	Sub-contractor	
Permit	5% of direct materials, equipment, & labor											
Construction Management	10% of direct materials, equipment, & labor											
Project Management	10% of direct materials, equipment, & labor											
Overhead, Profit & Bond	20 5% of direct materials, equipment, & labor											
Subcontractor Fee	10% of subcontractor costs											
Subtotal Indirect Capital Costs												
Contingency	30% of direct and indirect capital costs											
Total Capital Costs												

Annual O&M Direct Costs												
Subtotal O&M Direct Costs												
Total O&M Costs												

Annual Post Closure Direct Costs												
Soil Sampling	Collect Groundwater Samples	12	ea									
Groundwater Monitoring	Semiannual Groundwater Sampling	14	ea									
Regulation	10% of Excavated Area/Yr	200	SY	30.22	30.06	30.06						
Subtotal Post Closure Direct Costs												
Annual Post Closure Indirect Costs												
Project Management	10% of post closure direct materials, equipment, & labor costs											
Subcontractor Fee	10% of post closure subcontractor costs											
Subtotal Post Closure Indirect Costs												
Contingency	30% of total post closure direct and indirect costs											
Total Annual Post Closure Costs												
Total Post Closure Costs (20 Yr @ 5% discount rate)												

Total Cost of Alternative												
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(1) Cost represents annual operating cost as presented in the Phase I Preliminary Plan for Future Utilization of Existing Water Treatment Facilities at Rocky Flats Plant, Draft Report (June 15, 1994). It has been assumed for cost estimating purposes that the French Drain and OU 1 Water Treatment Plant will operate for one year during excavation and treatment.

OU-1 DOMESTIC WATER SUPPLY SIMULATIONS

**The results of computer simulations of domestic
water production capabilities from subsurface units beneath
OU 1 at the Rocky Flats Plant Golden Colorado**

**This work was performed by the Geosciences Division
in support of risk analysis studies**

December 14 1992

INTRODUCTION

To investigate the water production capabilities of the colluvial materials beneath Operable unit 1 at the Rocky Flats Plant several transient pumping computer simulations were performed. These simulations were designed to determine whether these saturated materials could produce sufficient water to supply a hypothetical four member household. A daily pumping requirement of 240 gallons per day (gpd) was assumed based on a daily water requirement of 60 gallons per person.

METHOD

Simulations were performed using the USGS MODFLOW groundwater flow simulation package (McDonald and Harbaugh 1988). Input parameters common to all simulations are listed in Table 1. Simulations were run using a daily time frame until the pumping well grid cell went dry or the end of the simulation (365 days) was reached.

The pumping well was located at the center of the 19 by 19 grid cell array. A variable grid spacing ranging from 5 feet at the well to 50 feet at the boundaries was used to provide realistic drawdown conditions near the well. The grid spacing for each scenario are given in Table 1 and shown in Figure 1. The specific yield came from lab analyses of core samples and example values from the literature for fine grained materials (Fetter 1980 pg 68). Boundary conditions were constant head equal to the initial head.

Table 1

PARAMETER	VALUE	SOURCE
Hydraulic Conductivity	1E 4 to 1E 5 cm/sec	Table 3.6 of OU1 Phase III Report
Specific Yield	0.10	Lab analyses/literature
Grid Spacing (variable)	from 5 to 50 ft	Assumed
Hydrogeologic Unit Character	Unconfined	On site observation
Initial Saturated Thickness	10 ft	Figure 3.36 of OU1 Phase III Report
Boundary Conditions	Constant head	Assumed

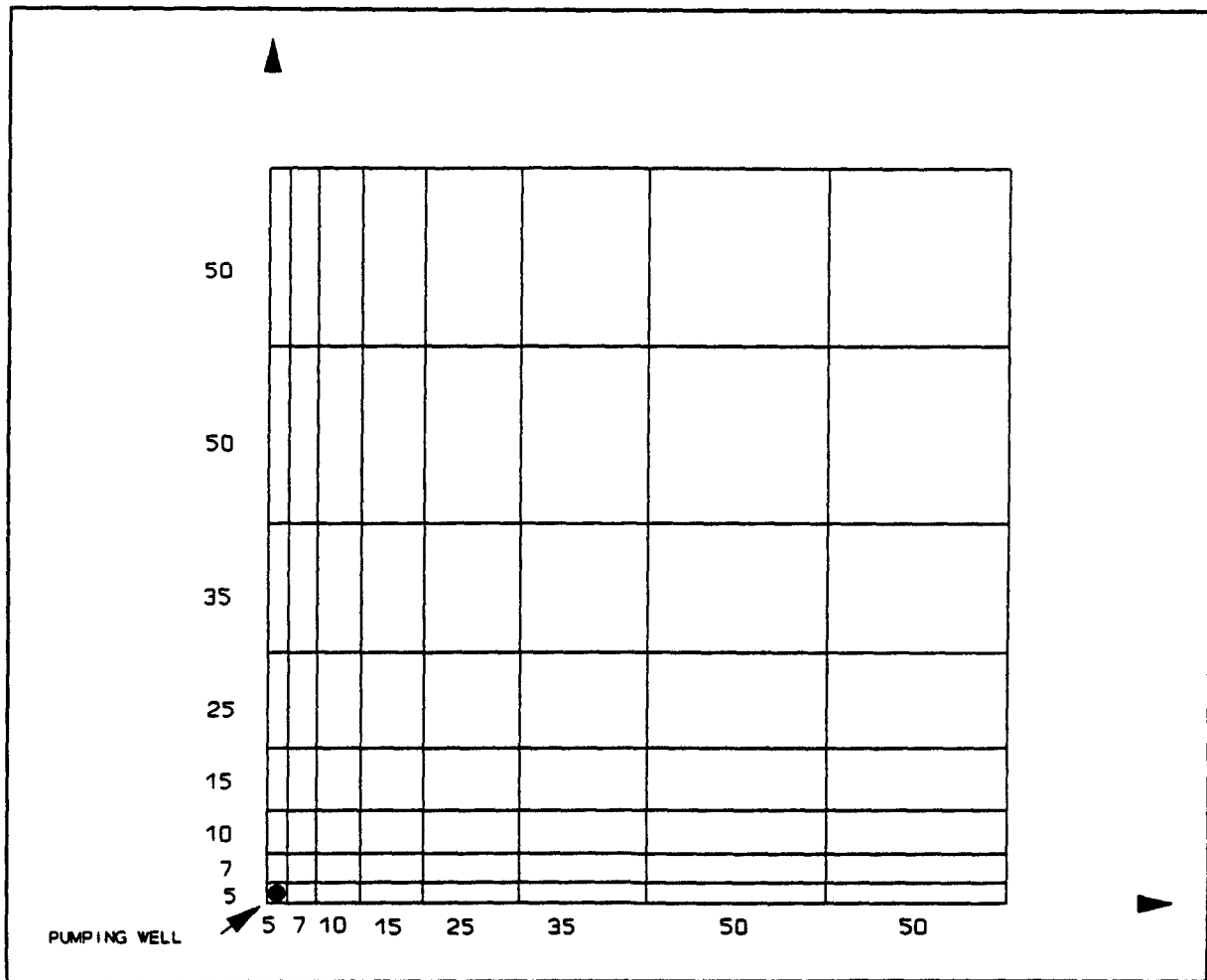


Figure 1 Figure shows 1/4 (upper right hand quadrant) of an example model grid. In model well is at center of grid. Grid spacings in feet. The number of grid nodes for each model may differ but grid spacings are similar. Not to scale.

1.5 GPM SCENARIO

For this scenario a pumping rate of 1.5 gpm was used. This rate is below the 3.5 gpm rate commonly used for domestic wells and as such is conservative. Each day of the transient simulation was divided into two stress periods and each period was divided into two timesteps. The first 2.7 hours of each day was used as a pumping period. It was assumed that the household maintained water storage capabilities and that this pumping period was used to replenish the water storage system. The pumping period was based on the total daily water requirement (240 gal) and the pumping rate (1.5 gpm).

$$240 \text{ gal} / (1.5 \text{ gal/min} \cdot 60 \text{ min/hr}) = 2.7 \text{ hrs}$$

The remaining 21.3 hours of each day allowed water level recovery to take place.

To determine the effect of uncertainty in hydraulic conductivity, two simulations with different conductivity parameters were run. The results from these simulations are shown in the following table.

Summary of simulation results for 1.5 gpm scenario

HYDRAULIC CONDUCTIVITY (CM/SEC)	WATER PRODUCTION DAYS
1E 5	< 1
1E 4	< 1

Results

For the 1.5 gpm scenario the pumping well grid cell went dry within the first day of the simulation regardless of which hydraulic conductivity was used. This is consistent with the low hydraulic conductivity and small saturated thickness observed for 881 Hillside colluvial materials.

MAXIMUM POTENTIAL WATER PRODUCTION

To further investigate the potential for water production from the colluvial materials on the 881 Hillside several simulations with differing pumping rates were performed. These simulations were not designed to produce 240 gallons of water per day but instead were intended to determine a potential maximum water production. For this reason each day of the transient simulation was divided into two stress periods with each period divided into two timesteps. The first 12 hour stress period was a pumping period and second 12 hour segment was a recovery phase. Again two different hydraulic conductivities were examined. All other simulation parameters are as listed in Table 1 and shown in Figure 1.

Results from simulations with a hydraulic conductivity of $1\text{e-}4$ cm/sec are shown in the following table. Each row represents a different pumping rate (given both in cubic feet per day and gallons per minute). The "Daily Water Production" column gives the equivalent daily water production rate in gallons. This is the rate at which water was being produced prior to any desaturation of the well cell within the model and assumes a 12 hour pumping period. The "Water Production Days" column gives the number of simulated days before the well cell was desaturated (dried up). Values for "Water Production Days" greater than 365 indicate the well cell did not desaturate during the simulation.

Simulation Results with $K = 1\text{e-}4$ cm/sec

PUMPING RATE FT ³ /DAY	PUMPING RATE GPM	DAILY WATER PRODUCTION (GAL)	WATER PRODUCTION DAYS
100	0.52	374	< 1
50	0.26	187	3.5
35	0.18	130	43.5
30	0.16	115	221.5
27	0.14	101	> 365

Results from simulations with a hydraulic conductivity of 1×10^{-5} cm/sec are shown in the following table. Column and row descriptions are as listed for the previous table. Note that pumping rates are lower than those in the previous simulation.

Simulation Results with $K = 1 \times 10^{-5}$ cm/sec

PUMPING RATE FT ³ /DAY	PUMPING RATE GPM	DAILY WATER PRODUCTION (GAL)	WATER PRODUCTION DAYS
27	0.14	101	2.25
10	0.052	37	9.25
5	0.026	19	70.5
2.5	0.013	9	> 365

An additional simulation was run using a hydraulic conductivity based on OU 1 field measurements. The geometric mean of single well tests in colluvial materials was 1.75×10^{-5} cm/sec. Using this K and the same values presented for other parameters gives a maximum pumping rate of 6.0 ft³/day (or 22.4 gallons per day) for a 12 hour pump period without desaturating the well.

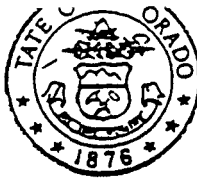
Results

The results from these simulations to investigate the maximum potential water production capabilities from the 881 Hillside colluvium indicate maximum expected production capabilities that are less than 10% of that required to supply a family of four (240 gallons). In reality long term production rates would be lower because of the constant head boundary conditions assumed in the model. This type of boundary condition would represent an infinite water source to the well given a sufficiently low pumping rate. Actual field conditions on the 881 Hillside consist of saturated regions often surrounded by desaturated zones which would limit long term water production capabilities. The simulation also assumed a constant saturated thickness across the model domain. Field data from the 881 Hillside indicate that the thickness of saturated colluvium varies, often thinning below the 10 foot saturated thickness assumed in the modeling. The combination of these factors suggest that the model determined pumping rates would be higher than would be expected from an actual water production well on the 881 Hillside.

References

Fetter C W Jr 1980 Applied Hydrogeology Merrill Publishing Company
Columbus 488 p

McDonald Michael G and Harbaugh Arlen W 1988 Techniques of Water
Resources Investigations of the United States Geological Survey Book 6
Chapter A1 A Modular Three dimensional Finite difference Groundwater Flow
Model



OFFICE OF THE STATE ENGINEER
DIVISION OF WATER RESOURCES

1313 Sherman Street Room 818
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(303) 866 3581
FAX (303) 866 3589
March 12 1992

Mr. Scott Grace
United States Department of Energy
Rocky Flats Office
P O Box 928
Golden CO 80402-0928

Dear Mr. Grace

We have reviewed the document submitted entitled Public Health Risk Assessment B81 Hillside Area (DUI) Technical Memorandum No. 6 Exposure Scenarios Revision 3.0 dated March 1992. The purpose of our review was to specifically evaluate the findings presented in Appendix B Investigation and Simulation of Water Production Capabilities.

The basic conclusion of this appendix is that neither the shallow alluvial aquifer (Rocky Flats Alluvium) nor the underlying Arapahoe Aquifer is capable of producing sufficient water for even domestic purposes. This conclusion was derived from model simulation runs utilizing the USGS MODFLOW ground water flow simulation package. This conclusion is applied only to the B81 Hillside area.

While the basic input parameters are given in the appendix, actual model setup and output were not submitted. Basically, the parameters selected and presented in Table B-3 and Table B-4 appear to be reasonable with the exception of the specific yield value for the Arapahoe Aquifer. Based on previous work by the USGS and on research funded by this office and the Colorado Water Conservation Board, the actual specific yield of the Arapahoe Aquifer ranges between 0.15 and 0.20. The simulation runs used a value of 0.30. The use of the higher value will result in more water being released from storage and a more rapid depletion. This will cause wells to dry up more quickly than they may in actuality. Although we suggest that the model be rerun with a specific yield of no more than 0.20, we do not feel that the result will significantly change the conclusion. It will change the length of time necessary to deplete wells.

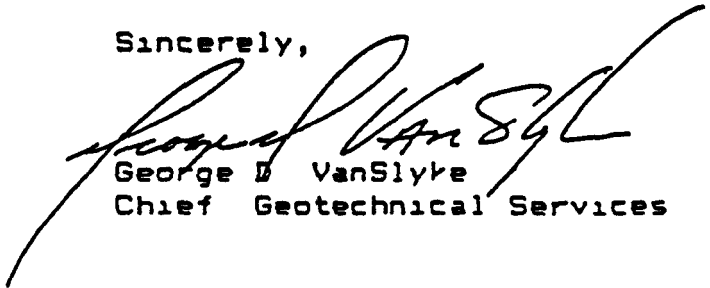
Based on these comments, we feel that the conclusion that neither aquifer is a potential source for domestic water supplies in the B81 Hillside area is valid when considering future land use.

We would like to comment on several statements made in the document which are not necessarily correct and should be corrected prior to issuance of the final document.

- 1 Page B-5 Paragraph 4 -- This paragraph states that domestic wells drilled to the Laramie-Fox Hills Aquifer (500 to 700 feet) are not an economically viable alternative This is not true It is quite common in the Denver Basin for domestic wells to be drilled to depths in excess of 1000 feet Therefore Laramie-Fox Hills wells for domestic purposes are very likely in the future depending on the permitted land use
- 2 Page B-8, last paragraph -- It is stated that well yields listed in Table B-5 are the maximum permissible pumping rates Actually the rates listed for the domestic wells are those reported by the driller at the time the well was completed and actual permissible pumping rates may be either 15 gpm or 25 gpm depending on the year the well was permitted It is true that the permissible rate is independent of the actual sustained yield Permitted pumping rates for wells other than domestic and stock (permit numbers with the suffix F) may also be different than either the maximum pumping rate or the sustained yield
- 3 Page B-13 first paragraph -- Permitted well yields of less than 15 gpm do not necessarily mean that a well is limited to domestic or stock use
- 4 Page B-17 last paragraph -- It is stated that the bedrock dips approximately 1 degree However Page 2 states that the dip is 2 degrees

We hope that these comments are helpful Should you have any questions please contact me at (303) 866-3585

Sincerely,



George D. VanSlyke
Chief Geotechnical Services

cc Hal Simpson, Acting State Engineer
Gary Baughman Colorado Department of Health Rocky Flats Unit
Ron Cattani Executive Director's Office CDNR

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AQUIFER PUMPING TEST DATA SHEET

DATE 5/21/92

PERSON RECORDING DATA R Smith - EG&G

WELL # 0487 (881 Hills. 2E)

HYDROSTRATIGRAPHIC UNIT Colluvium

SCREENED INTERVAL 9.94 ft to 21.49 ft from TOC-SS (SATURATED SCREENED INTERVAL)

STATIC WATER LEVEL 9.94 ft @ TOC-SS PUMPING WELL ID 2 in (CASING)

DISTANCE TO PUMPING WELL N/A ft WELL ID 21.94 ft @ TOC-SS

PUMP INTAKE DEPTH 21.17 ft

TEST START TIME 13 19 55

INITIAL SATURATED THICKNESS 12.00 ft

ELAPSED TIME (Units)/(min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units)/(hr)	WATER LEVEL
00	ON				9.94
122	OFF	305	122	25	
322	ON				
408	OFF	0	(pump NOT PRIMING)		
858	ON				
915	OFF	0.75	793	0.09	
1108	ON				
12.58	OFF	0	(pump NOT PRIMING)		
16.08	ON				
17.53	OFF	0.5	8.38	0.060	
24.58	ON				
26.42	OFF	0.5	8.89	0.056	
34.58	ON				

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PERSON RECORDING DATA _____

WELL # 0487

HYDROSTRATIGRAPHIC UNIT _____

(see Page 1)

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL ID _____ in

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME _____

ELAPSED TIME (Units)(min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units)(gpm)	WATER LEVEL (ft)
36 05	OFF	0 5	9 63	0 052	
45 58	ON				
46 77	OFF	0 5	10 72	0 047	
56 92	ON				
57 30	OFF	0 45	10 52	0 043	
68 58	ON				
70 16	OFF	0 5	12 86	0 039	
82 75	ON				
84 50	OFF	0 6	14 34	0 042	
97 58	ON				
99 42 (0)	OFF	0 65	14 92	0 044	
101 25 (1 83)		(RECOVERY DATA)			21 28
102 33 (2 91)					20 98

END OF
TEST

(4011)(AQT-5)(10-9/91)

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PERSON RECORDING DATA _____

WELL # 0487

HYDROSTRATIGRAPHIC UNIT

(see page 1)

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL I.D. _____ in

DISTANCE TO PUMPING WELL ft.

TEST START TIME / /

[illegible]

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DATE 5/21/92

PERSON RECORDING DATA R Smith - EG&G

WELL # 37191 (881 Hillside)

HYDROSTRATIGRAPHIC UNIT Colluvium

SCREENED INTERVAL 13 9 ft to 23 8 ft from TOC-PVC

STATIC WATER LEVEL 7.55 ft PUMPING WELL ID 2 in (casing)
@ TOC-PVC

DISTANCE TO PUMPING WELL N/A ft WELL T.D. 2580 ft @ TOC-PVC

PUMP INTAKE DEPTH 2503 ft

TEST START TIME 09 38 25

INITIAL SATURATED THICKNESS 51725

ELAPSED TIME (Units)/(min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units)/(min)	WATER LEVEL
<u>00</u>	<u>ON</u>				<u>755</u>
<u>25</u>	<u>OFF</u>	<u>30</u>	<u>25</u>	<u>120</u>	
<u>45</u>	<u>ON</u>				
<u>55</u>	<u>OFF</u>	<u>12</u>	<u>30</u>	<u>040</u>	
<u>75</u>	<u>ON</u>				
<u>825</u>	<u>OFF</u>	<u>10</u>	<u>275</u>	<u>036</u>	
<u>1025</u>	<u>ON</u>				
<u>1141</u>	<u>OFF</u>	<u>05</u>	<u>316</u>	<u>010</u>	
<u>1441</u>	<u>ON</u>				
<u>1524</u>	<u>OFF</u>	<u>04</u>	<u>323</u>	<u>010</u>	
<u>1924</u>	<u>ON</u>				
<u>1974</u>	<u>OFF</u>	<u>04</u>	<u>450</u>	<u>009</u>	
<u>2474</u>	<u>ON</u>				

(4011)(AOT-5T)(1079/01)

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PERSON RECORDING DATA _____

WELL # 37191

(SEE Page 1)

HYDROSTRATIGRAPHIC UNIT _____

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL I.D. _____ in

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME _____

ELAPSED TIME (Units) (min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	ΔT (min)	AVERAGE Q (pumping well) (Units)/(min)	WATER LEVEL (ft)
<u>25.07</u>	<u>OFF</u>	<u>0.3</u>	<u>5.33</u>	<u>0.056</u>	_____
<u>33.07</u>	<u>ON</u>	_____	_____	_____	_____
<u>33.66</u>	<u>OFF</u>	<u>0.5</u>	<u>8.59</u>	<u>0.058</u>	_____
<u>39.66</u>	<u>ON</u>	_____	_____	_____	_____
<u>40.22</u>	<u>OFF</u>	<u>0.5</u>	<u>6.66</u>	<u>0.075</u>	_____
<u>47.33</u>	<u>ON</u>	_____	_____	_____	_____
<u>48.25</u>	<u>OFF</u>	<u>0.4</u>	<u>7.93</u>	<u>0.050</u>	_____
<u>57.25</u>	<u>ON</u>	_____	_____	_____	_____
<u>58.11</u>	<u>OFF</u>	<u>0.55</u>	<u>9.91</u>	<u>0.055</u>	_____
<u>67.16</u>	<u>ON</u>	_____	_____	_____	_____
<u>68.07</u>	<u>OFF</u>	<u>0.5</u>	<u>9.91</u>	<u>0.050</u>	_____
<u>78.08</u>	<u>ON</u>	_____	_____	_____	_____
<u>78.86</u>	<u>OFF</u>	<u>0.5</u>	<u>10.79</u>	<u>0.046</u>	_____

(4011)(1QTEST)(107991)

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PERSON RECORDING DATA _____

WELL # 37191

HYDROSTRATIGRAPHIC UNIT _____ (see page 1)

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL I.D. _____ in

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME _____

ELAPSED TIME (Units)/(min) ^{ELAPSED RECOVERY TIME}	Pump ON/OFF	VOLUME Pumped (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units)/(gals)	WATER LEVEL
90.87	ON				
91.87	OFF	0.6	13.01	0.046	
104.87	ON				
105.87	OFF	0.75	14.00	0.054	
116.87	ON				
118.55 (10)	OFF	0.70	12.68	0.055	
119.83 (128)		(RECOVERY DATA)			24.40
120.83 (228)					23.71
121.83 (328)					23.19
122.83 (428)					22.79
123.83 (528)					22.36
124.83 (628)					22.08
125.83 (728)					21.75

END OF
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WELL # 37191

HYDROSTRATIGRAPHIC UNIT

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft / PUMPING WELL I.D. _____ in.

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME / / [illegible]

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PERSON RECORDING DATA R Smith - EG&G

WELL # 6286 (881 HILLSIDE)

HYDROSTRATIGRAPHIC UNIT UPPER LARAMIE SANDSTONE

SCREENED INTERVAL 26.43 ft to 36.40 ft from TOC-SS

STATIC WATER LEVEL 26.41 ft @ TOC-SS PUMPING WELL ID 2 in (casing)

DISTANCE TO PUMPING WELL N/A ft WELL T ID 36.85 ft @ TOC-SS
PUMP INTAKE DEPTH 36.08 ft

TEST START TIME 09 18 50 INITIAL SATURATED THICKNESS 10.44 ft

ELAPSED TIME (Units)/(min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units)/(gpm)	WATER LEVEL (ft)
0.0	ON				26.41
2.0	OFF	2.3	2.0	1.15	
11.66	ON				
13.58	OFF	0.55	11.58	0.047	
25.83	ON				
26.32	OFF	0.5	12.74	0.039	
36.83	ON				
38.58	OFF	0	(Pump NOT PRIMING)		
38.74	ON				
39.25	OFF	0.45	12.93	0.035	
51.17	ON				
51.50	OFF	0.45	12.25	0.037	
63.67	ON				

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PERSON RECORDING DATA _____

WELL # 1286

HYDROSTRATIGRAPHIC UNIT _____ (see page 4)

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL ID _____ in

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME _____

ELAPSED TIME (Units) (min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL	DT (min)	AVERAGE Q (pumping well) (Units) (gpm)	WATER LEV. LIST
		(Units)			
64.88	OFF	0.4	13.38	0.030	
66.00	ON				
77 12 67.12 165	OFF	0.4	12.24	0.033	
89.25	ON				
89.83	OFF	0.35	12.71	0.028	
101.33	ON				
102.03	OFF	0.2	12.20	0.016	
117.7	ON				
118.10	OFF	0.45	16.07	0.028	
132.66	ON				
133.23 (0)	OFF	0.4	15.13	0.026	
135.08 (185)		(RECOVERY DATA)			35.51
136.23 (3)		↓			35.25

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PERSON RECORDING DATA _____

WELL # 6286

HYDROSTRATIGRAPHIC UNIT _____ (see Page 1)

SCREENED INTERVAL _____ ft to _____ ft

STATIC WATER LEVEL _____ ft PUMPING WELL ID _____ in

DISTANCE TO PUMPING WELL _____ ft

TEST START TIME _____

ELAPSED TIME (Units) (min)	Pump ON/OFF	VOLUME PUMPED (gals) WATER LEVEL (Units)	DT (min)	AVERAGE Q (pumping well) (Units) (gpm)	WATER LEVEL
137 23 (4)		(RECOVERY DATA)			35.02
138 23 (5)					34.87
139 23 (6)					34.72
140 23 (7)					34.60
141 23 (8)					34.49
142 23 (9)					34.40
143 23 (10)					34.31
144 23 (11)					34.22
145 23 (12)					probe failure
					(UNABLE TO
					DETERMINE
					PROBE)

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Source: *From Confrontation to Reconciliation*

Debitum: MAG22

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111



BIG ROCKY FLATS

Rocky Flats Environmental technology site
Page 404

1

Figure 1.5
881 Hillside Area
Operable Unit 1
Average Selenium Values
in Alluvial Groundwater
all units = UG/L

EXPLANATION

- 0-100
- 100-1000
- 1000-10000
- 10000
- LOCATION WITH NO DAT
- BEDROCK LOCATION
- INDIVIDUAL HAZARDOUS SUBSTANCE SITE
- FRENCH DRAIN

Standard Map Features

- Buildings or other structures
- Lakes and ponds
- Streams, ditches, or other drainage features
- Fences
- Rocky Run boundary
- Paved roads
- Dirt roads

NOTE: The map shows the location of the Rocky Run boundary and the location of the French drain. The map also shows the location of the buildings and other structures, lakes and ponds, streams, ditches, or other drainage features, fences, paved roads, and dirt roads.

NOTE: The map shows the location of the Rocky Run boundary and the location of the French drain. The map also shows the location of the buildings and other structures, lakes and ponds, streams, ditches, or other drainage features, fences, paved roads, and dirt roads.

Scale: 1 inch = 1000 feet
Scale: 1 inch = 1000 feet
Scale: 1 inch = 1000 feet
Scale: 1 inch = 1000 feet

U.S. Department of Energy
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Prepared by

EG&B ROCKY FLATS

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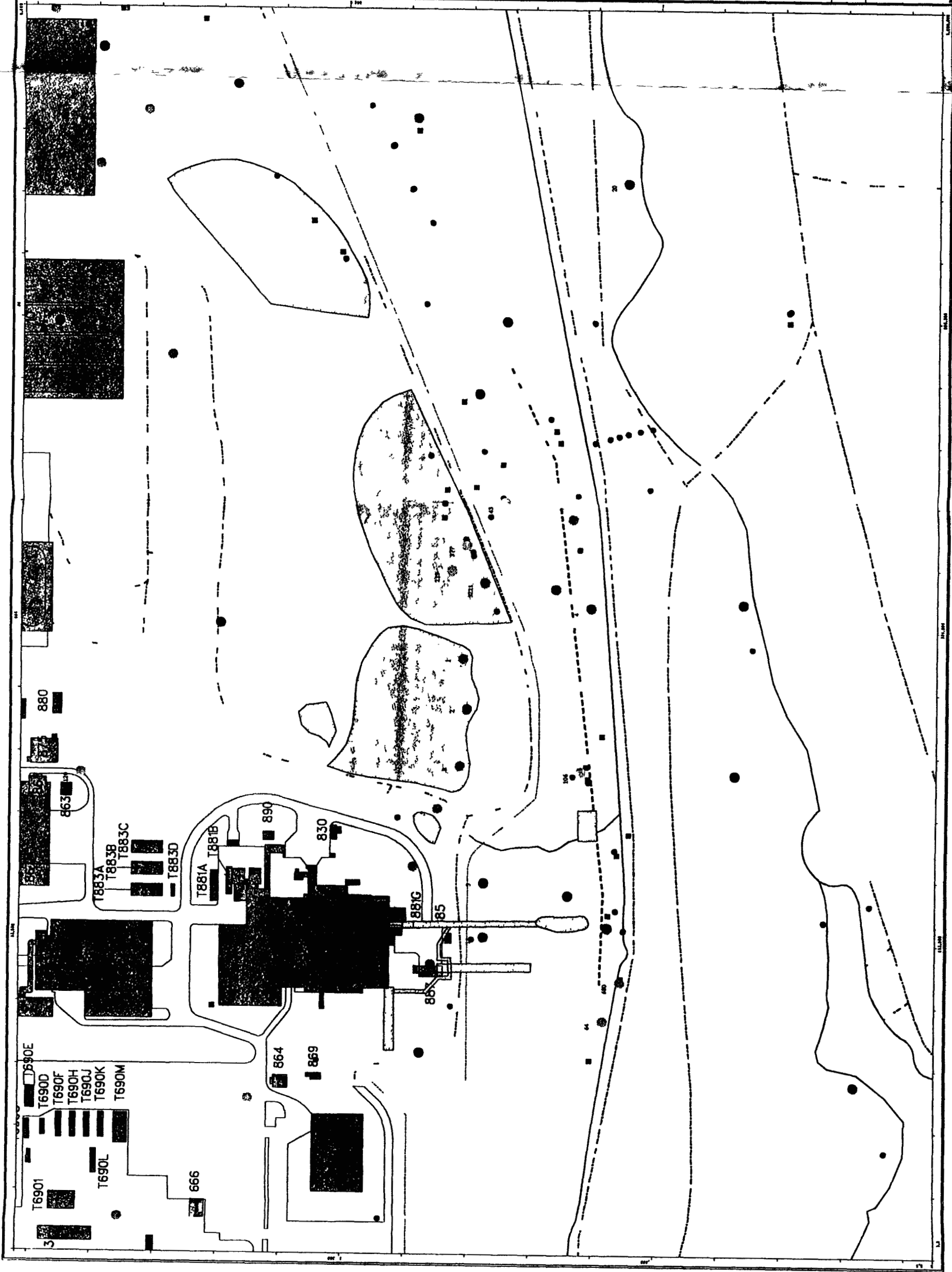


Figure 1-6
881 Hillside Area
Operable Unit 1
Average Target PAHs
in Surface Soils
all units = UG/KG

EXPLANATION

- 0- 00
- 00- 000
- 000- 0000
- 0000
- LOCATION WITH NO DA
- INDIVIDUAL HAZARDOUS WASTE SITE
- FRENCH DRAIN

Standard Map Features

- Buildings or other structures
- Lakes and ponds
- Streams, ditches, or other drainage features
- Fences
- Rocky Flats boundary
- Paved roads
- Dirt roads

Also include:
 - Area of low water
 - Area of high water
 - Area of very high water
 - Area of extreme high water

Notes:
 - The map shows the location of the area of interest relative to the surrounding area.
 - The map shows the location of the area of interest relative to the surrounding area.
 - The map shows the location of the area of interest relative to the surrounding area.

Scale:
 1 inch = 1 mile
 1 centimeter = 100 meters
 1 kilometer = 1000 meters

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ROCKY FLATS
 Rocky Flats Environmental Technology Site
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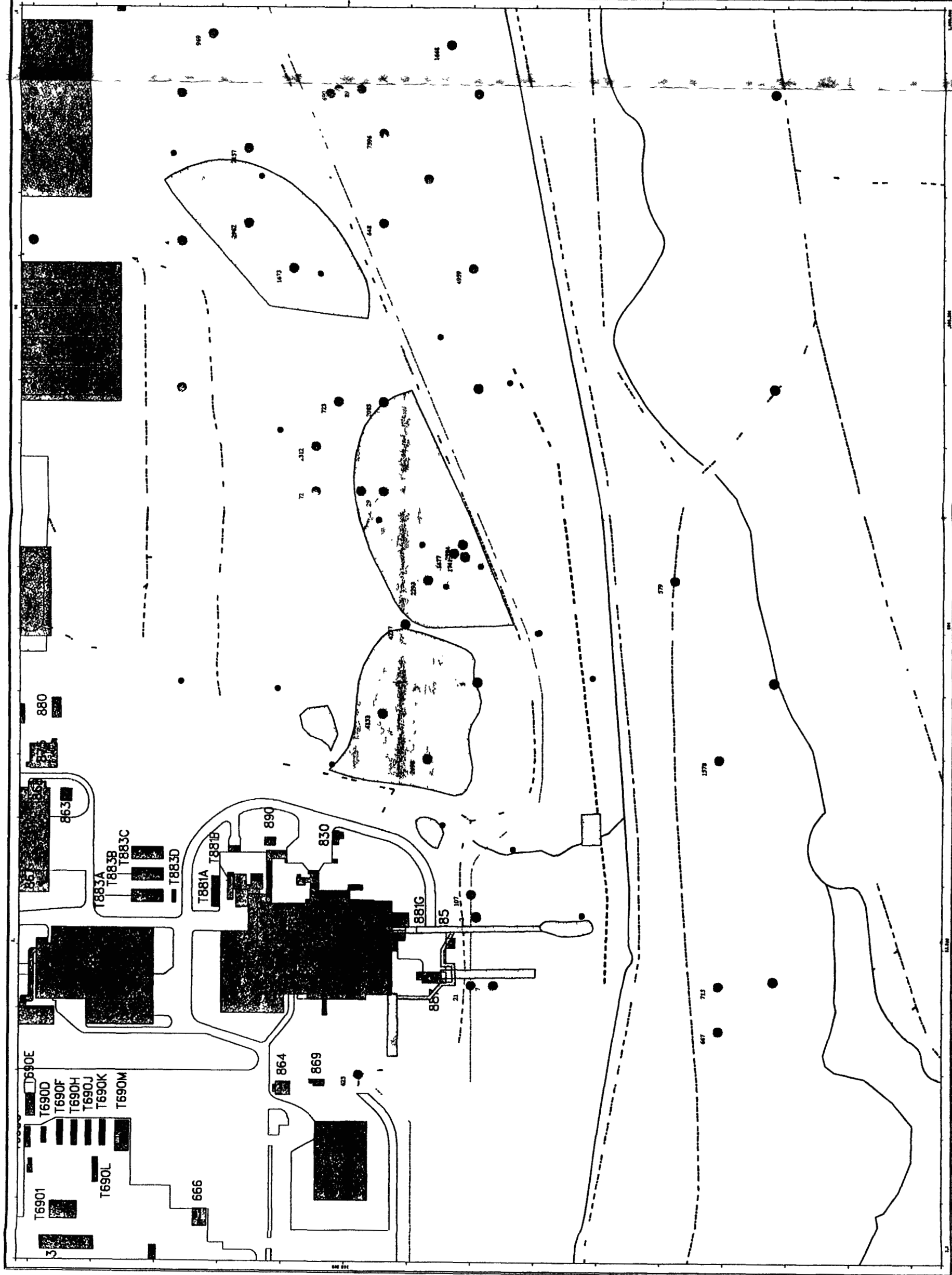


Figure 17
881 Hillside Area
Operable Unit 1
Average Target PCBs
in Surface Soils
all units = UG/KG

EXPLANATION

- 0- 00
- 00- 000
- 000- 0000
- 0000
- LOCATION WITH NO DAT
- INDIVIDUAL HAZARDOUS
- ISTANCE SITE
- FRENCH DRAIN

Standard Map Features

- Buildings or other structures
- Lakes and ponds
- Streams, ditches, or other drainage features
- Fences
- Rocky Flats boundary
- Paved roads
- Dirt roads

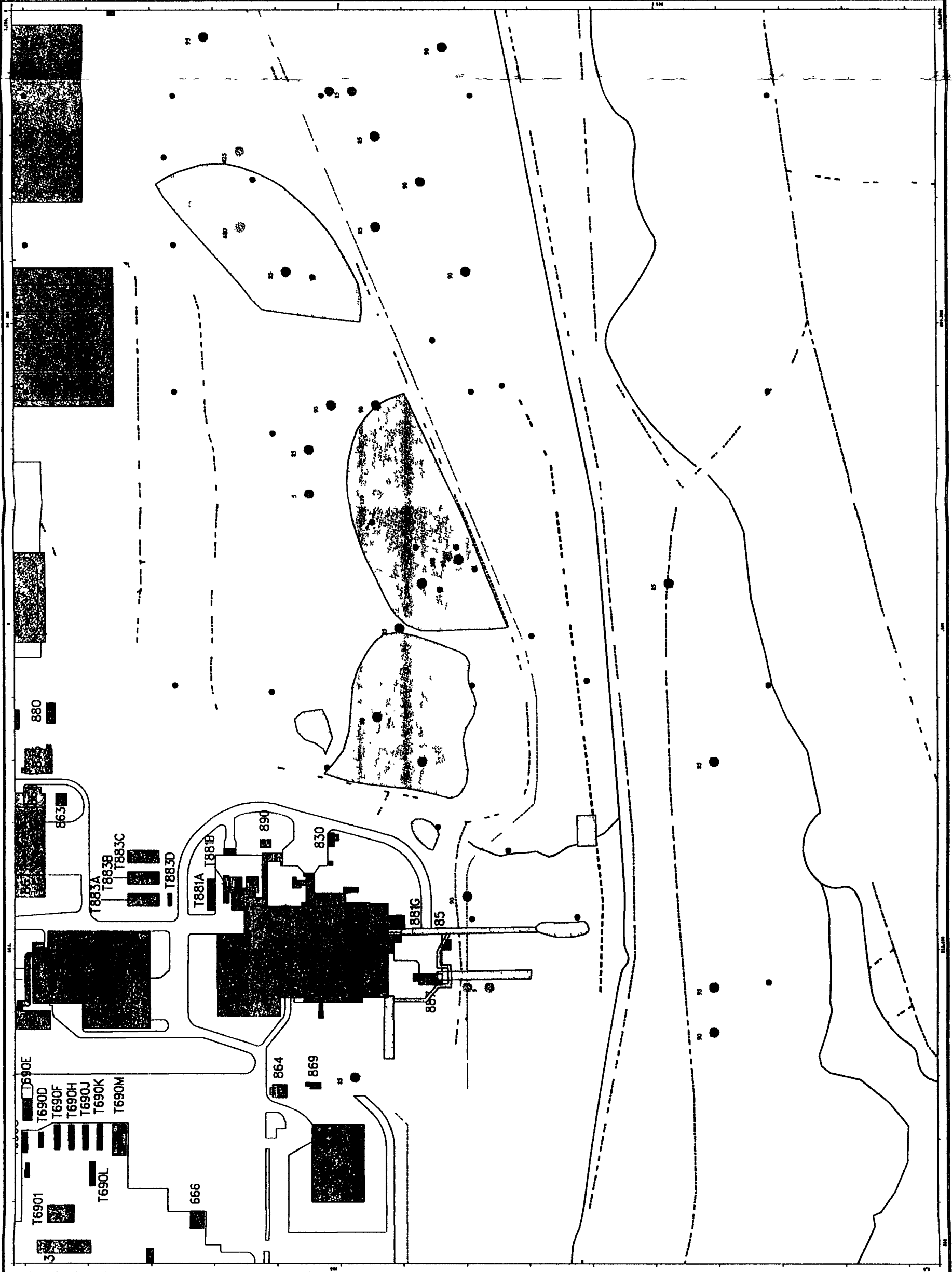
Scale 1:50,000
North arrow pointing up
Rocky Flats boundary
Scale 1:50,000
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Rocky Flats boundary
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North arrow pointing up
Rocky Flats boundary
Scale 1:50,000
North arrow pointing up

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EXPLANATION

11

11

1 10

10-100

100-1000

1000-10000

10000

LOCATION WITH NO DATA

INDIVIDUAL HAZARDOUS SUBSTANCE SITE

BENCH DRAIN

Standard Map Features

Buildings or other structures

silica and bonds

streams, ditches, drainage features

Since

Rocky Flats boundary

sewed roads

virt oada

Author's address:
 Biology, Health and Society Institute
 Public Health
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 Hygiene, London, UK
 1996

1. The following information is being furnished to you for your information only. It is not intended to be used for any other purpose.



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U.S. Department of Energy

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Prepared by

STANDARD ROCKY FLATS

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University of California

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Figure 19
881 Hillside Area
Operable Unit 1
IHSS Locations
Potential Groundwater
Migration Pathways

- EXPLANATION**
- INDIVIDUAL HAZARDOUS SUBSTANCE SITE
 - FRENCH DRAIN
 - ACTIVE W PATHWAYS
 - PRE-FRNC BAIN THW

Standard Map Features

- Buildings or other structures
- Lakes and ponds
- Streams, ditches, or other drainage features
- Fences
- Contours (20' intervals)
- Rocky Rise boundary
- Paved roads
- Dirt roads

Scale: 1 inch = 200 feet
North Arrow

Scale: 1 inch = 200 feet
North Arrow

Scale: 1 inch = 200 feet
North Arrow

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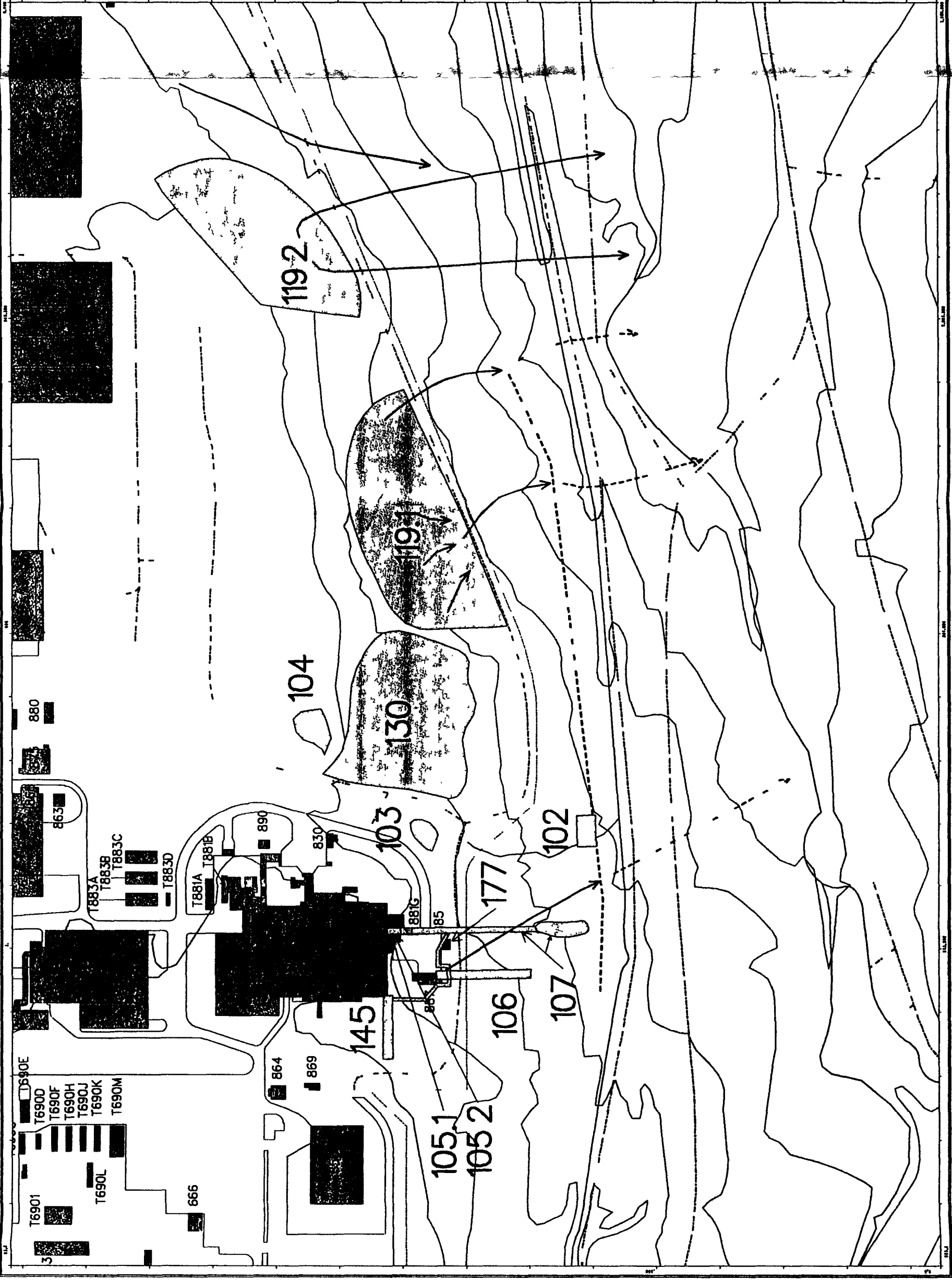
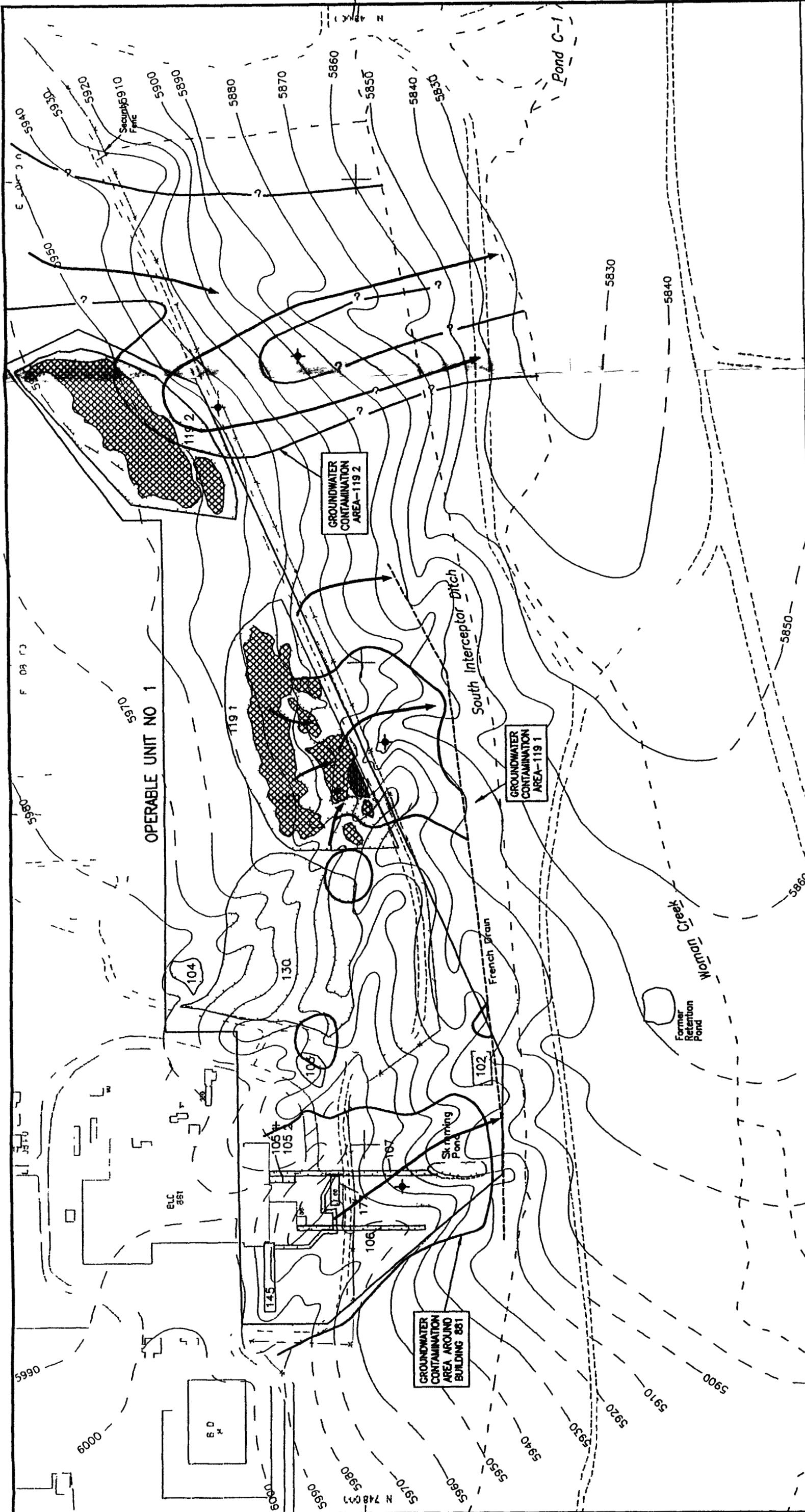


Table 3-1
Summary of Groundwater Remedial Action Alternative Development^a

GENERAL RESPONSE ACTION	PROCESS OPTION	PROPOSED REMEDIAL ACTION ALTERNATIVES							
		0	1	2	3	4	5	6	7
		No Action	Institutional Controls without the French Drain	Institutional Controls with the French Drain	Modified French Drain with Additional Extraction Wells	Groundwater Pumping and Soil Vapor Extraction	Groundwater Pumping and SVE with Thermal Enhancement	Hot Air Injection with Mechanical Mixing	Groundwater Removal by Soil Excavation and Sump Pumps
		N/A	All IHSSs	All IHSSs	All IHSSs	IHSS 119 1	IHSS 119 1	IHSS 119 1	IHSS 119 1
No action	Groundwater monitoring	✓	✓	✓	✓	✓	✓	✓	✓
Institutional controls	Legal restrictions on land use		✓	✓	✓				✓
	Legal restrictions on well placement		✓	✓	✓				✓
Containment	Subsurface drains (existing French Drain)			✓	✓	✓	✓	✓	✓
	Environmental isolation enclosure (optional)								✓
	Surface cap								✓
Removal	Subsurface drains (existing French Drain)			✓	✓	✓	✓	✓	✓
	Horizontal and/or vertical extraction wells or sumps			✓	✓	✓	✓	✓	✓
	Loader/dozer/excavator								✓
	RF/ohmic heating						✓		
In situ treatment of chlorinated solvents	Soil vapor extraction					✓	✓		
	Hot air/steam stripping with mechanical mixing							✓	
Ex situ treatment of chlorinated solvents	Ultraviolet photolysis with chemical oxidation			✓	✓	✓	✓	✓	✓
Ex situ treatment of inorganics	Ion exchange			✓	✓	✓	✓	✓	✓

^a Shaded alternatives were screened from further consideration on the basis of effectiveness, implementability, or cost.



U S DEPARTMENT OF ENERGY
Rocky Flats Environmental Technology Site
Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO 1

**Potential Locations for
Additional Extraction Wells**

EXPLANATION

- FENCE
- CREEK/DRAINAGE
- INDIVIDUAL HAZARDOUS SUBSTANCE SITE (IHSS) AND IHSS DESIGNATION
- DASHED WHERE DISTURBED DURING CONSTRUCTION OF FRENCH DRAIN
- ACTUAL SCRAP METAL AND DRUM STORAGE AREAS IN IHSS 119 1 BASED ON AERIAL PHOTOGRAPHS
- ACTUAL DRUM STORAGE AREA IN IHSS 119 1 BASED ON AERIAL PHOTOGRAPHS
- C1 = 10ft. (DASHED WHERE INFERRED)
- Inferred Extent of Contamination Based on 1/92 Detections
- Potentially Active Contaminant Migration Pathway
- Potential Extraction Well Location

0 100 200
SCALE 1" = 200'

Figure 3--1

Table 4-1
Summary of Detailed Analysis of Alternatives

[illegible]

Table 4.1
(Continued)

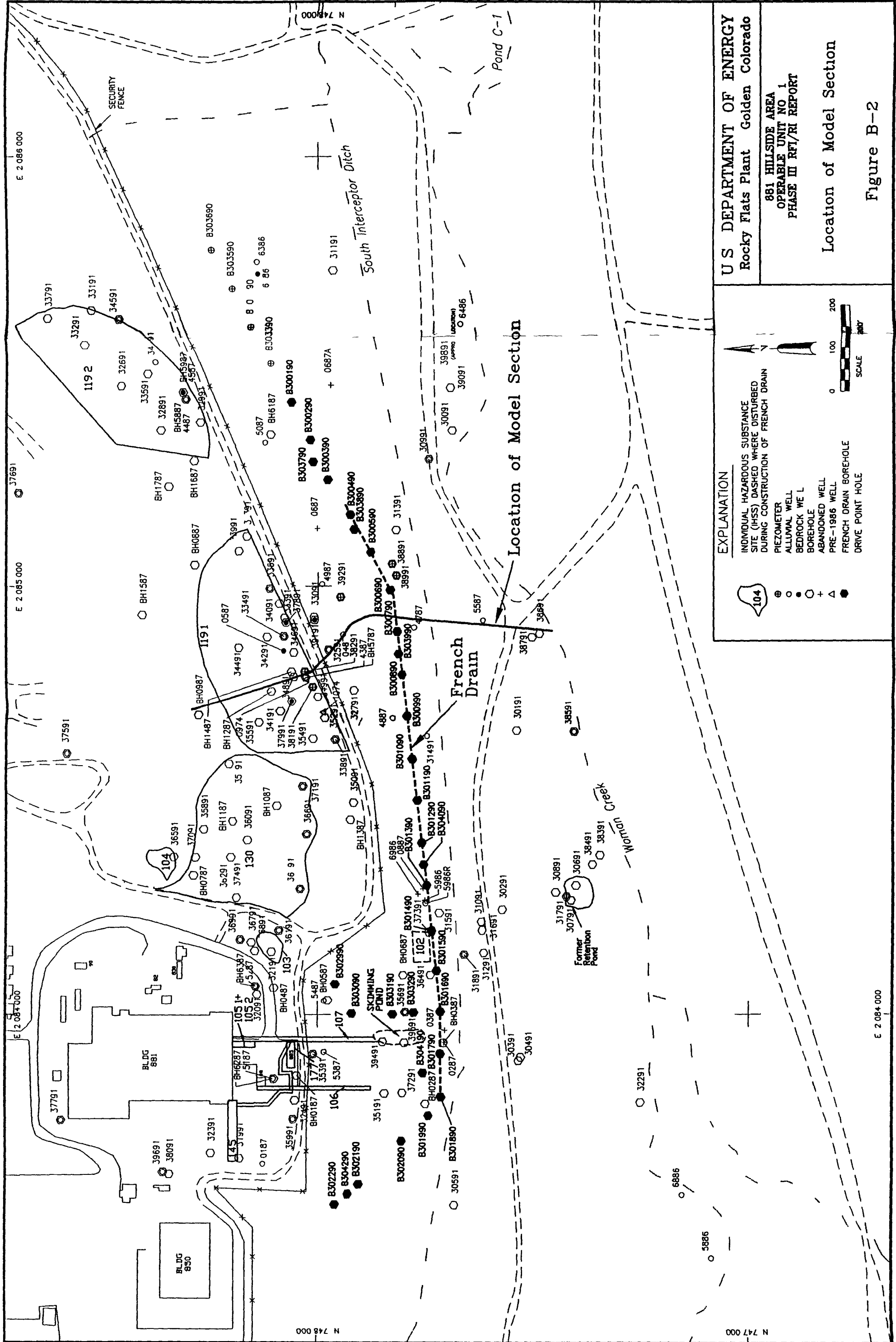
CYRCA Alternative Criteria	ALTERNATIVE Conventional Criteria Without the French Drain	ALTERNATIVE French Drain System	ALTERNATIVE Modified French Drain With Additional Extraction Wells	ALTERNATIVE SVE and SVE Extraction Wells and SVE Extraction Wells	ALTERNATIVE RF Heating and Extraction Wells and SVE Extraction Wells	ALTERNATIVE Hot Air Injection and SVE Extraction Wells	ALTERNATIVE Excavation and SVE Extraction Wells
Reduction of Toxicity Mobility or Volume Through Treatment							
Treatment Process Used	None	None	None	Extracted groundwater would be treated by existing Building 891 UV/peroxide/ion exchange process	Extracted groundwater would be treated by existing Building 891 UV/peroxide/ion exchange process	SVE is a proven technology Recovered soil gas would be treated with activated carbon Extracted groundwater would be treated by existing Building 891 UV/peroxide/ion exchange process	RF heating as an SVE enhancement is an innovative technology Recovered soil gas would be treated with activated carbon Extracted groundwater would be treated by existing Building 891 UV/peroxide/ion exchange process
Amount Destroyed or Treated	None	None	None	Small quantities of COCs would be treated due to low concentrations in groundwater and low extraction rate	Small quantities of COCs would be treated due to low concentrations in groundwater and low extraction rate	SVE may not effectively remove COCs due to low permeability of soils	Greater quantity of COCs may be removed and treated than for Alternatives 4 due to RF heating
Reduction of Toxicity Mobility or Volume	None	None	None	French drain would continue to reduce mobility and volume of COCs Toxicity would be reduced through UV/peroxide/ion exchange process	French drain and additional extraction wells would reduce mobility and volume of COCs Toxicity would be reduced through UV/peroxide/ion exchange process	SVE would effectively reduce volume and mobility of COCs Toxicity would be reduced through carbon regeneration process and UV/peroxide/ion exchange process	Reduction of volume and mobility and toxicity may be slightly more effective than for Alternatives 4 and 5 due to hot air injection and mechanical mixing
Irreversible Treatment	Not Applicable	Not Applicable	Not Applicable	Contaminant removal would be irreversible However DNAPLs may continue to act as source	Contaminant removal would be irreversible However DNAPLs may continue to act as source	Contaminant removal would be irreversible However DNAPLs may continue to act as source	Contaminant removal would be irreversible assuming all DNAPL sources are removed
Type and Quantity of Residuals Remaining after Treatment	Existing contaminant concentrations would remain unchanged	Existing contaminant concentrations would remain unchanged	Existing contaminant concentrations would remain unchanged	Continued operation of the French drain would further reduce volume of contaminants Residual concentrations of COCs would remain in subsurface soils and groundwater	Operation of French drain and additional extraction wells would further reduce volume of contaminants Residual concentrations of COCs would remain in subsurface soils and groundwater	Residual concentrations of COCs may remain at IHSS 119.1 following treatment. Low concentrations of COCs would remain in downgradient groundwater	Residual concentrations of COCs may remain at IHSS 119.1 following excavation. Low concentrations of COCs would remain in downgradient groundwater
Statutory Preferences for Treatment	Does not satisfy preference for treatment.	Does not satisfy preference for treatment.	Does not satisfy preference for treatment.	Satisfies preference for treatment	Satisfies preference for treatment.	Satisfies preference for treatment	Satisfies preference for treatment.

Table 4-1

	ALTERNATIVE No. 1	ALTERNATIVE No. 2	ALTERNATIVE No. 3	ALTERNATIVE No. 4	ALTERNATIVE No. 5	ALTERNATIVE No. 6	ALTERNATIVE No. 7	ALTERNATIVE No. 8	ALTERNATIVE No. 9	ALTERNATIVE No. 10
Short Term Effectiveness										
Community Protection	No increase in potential risks to the public	Reduces potential risks to public by restricting site access	No increase in potential risks to the public	No increase in potential risks to workers	No increase in potential risks to the public	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices	No significant increase in potential risks to the public	No significant increase in potential risks to the public	No significant increase in potential risks to the public	Potential risks to public due to air borne dust generated during excavation activities and transport of excavated soil off site
Worker Protection	No increase in potential risks to workers	No increase in potential risks to workers	No increase in potential risks to workers	No increase in potential risks to workers	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices	Potential risks from exposure to COCs in groundwater or soil vapor associated with operating specialized equipment. Risks would be minimized through standard health and safety practices	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices	Potential risks from exposure to COCs in groundwater or soil vapor associated with drilling and construction Risks would be minimized through standard health and safety practices
Environmental Impacts	No additional environmental impacts	No additional environmental impacts	No additional environmental impacts	No additional environmental impacts	Minor impacts to soil including limited loss of vegetation	Minor loss of vegetation associated with construction activities and heating may have significant impact on subsurface	Significant impact on environment due to rigorous soil mixing and in situ heating In addition soil stability problems at OU 1 would be exacerbated by the process	Significant impact on environment due to rigorous soil mixing and in situ heating In addition soil stability problems at OU 1 would be exacerbated by the process	Significant impact on environment due to rigorous soil mixing and in situ heating In addition soil stability problems at OU 1 would be exacerbated by the process	Excavation would have significant short term impacts on environment
Time Until Response Objectives are Achieved	Groundwater would continue to meet MCLs at Woman Creek.	Groundwater would continue to meet MCLs at Woman Creek.	Groundwater would continue to meet MCLs at Woman Creek.	Groundwater would continue to meet MCLs at Woman Creek.	SVE would remove source in 5 years Groundwater would continue to meet MCLs at Woman Creek	Enhanced SVE would remove source in 3 years Groundwater would continue to meet MCLs at Woman Creek.	Sources would be removed in less than 1 year Groundwater would continue to meet MCLs at Woman Creek.	Sources would be removed in less than 1 year Groundwater would continue to meet MCLs at Woman Creek.	Sources would be removed in less than 1 year Groundwater would continue to meet MCLs at Woman Creek.	Sources would be removed and soil treated in less than 1 year Groundwater would continue to meet MCLs at Woman Creek
Implementability										
Ability to Construct and Operate	Not Applicable	Not Applicable	Would only include operation of existing equipment.	Simple to construct and operate new wells using conventional readily available technology	Simple to construct using conventional readily available technology	Simple to construct and operate using conventional readily available technology	Simple to implement using readily available technology However soil stability concerns may limit operation of mixing device	Excavation can be implemented using standard earth-moving equipment and dewatering pumps However potential radionuclide contamination in subsurface soils may limit ability to transfer soils off site	Excavation can be implemented using standard earth-moving equipment and dewatering pumps However potential radionuclide contamination in subsurface soils may limit ability to transfer soils off site	Excavation can be implemented using standard earth-moving equipment and dewatering pumps However potential radionuclide contamination in subsurface soils may limit ability to transfer soils off site
Ease of Doing More Action if Needed	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions	Would not limit the ability to perform future remedial actions
Ability to Monitor Effectiveness	Existing monitoring programs would track movement of COCs	Existing monitoring programs would track movement of COCs	Existing monitoring programs would track movement of COCs	Existing monitoring programs would track movement of COCs	Existing monitoring programs would track movement of COCs	Existing monitoring programs would track movement of COCs	Existing monitoring programs would determine effectiveness	Existing monitoring programs would determine effectiveness	Existing monitoring programs would determine effectiveness	Existing monitoring programs would determine effectiveness

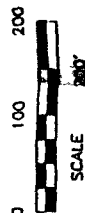
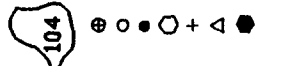
Table 4-1
(Continued)

CERCLA Analysis Criteria	ALTERNATIVE 0 No Action	ALTERNATIVE 1 Institutional Controls Without the French Drain	ALTERNATIVE 2 Institutional Controls With French Drain	ALTERNATIVE 3 Modified French Drain With Additional Extraction Wells	ALTERNATIVE 4 Groundwater Pumping and Soil Vapor Extraction	ALTERNATIVE 5 Groundwater Pumping and Soil Vapor Extraction With Enhanced Extraction	ALTERNATIVE 6 Hot Air Injection With Mechanical Mixing	ALTERNATIVE 7 Soil Excavation and Groundwater Removal With Sump Pumps
Ability to Obtain Permits/Coordination with Agencies	No problems anticipated	No problems anticipated	No problems anticipated	No problems anticipated	No problems anticipated	No problems anticipated	No problems anticipated	Potential radionuclide contamination in subsurface soils may limit ability to transfer soils off site
Availability of Services and Capacities	No services required	Minimal services required	Minimal additional services required	Services readily available	Services readily available	Services readily available	Services readily available	Services readily available
Availability of Equipment, Specialists and Materials	None required	None required	None required	Readily available	Readily available	Readily available	Likely to be readily available although technology is considered innovative	Readily available
Availability of Technologies	None required	None required	None required	Readily available would utilize common construction techniques	Readily available	Readily available	Likely to be readily available although technology is considered innovative	Readily available would utilize common construction techniques
Cost								
Present Worth (1994) Cost	\$1,895,100	\$1,895,100	\$1,493,300	\$1,648,700	\$8,141,800	\$7,503,100	\$5,030,800	\$13,093,700



EXPLANATION

- INDIVIDUAL HAZARDOUS SUBSTANCE SITE (IHSS) DASHED WHERE DISTURBED DURING CONSTRUCTION OF FRENCH DRAIN
- PIEZOMETER
- ALLUVIAL WELL
- BEDROCK WELL
- BOREHOLE
- ABANDONED WELL
- PRE-1986 WELL
- FRENCH DRAIN BOREHOLE
- DRIVE POINT HOLE



U S DEPARTMENT OF ENERGY
Rocky Flats Plant Golden Colorado

881 HILLSIDE AREA
OPERABLE UNIT NO. 1
PHASE III RFI/RI REPORT

Location of Model Section

Figure B-2

Table D 1
Potential Federal
Action Specific ARARs and TBCs for Proposed Remedial Action Alternatives

		1		3		6											
Standard Requirement (Criteria or Limitation)		Citation		No Action		Intitutonal Controls without the French Drain		Institu tional Controls with the French Drain		Modified French Drain with Additional Extraction Wells		(Groundwater Pumpin- and Syt with Thermal Enhancement ion		Steam Injection with Mechanical Mixing		Soil Excavation and (Groundwater Retoval with Sump Pumps	
Resource Conservation and Recovery Act (RCRA)		42 USC Secs 6901 6987				NA		NA		NA		A ²		A ²		A ²	
A Criteria for Classification of Solid Waste Disposal Facilities and Practices		40 CFR Part 257		NA		R/Y		R/Y		R/Y		A ² /Y		A ² /Y		A ² /Y	
B Hazardous Waste Management Systems General		40 CFR Part 260		R/Y		R/Y		R/Y		R/Y		A ² /Y		A ² /Y		A ² /Y	
C Identification and Listing of Hazardous Wastes		40 CFR Part 261		R/Y		R/Y		R/Y		R/Y		A ² /Y		A ² /Y		A ² /Y	
D Proposed Definition of Hazardous Waste to Exclude Environmental Media ¹ 58FR48156		40 CFR Part 260 261 261 4 261 42 and 268		C/Y ¹		C/Y ¹		C/Y ¹		C/Y ¹		C/Y ¹		C/Y ¹		C/Y ¹	
E Standards Applicable to Generators of Hazardous Waste		40 CFR Part 262		R/Y		R/Y		R/Y		R/Y		A ² /Y		A ² /Y		A ² /Y	
F Releases from Solid Waste Management Units		40 CFR Part 264 Subpart F		R/Y		R/Y		R/Y		R/Y		R/Y		R/Y		R/Y	
G Closure and Post Closure		40 CFR Part 264 112 Subpart G and 264 601		R/Y		R/Y		R/Y		R/Y		A/Y		A/Y		A/Y	
H Use and Management of Containers		40 CFR Part 264 Subpart I		NA		NA		NA		NA		A/Y		A/Y		A/Y	
I Landfills		40 CFR Part 264 Subpart N		NA		NA		NA		NA		NA		NA		NA	
J Miscellaneous Units		40 CFR Part 264 Subpart X		NA		NA		NA		NA		R/Y		R/Y		R/Y	
K Air Emission Standards for Process Vents		40 CFR Part 264 1032 and 264 1033 Subpart AA		NA		NA		NA		NA		A/Y		A/Y		A/Y	
L Air Emission Standards for Equipment Leaks		40 CFR Part 264 1056 1057 Subpart BB		NA		NA		NA		NA		A/Y		A/Y		A/Y	
M Proposed Air Emission Standards for Storage Units		40 CFR Part 264 1083 Subpart CC		NA		NA		NA		NA		C/Y ³		C/Y ³		C/Y ³	
N Temporary Unit		40 CFR Part 264 553 Subpart S		NA		NA		NA		NA		A/Y		A/Y		A/Y	

² Assumes requirements of 261.42 could be met, acceptable risk range 10⁻⁴ - 10⁻⁶ and levels in soil and groundwater do not pose human health hazard nor environmental hazard

³ Applies to new treatment system

May apply if concentration of organics in tank exceed 500 ppmv

Table D 1
Potential Federal
Action Specific ARARs and TBCs for Proposed Remedial Action Alternatives

		0	1	2	3	4	5	6	7
Standard Requirement Criteria or Limitation	Citation	No Action	Institutional Controls without the French Drain	Institutional Controls with the French Drain	Modified French Drain with Additional Extraction Wells	Groundwater Pumping and SVE with Thermal Enhancement	Steam Inject on with Mechanical Mixing	Soil Excavation and Groundwater Removal with Sump Pumps	
O Corrective Action for Solid Waste Management Unit (CAMU)	40 CFR 264 552 Subpart S	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y
P Interim Status Standards for Owners and Operators of Hazardous Waste Treatment Storage and Disposal Facilities	40 CFR Part 265	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y	R/Y
Q Interim Status Standards for Owners and Operators of New Hazardous Waste Land Disposal Facilities	40 CFR Part 267	NA	NA	NA	NA	NA	NA	NA	NA
R Land Disposal Restrictions	40 CFR Part 268	NA	NA	NA	NA	A/Y	A/Y	A/Y	A/Y
Toxic Substances Control Act	15 USC Secs 2601 2629	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y
A PCB Requirements	40 CFR Part 761								
Clean Water Act	33 USC Secs 1251 1376								
A Discharge of Effluent	40 CFR Sec 125 100								
FF CA CWA 90-1 NPDES Federal Facility Compliance Agreement	40 CFR Sec 122 41	A/Y	NA	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y
B Toxic Pollutant Effluent Standards	40 CFR 129	NA	NA	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y
C Discharge of Stormwater	40 CFR Sec 122 21 40 CFR Sec 122 26	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y	A/Y
Atomic Energy Act	42 USC Secs 2011 et seq								
A Radiation Protection and Radioactive Waste Management	10 CFR Part 20 1301 20 1302 Subpart D and K	NA	NA	NA	NA	NA	NA	NA	A/Y
B Performance Objectives in Licensing for Land Disposal of Radioactive Waste	10 CFR Part 61	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y

4 Applies to residuals of treatment system such as spent carbon, HEPA filters or ion exchange resins
5 Considered for impacts to groundwater

Table D 1
Potential Federal
Action Specific ARARs and TBCs for Proposed Remedial Action Alternatives

		0	1	2	3	4	5	6	7
Standard Requirement Criteria or Limitation	Citation	No Action	Insitutional Controls without the French Drain	Insitutional Controls with the French Drain	Modified French Drain with Additional Extraction Wells	Groundwater Pumping and Soil Vapor Extraction	Groundwater Pumping and SVE with Thermal Enhancement	Steam Injection with Mechanical Mixing	Soil Excavation and Groundwater Removal with Sump Pumps
Clean Air Act									
A Prevention of Significant Deterioration Requirements	42 USC Secs 7401 7642 40 CFR 52	NA	NA	NA	NA	NA	NA	NA	NA
B National Emission Standards for Hazardous Air Pollutants	40 CFR 61	NA	NA	NA	NA	R/Y	R/Y	R/Y	R/Y
Safe Drinking Water Act									
Undergrund Inspection Control Program Class V Wells	40 CFR 146 5	NA	NA	NA	NA	NA	NA	R/Y	NA
DOE Orders									
General Environmental Protection Program	5400 1	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y
Environmental Compliance Issue Coordination	5400 2A	NA	NA	NA	NA	NA	NA	NA	NA
Radiation Protection of the Public and Environment	5400 5	NA	NA	NA	NA	NA	NA	NA	C/Y
Environment Safety and Health Programs for DOE Operations	5480 1B	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y
Radioactive Waste Management	5820 2A	NA	NA	NA	NA	NA	NA	NA	C/Y
Hazardous and Radioactive Mixed Hazardous Waste Management	5400 3	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y
Environmental Protection Safety and Health Protection Standards	5480 4	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y	C/Y

Endnotes

A = Applicable
R = Relevant and Appropriate
NA = Not an ARAR
C = Considered
Y = in compliance or can be in compliance
N = not in compliance/standard exceeded

Table D 2
Potential State
Action Specific ARARs and TBCs for Proposed Remedial Action Alternatives

Standard Requirement Criteria or Limitation	Citation	0																4				6																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																		
		No Action				Institutional Controls without French Drain				Institutional Controls with French Drain				Modified French Drain with Aggregates Extraction				Groundwater Pumping and Solvent Extraction				Groundwater Pumping and Solvent Extraction with Thermal Treatment				Steam Injection with Mechanical Mixing				Excavation and Groundwater Removal with Sump Pumps																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																										
Colorado Hazardous Waste Act and State Hazardous Waste Siting Act	CRS § 25-15-101 et seq 25-15-200 et seq 25-15-301 et seq																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																							

1 Applies to new treatment system
2 Applies to residuals of treatment system such as spent carbon HEPA filter or exchange

Table D 2
Potential State
Action Specific ARARs and TBCs for Proposed Remedial Action Alternatives

Standard Requirement Criteria or Limitation	Citation	0 1 2 3 4 5 6					
		No Action	Institutional Control without French Drain	Institutional Controls with French Drain	Modified French Drain with Additional Extraction	Groundwater Pumping and Soil Vapor Extraction	Grout Water Partitioning and Equalization
Colorado Air Pollution Prevention Control Act, as amended	CRS 25 7 112						
Colorado Air Pollution Control Regulations Air Pollutant Emission Notice Requirements	5 CCR 1001 5 Regulation 3 Subpart A	NA	NA	NA	NA	NA	NA
State Construction Permits	5 CCR 1001 5 Regulation 3 Subpart B	NA	NA	NA	NA	NA	NA
Operating Permit Program	5 CCR 1001 5 Regulation 3 Subpart C	NA	NA	NA	NA	NA	NA
Control of Emissions Volatile Organic Compound	Regulation 7 General Provisions	NA	NA	NA	NA	R/Y	R/Y
Soil Erosion Dust Blowing Act	CRS 35 72 101 et seq	NA	NA	NA	R/Y	R/Y	R/Y
Act to Establish Power and Duties of Board of Health Department of Health	CRS § 25 1 107 25 1 108 and 25 11 104	NA	NA	NA	NA	NA	NA
Colorado Rules and Regulations Pertaining to Radiation Control	See below						
A Radioactive Material Other than Source Material	6 CCR 1007 1 1 Part III RH 3 3 1 Schedule A	NA	NA	NA	NA	NA	NA
B Standards for Protection Against Radiation	6 CCR 1007 1 Part IV RH 4 2 1-4 2 2 3	NA	NA	NA	NA	NA	NA
Colorado Noise Abatement Statute	CRS 25 12 101 et seq	NA	NA	NA	NA	NA	NA
Storage Tank Facility Owner/Operator Guidance Documents	Colorado Department of Health December 1992 ¹	NA	NA	NA	NA	NA	NA
State Engineers Authorities							
Colorado Water Well & Pump Installation Regulations	CRS 37 91 101 112 2CCR402 2	NA	NA	NA	NA	NA	NA

3 Construction requirements do not apply to treatment alternative source (without consideration of other sources)- although some chemicals could trigger a requirement for an operating permit, substantive requirements are found in Regulation 7 for RACT
4 Minimal soil disturbance french drain remains in place
5 Assumes no action that would newly disturb rocks or soil

Table D 3
Potential Federal and State
Location Specific ARARs and TBCs for Proposed Remedial Action Alternatives

[illegible]

